

[Click here to view linked References](#)

<https://dx.doi.org/10.1016/j.rse.2022.113136>

This is the Pre-Published Version.

1 **Journal:** Remote Sensing of Environment

2

3 **Title:** A new object and class based gap-filling method for PlanetScope satellite image time series

4

5 **Author List:** Jing Wang¹, Calvin KF Lee¹, Xiaolin Zhu², Ruyin Cao³, Yating Gu¹, Shengbiao Wu¹,
6 Jin Wu^{1,4,5*}

7

8 **Author Affiliations:**

9 (1) Research Areas of Ecology and Biodiversity, School for Biological Sciences, The University
10 of Hong Kong, Hong Kong, China

11 (2) Department of Land Surveying and Geo-Informatics, The Hong Kong Polytechnic University,
12 Hong Kong, China

13 (3) School of Resources and Environment, University of Electronic Science and Technology of
14 China, Sichuan, China

15 (4) Institute for Climate and Carbon Neutrality, The University of Hong Kong, Hong Kong, China

16 (5) State Key Laboratory of Agrobiotechnology, The Chinese University of Hong Kong, Shatin,
17 Hong Kong, China

18

19

20 *** Corresponding Author:** Jin Wu

21 School for Biological Sciences, The University of Hong Kong, Pokfulam Road, Hong Kong (email:
22 jinwu@hku.hk; phone: +852 2299-0655)

23

24 **Figures/Tables Record:**

25 Number of figures/tables: 10 figures, 4 tables

26 Supporting materials: 13 figures, 1 table

27

28 **Abstract**

29 PlanetScope CubeSats data with a 3-meter resolution, frequent revisits, and global coverage have
30 provided an unprecedented opportunity to advance land surface monitoring over the recent years.
31 Similar to other optical satellites, cloud-induced data missing in PlanetScope satellites
32 substantially hinders its use for broad applications. However, effective gap-filling in PlanetScope
33 image time series remains challenging and is subject to whether it can 1) consistently generate
34 high accuracy results regardless of different gap sizes, especially for heterogeneous landscapes,
35 and 2) effectively recover the missing pixels associated with rapid land cover changes. To address
36 these challenges, we proposed an object and class based gap-filling ('OCBGF') method. Two
37 major novelties of OCBGF include 1) adopting an **object**-based segmentation method in
38 conjunction with an unsupervised **classification** method to help characterize the landscape
39 heterogeneity and facilitate the search of neighboring valid pixels for gap-filling, improving its
40 applicability regardless of the gap size; 2) employing a scenario-specific gap-filling approach that
41 enables effective gap-filling of areas with rapid land cover change. We tested OCBGF at four sites
42 representative of different land cover types (plantation, cropland, urban, and forest). For each site,
43 we evaluated the performance of OCBGF on both simulated and real cloud-contaminated scenarios,
44 and compared our results with three state-of-the-art methods, namely Neighborhood Similar Pixel
45 Interpolator (NSPI), AutoRegression to Remove Clouds (ARRC), and Spectral-Angle-Mapper
46 Based Spatio-Temporal Similarity (SAMSTS). Our results show that across all four sites, OCBGF
47 consistently obtains the highest accuracy in gap-filling when applied to scenarios with various gap
48 sizes (RMSE=0.0065, 0.0090, 0.0092, and 0.0113 for OCBGF, SAMSTS, ARRC, and NSPI,
49 respectively) and with/without rapid land cover changes (RMSE=0.0082, 0.0112, 0.0119, and
50 0.0120 for OCBGF, SAMSTS, ARRC, and NSPI, respectively). These results demonstrate the
51 effectiveness of OCBGF for gap-filling PlanetScope image time series, with potential to be
52 extended to other satellites.

53 **Keywords:** Gap-filling, CubeSats, image reconstruction, cloud removal, object-based
54 segmentation

55

56 **1. Introduction**

57 Recent advances in Earth observation satellites with increasing spatial and temporal resolutions
58 have created unprecedented opportunities for monitoring rapid and fine-scale changes on the
59 Earth's surface. One typical example of these advances is the PlanetScope constellation that is
60 made of 180+ micro satellites (CubeSats) (Planet, 2021). These CubeSats altogether provide a
61 daily-to-weekly global coverage at a 3-meter spatial resolution with four spectral bands (i.e. red,
62 green, blue, and near-infrared (NIR)) (Roy et al., 2021; Wang et al., 2020). As a result of these
63 specifications, PlanetScope satellites have been increasingly suggested as a powerful and new
64 means to improve fine-scale Earth's surface monitoring. Similar to other optical satellite
65 observations, PlanetScope data are subject to cloud/cloud shadow contamination. According to the
66 global statistics (Ju and Roy, 2008; Norris et al., 2016), around one-third of the global land surface
67 is covered by clouds within a year, resulting in large quantities of missing data that hinder the
68 wider applications of optical satellite measurements including PlanetScope (Roy et al., 2021).
69 Therefore, accurate and effective gap-filling is important to advance the use of PlanetScope data
70 for monitoring those rapid changes in land surface statuses and processes, such as fine-scale land
71 use development, agricultural expansion, natural disasters and associated impact assessments
72 (Feng et al., 2022; Halls and Magolan, 2019; Wang et al., 2019; Zeng et al., 2018, 2021),
73 forecasting seasonal crop growth (Kimm et al., 2020; Sadeh et al., 2021), quantifying fine-scale
74 plant phenology (Chen et al., 2019; Wang et al., 2020; Wu et al., 2021), and characterizing surface
75 carbon and water fluxes (Dechant et al., 2022; Kong et al., 2022; McCabe et al., 2017).

76 To gap-fill missing data due to cloud/cloud shadow contamination in satellite images, many
77 methods have been developed. These methods can be grouped into four categories according to
78 the auxiliary information used, including fusion-, spatial-, temporal-, and spatiotemporal- based
79 methods (Cao et al., 2020; Shen et al., 2015). Fusion-based methods have been used for gap-filling
80 based on the integration of multisource images from other optical satellites (Luo et al., 2018; Roy
81 et al., 2008) or synthetic aperture radar (SAR) images (Huang et al., 2015; Li et al., 2020b).
82 However, these methods suffer from inconsistencies in spatial resolution and radiometric
83 characteristics among different data types (Cao et al., 2020). Spatial-based methods rely on the
84 assumption that adjacent pixels tend to be more similar due to spatial autocorrelation, and often
85 use the neighboring clear pixels to fill gaps via various approaches such as spatial interpolation

86 methods (Pringle et al., 2009; Zhang et al., 2007) and inpainting methods (Lorenzi et al., 2011;
87 Maalouf et al., 2009). These methods have been demonstrated to be effective for filling data gaps
88 associated with small cloud regions or gaps due to instrument errors such as Landsat Scan Line
89 Corrector (SLC)-off images, but uncertainty massively increases with the size of data gaps (Shen
90 et al., 2015). Temporal-based methods rely on the assumption that there are no obvious land cover
91 changes over a short period of time, and missing pixels in the cloudy image ('target image') can
92 be substituted or modeled with the pixels from cloud-free images acquired on adjacent dates
93 ('reference images') after a relative normalization process. These methods include temporal
94 filtering, temporal replacement, and machine learning based methods (Li et al., 2020a; Lin et al.,
95 2013; Yan and Roy, 2020; Zeng et al., 2013). Nevertheless, these methods are sensitive to the
96 selection of the reference images and generate large uncertainty in heterogeneous landscapes as
97 they usually assume that spectrally similar pixels come from the same class and have the same
98 temporal changing patterns without accounting for the intra-class temporal variability (Chen et al.,
99 2011; Shen et al., 2015). The fourth type is spatiotemporal-based methods that integrate both
100 spatial and temporal information for gap-filling (Zhang et al., 2018; Zhu et al., 2012a). For
101 example, a neighborhood similar pixel interpolator (NSPI) method that combines both temporal
102 predictions from the cloud-free reference image and spatial predictions from pixels outside the
103 data gaps of the target image has been proposed to gap-fill the SLC-off data missing in individual
104 Landsat images (Zhu et al., 2012a), which was further modified to gap-fill the cloud/cloud shadow-
105 induced data missing in a time series of Landsat images (Zhu et al., 2018). However, it is criticized
106 for the sensitivity to reference images used (Cao et al., 2020). To reduce the dependence on specific
107 reference images, Cao et al. (2020) developed an autoregression method to remove clouds (ARRC)
108 for Landsat data that explicitly uses the autoregression of Landsat image time series of the adjacent
109 clear pixels to gap-fill those missing pixels.

110 Among the four categories of gap-filling methods described above, spatiotemporal-based methods
111 are more commonly used since they generally obtain satisfactory results (Cao et al., 2020; Shen et
112 al., 2015). For example, recently developed NSPI and ARRC have been tested across multiple
113 sites across the globe with demonstrated high performance in gap-filling (Cao et al., 2020; Zhu et
114 al., 2018). However, most of these methods are tested on medium to coarse spatial resolution
115 satellite data (e.g. Moderate Resolution Imaging Spectroradiometer (MODIS), Landsat, and

116 Sentinel-2) (Griffiths et al., 2019; Liu et al., 2017; Shen et al., 2015), and have rarely been explored
117 in high spatial resolution satellite data, such as PlanetScope with a 3-m spatial resolution.
118 Therefore, their applicability to the high-resolution PlanetScope remains unknown and may be
119 constrained by the following two factors.

120 The first one is associated with the gap size concern. For the spatial component in spatiotemporal-
121 based methods, the accuracy and efficiency have been demonstrated to decrease significantly with
122 increasing data gap size (or the number of pixels with missing values). Meanwhile, the temporal-
123 based component in the spatiotemporal-based methods relies on the spectrally similar pixels in the
124 neighborhood of data missing pixels to model the temporal changing pattern between the target
125 and reference images, and the search for the spectrally similar pixels for each data missing pixel
126 is time-consuming, especially for the large data gaps. Moreover, due to a much higher spatial
127 resolution of PlanetScope images than traditional satellite data, two novel issues could also
128 emerge. One is even for the same cloud/shadow size (in terms of m^2), there would be more pixels
129 with data missing in PlanetScope images than others. Second, due to the spatial heterogeneity of
130 land cover in nature, there could be multiple objects within the same class or multiple classes
131 within the same object in high-resolution PlanetScope images (e.g. Fig. S1). In other words,
132 concerns about the spatial heterogeneity issue can be more serious in PlanetScope images than in
133 traditional satellite data, especially for those big-size data gaps. Thus, methodological
134 improvements are needed to address the above issues that are particularly important to high-
135 resolution PlanetScope data. Object-based segmentation methods have been increasingly used to
136 automatically segment spectrally and spatially similar pixels into independent objects in satellite
137 images (Hossain and Chen, 2019; Myint et al., 2011), and have been recently implemented in gap-
138 filling (Case and Vitti, 2021; Maxwell et al., 2007; Wu et al., 2018; Yan and Roy, 2018) and up to
139 spatiotemporal fusion methods (Guan et al., 2017; Huang and Zhang, 2014; Luo et al., 2018; Xi et
140 al., 2019). For example, a Spectral-Angle-Mapper Based Spatio-Temporal Similarity (SAMSTS)
141 method (Yan and Roy, 2018) that segmented image time series into a segmentation map has been
142 demonstrated effective for gap-filling large gaps in Landsat image time series. Such object-based
143 segmentation methods may be able to address the above concern by reducing the sensitivity on
144 gap size while improving the efficiency for filling large gaps in very high-resolution satellite

145 images (Case and Vitti, 2021), but the relevant workflow has not yet been tested on PlanetScope
146 images.

147 In addition to the gap size concern, the appropriate selection of reference images is vital to achieve
148 successful and accurate gap-filling, but could result in large uncertainties when rapid land cover
149 changes occur. Gap-filling methods often use either single or multiple reference images. The
150 single-reference-image methods require a clear reference image that is acquired at a close date to
151 the target image for gap-filling (Zhu et al., 2012b). Although PlanetScope has a daily-to-weekly
152 temporal coverage, this data requirement remains limited in tropical or subtropical areas due to
153 continuous cloud contamination, especially during high rainfall wet seasons (Roy et. al., 2021;
154 Wang et al., 2020). To address this issue, the multiple-reference-image methods select multiple
155 reference images that are cloud-free and have the highest similarity with the target image for each
156 cloud patch instead of the whole image (Lin et al., 2014). Some studies have demonstrated that
157 multiple-reference-image methods have better performance than single-reference-image methods
158 (Chen et al., 2017; Lin et al., 2013), but it remains difficult to obtain cloud-free reference images
159 for large cloud patches, as the similarities between the reference and target images for all objects
160 even within a cloud patch can vary considerably (Cao et al., 2020), particularly for the
161 heterogeneous landscapes that appear more often in the high-resolution PlanetScope image (as Fig.
162 S1). For this, here we test whether adaptively selecting either single or two reference images on
163 an object-class basis rather than at the cloud patch level can help improve gap-filling performance
164 by reducing the uncertainties associated with both spatial heterogeneity and diverse (with/without
165 rapid) land cover change scenarios.

166 The goal of this study thus aims to develop an accurate and robust gap-filling method to
167 automatically reconstruct missing data in PlanetScope image time series. Given the challenges in
168 gap-filling PlanetScope satellites, we highlighted two major novelties of this work as follows.
169 First, we employed an object-based segmentation method in conjunction with an unsupervised
170 classification method to effectively reconstruct the missing information regardless of image gap
171 size (corresponding to the number of image pixels contaminated by clouds/cloud shadows).
172 Second, for each object-class (i.e. a group of pixels of the same object and the same class), we
173 adopted an adaptive method to automatically select one or two reference images from the image
174 time series for gap-filling, and assigned the best guess of the temporal change scenarios

175 (with/without rapid land cover change) based on their temporal changing patterns. We named this
176 method the object and class based gap-filling (OCBGF) method. With this new method, we hope
177 to effectively gap-fill missing pixels in PlanetScope image time series and enhance the data
178 availability and continuity for fine-scale land surface monitoring. As a proof-of-concept, we
179 selected the four sites with different land cover types covering a variety of climate zones on the
180 global land surface, and compared OCBGF with three state-of-the-art methods, NSPI, ARRC, and
181 SAMSTS.

182 **2. Study sites and materials**

183 **2.1 Study sites**

184 To test the OCBGF method, we selected four sites that are representative of different land cover
185 types and climate zones across the global land surface (Fig. 1). These include 1) a managed forest
186 landscape from a Eucalyptus plantation site in South Brazil (henceforth ‘Euc-plantation’), 2) a
187 cropland landscape from Iowa in the United States (henceforth ‘Iowa-cropland’), 3) a metropolitan
188 urban area from Beijing city in China (henceforth ‘Beijing-urban’), and 4) a moist forest landscape
189 from the Barro Colorado Island in Panama (henceforth ‘BCI-forest’). Details regarding the
190 location, climate, seasons, and spatial extent of these four sites are shown in Table 1.

191 We selected these study sites for two reasons. First, these four sites represent different land cover
192 types and climate zones with various levels of landscape heterogeneity and span a large range of
193 annual precipitation from 492-2052 mm per year (Table 1). For example, the Beijing-urban site
194 contains a large number of built-ups, such as buildings, highways, airport, as well as vegetated
195 areas in north-eastern suburbs; the Iowa-cropland site is typical heterogeneous cropland with
196 various crops types, such as corn, soybeans, and oats; the BCI-forest site includes a moist forest
197 mixed with deciduous and evergreen broadleaf species, roads, buildings, and bare soils in
198 surrounding areas of the Barro Colorado Island; and the Euc-plantation site has mixed land cover
199 types, including commercial Eucalyptus, forest patches, buildings, roads, and bare soils. In
200 addition, among these sites, the BCI-forest site has a tropical climate with the most serious cloud
201 contamination resulting in continuous data missing in a year, especially during the wet season with
202 monthly rainfall higher than 100 mm.

203 Second, there is modest to strong temporal variability in surface reflectance at these sites. For most
204 gap-filling methods, large temporal variations in surface reflectance caused by plant phenology
205 and abrupt land cover change could cause uncertainties (Cao et al., 2020). The Beijing-urban site
206 has experienced substantial urbanization in recent years, converting vegetated surfaces into
207 buildings (Cao et al., 2020). The Iowa-cropland site has a significant crop rotation between corn
208 and soybean and vegetation phenology changes, such as seeding, growing, and harvesting (Cao et
209 al., 2020). The BCI-forest site exhibits modest to large seasonal reflectance variability with
210 significant amounts of leaf shedding and leaf exchange during the high-light dry season (Detto et
211 al., 2018; Park et al., 2019). The Euc-plantation site experiences strong seasonal reflectance
212 changes caused by harvesting, and land cover changes caused by deforestation and reforestation
213 (Lopez-Poma et al., 2020; Qin et al., 2019).

214 For more details about the ecological, hydrological, and topographic characteristics of these four
215 sites, refer to previous studies (Campoe et al., 2012; Leigh, 1999; Liu et al., 2018; Qi et al., 2011).

216 **2.2 Materials**

217 The four-band, 3-m resolution PlanetScope data from Planet Labs PBC. (San Francisco, CA, USA)
218 were used in this study. We accessed the data from <https://www.planet.com/> through a research
219 and education license with Planet Labs PBC. We downloaded the PlanetScope data for all four
220 sites that span the whole annual cycle covering a wide range of percent data missing for each site
221 (Fig. 1c). The level 3B surface reflectance product of PlanetScope was used, which has been
222 orthorectified and pre-processed (including geometric, radiometric, and atmospheric corrections)
223 (Planet, 2021). To retain as much good data as possible, we assessed all available data that met the
224 following criteria: 1) "standard" quality level (that refers to an image meeting a variety of quality
225 standards; Planet, 2021), 2) rectification with ground control points, 3) solar zenith angle $< 80^\circ$
226 and view zenith angle $< 5^\circ$, 4) snow cover $< 5\%$, 5) cloud cover $< 80\%$, 6) thin cloud cover $< 5\%$.
227 Consequently, across the four sites throughout the entire annual cycle of 2018, a total of 271 days
228 of PlanetScope land surface reflectance images were accessed (Table 1). Each of these images is
229 from the sensor type of PS2 (denoting Dove Classic) and has a spatial extent of $20\text{km} \times 20\text{km}$
230 (6667×6667 pixels). Notably, since the BCI-forest site has some water surfaces occupied by lakes
231 and rivers, prior to testing our method, we masked out these elements in the corresponding

232 PlanetScope images using the same masking method as Wang et al. (2021). This was conducted
233 to minimize the gap-filling uncertainty associated with large spectral variability with the water
234 surfaces across both space and time (Qiu et al., 2019; Zhu et al., 2015; Zhu and Woodcock, 2012)
235 and misclassification error encountered in cloud and shadow detection on water surfaces (Zhu and
236 Helmer, 2018).

237 **3. Methods**

238 We applied the OCBGF method for gap-filling PlanetScope time-series images using one target
239 image (I_{t_0} , where t_0 is the acquisition date of the target image) in this time series as an example.
240 The same procedure also applies to any other images in the time series that need to be gap-filled.
241 The time-series images can be fully or partly cloud-free (e.g. percent data missing ranging from 0-
242 100%; Fig. 1c), and OCBGF automatically selects those image pixels in the time series with valid
243 values at or nearby the locations of gaps to gap-fill I_{t_0} . Particularly, we assumed that pixels with
244 both high spectral similarity and spatial continuity tend to have similar temporal changing patterns
245 (Chen et al., 2011; Zhu et al., 2012b). Based on this assumption, we first developed a model to
246 capture the temporal changing patterns using spectrally and spatially similar pixels outside the
247 gaps, and then applied this model to gap-fill the missing pixels in I_{t_0} . There are four tasks in
248 OCBGF for gap-filling (Fig. 2). First, we conducted pixel-level quality control for each image in
249 the time series to minimize potential cloud and cloud shadow impacts. Second, before gap-filling,
250 we integrated a cloud-free time series (I_{t_s}) covering around a one-month time period with the
251 minimum data missing nearby each I_{t_0} , then segmented and classified I_{t_s} into independent objects
252 of different classes. Third, for each object-class (i.e. a group of pixels of the same object and the
253 same class) with missing pixels in I_{t_0} , we performed gap-filling with two scenarios: single-
254 reference-image (no rapid land cover change) and two-reference-images (rapid land cover change)
255 scenarios. Finally, we conducted post-image-processing with a guided filtering approach to further
256 reduce random noises in the gap-filled images while retaining detailed information.

257 **3.1 Pixel-level quality control (Task 1)**

258 Before gap-filling, strict pixel-level quality control is essential (Chen et al., 2004). For this, we
259 adopted a two-step approach. First, we applied a recently developed automatic cloud/cloud shadow
260 detection method, STI-ACSS (Wang et al, 2021), for initial and automatic screening of cloud/cloud

261 shadow across the entire image time series. The STI-ACSS method was used because it has been
262 demonstrated to be more effective in cloud/cloud shadow screening in PlanetScope images
263 compared with the default PlanetScope quality control layers and other state-of-the-art cloud/cloud
264 shadow methods (Wang et al., 2021). Second, since no cloud/cloud shadow detection algorithm is
265 perfect, to minimize the uncertainty associated with the residual effects in STI-ACSS, we applied
266 additional pixel-level quality control to automatically mask out any remaining cloud/cloud shadow
267 pixels with two sub-steps: (1) to minimize the potential effects associated with thin clouds/cloud
268 shadows surrounding the detected clouds/cloud shadows, we used morphological dilation with a
269 structure element on each cloud and cloud shadow mask to label the border areas of clouds/cloud
270 shadows (Soille, 1999; Soille and Pesaresi, 2002). The structure element is a disk-shaped matrix
271 (Zheng, 1995) with a user-specified size (5×5 pixels in this study); and (2) to best filter remaining
272 cloud/cloud shadow pixels that have larger temporal variability in their spectral reflectance than
273 clear-sky pixels (Wang et al., 2021), we examined their band-specific pixel values with a user-
274 specified threshold pair (e.g. 1 and 99 percentiles) across the entire image time series and masked
275 out the pixels with values either greater than the upper bound (e.g. 99 percentile) or smaller than
276 the lower bound (e.g. 1 percentile). More details regarding the three parameters used in this task,
277 i.e., the size of the structure element, and a pair of percentile thresholds for additional data quality
278 control, the way to determine their values, and the values of recommendation are summarized in
279 Table S1.

280 **3.2 Object segmentation and classification (Task 2)**

281 Because the natural landscape is often mixed with multiple objects within the same class or
282 multiple classes belonging to the same object like the example shown in Fig. S1, the conventional
283 approaches that either rely on any given moving window or are based on the object-based image
284 segmentation are not sufficiently accurate to address this issue. Thus, here we developed a new
285 method relying on the concept of object-class, and also compared this new method with the
286 conventional approaches that relied on either object or class alone. The new object-class method
287 integrated an object-based segmentation method with an unsupervised classification method for
288 automatic identifications of pixels, by which it helps to group those spectrally and spatially similar
289 pixels into the same object and the same class.

290 Specifically, we first used an object-based image analysis (OBIA) method as implemented by
291 Watkins and Van Niekerk (2019a, 2019b) to segment all pixels in a composited cloud-free time
292 series (I_{ts}) into individual objects with irregular sizes based on their spectral and spatial
293 characteristics. This OBIA method was adopted because of its high accuracy in detecting object
294 boundaries as well as its capability to operate without any prior knowledge (Watkins and Van
295 Niekerk, 2019a). This method includes the following four steps, through which we generated a
296 segmentation map (S_{ts}) for I_{ts} :

297 (Step 1) We generated a composited cloud-free time series I_{ts} for automatic image segmentation.
298 For this, the 10 temporally-adjacent images (i.e. 5 images before/after I_{t_0}) were automatically
299 selected from the whole image time series, with which we further derived I_{ts} by identifying the
300 clearest images with minimum missing pixels from each 5-image group (Fig. S2). Each 5-images
301 before and after I_{t_0} were selected because this roughly represents a short temporal period of one
302 month in PlanetScope images (Roy et al., 2021). The two selected clearest images (I_{ts} ; $n=2$, with
303 one before I_{t_0} and one after I_{t_0}) were subsequently used for image automatic segmentation as
304 described in steps 2-4. It is noted that i) when there are more than one clearest images within each
305 5-image group, the image with the least temporal distance to I_{t_0} is selected; and ii) these two
306 clearest images both need to cover the full area of I_{t_0} ; otherwise, we would use the clearest image
307 across the entire time series. It is also important to note that the approach used here is empirical
308 and usable for PlanetScope satellites of high temporal resolution, but might need some fine-tuning
309 when applying to other satellites of different temporal resolutions.

310 (Step 2) We employed an image high-pass filtering approach with a 4-neighborhood Laplacian
311 filter (Gonzalez and Woods, 2006; Solomon and Breckon, 2010) for each image in I_{ts} , through
312 which we could include more spatial details to facilitate the subsequent image segmentation.

313 (Step 3) We applied a commonly-used Canny edge-detection operator (Canny, 1986) on each
314 sharpened image in I_{ts} and aggregated the edge layers with a union operation to generate one
315 composite edge layer for image segmentation.

316 (Step 4) We employed a widely-used region-based segmentation approach, watershed
317 segmentation (Li et al., 2010), on the above-derived edge layer for deriving S_{ts} (Fig. 2). This
318 approach divides regions of local minima (catchment basins) into individual objects based on

319 edges with high gradient magnitudes (Salman, 2006). This watershed segmentation approach used
320 8-connected connectivity to specify the directions of adjacent pixels in the neighborhood of a given
321 pixel.

322 After deriving individual objects using the above image segmentation method, we employed a
323 commonly-used unsupervised classifier, k -means (Lloyd, 1982), on I_{ts} . The k -means classifier
324 automatically classifies all valid pixels of I_{ts} into K classes based on their spectral similarity, and
325 generates a classification map C_{ts} (Fig. 2). The classifier minimized the sum of the squared
326 Euclidean distance of spectral reflectance between each pixel and the class centroid to estimate
327 classification results. There is only one parameter, K (the number of classes), in this unsupervised
328 classification. To determine K , we specified a range of K values (5-10 in our study) and then
329 followed the Calinski-Harabasz clustering evaluation criterion (Calinski and Harabasz, 1974) to
330 automatically determine the optimal K within this range. A further sensitivity analysis based on
331 the setting of different K ranges (Fig. S3) confirms that the K range set for this study is valid. It is
332 worthy to note that the setting of K range can vary with the level of landscape heterogeneity, thus
333 we included this recommendation in Table S1 for the readers' reference.

334 With the S_{ts} and C_{ts} derived above, we finally identified the object and class type for each pixel,
335 and grouped the pixels belonging to the same object and the same class as a representative gap-
336 filling unit of 'object-class'.

337 **3.3 Gap-filling with two scenarios (Task 3)**

338 Since rapid land cover change could also introduce uncertainty into gap-filling, we first
339 differentiated the two land cover change scenarios (with or without rapid land cover change) and
340 then adopted scenario-specific gap-filling procedures. These two steps were conducted on an
341 object-class basis as follows.

342 **3.3.1 Differentiations of two land cover change scenarios**

343 For each given object-class with missing pixels in I_{t0} , we divided them into the following two
344 cases: (1) only part of the pixels were missing, and (2) all pixels were missing.

345 For case 1, we first searched each target object-class throughout the full image time series, and
 346 then identified the reference images that had valid pixels for gap-filling. For each object-class,
 347 based on their temporal distances to I_{t_0} , we further determined the reference images with the
 348 closest temporal distance before I_{t_0} (noted as $I_{t_{0-1}}$) and after I_{t_0} (noted as $I_{t_{0+1}}$) as well as the
 349 image with the closest absolute temporal distance to I_{t_0} (denoted as I_{t_n} , which could be either
 350 $I_{t_{0-1}}$ or $I_{t_{0+1}}$; Fig. 2). For case 2, we turned to the neighboring object-class that had valid pixels
 351 in I_{t_0} , belonged to the same class, and was spatially closest to the target object-class, and repeated
 352 the same process as case 1, through which we identified the corresponding reference images of
 353 $I_{t_{0-1}}$, $I_{t_{0+1}}$, and I_{t_n} . Since the reference image(s) were selected on an object-class basis, our
 354 approach can thus optimize the temporal distance between reference image(s) and the target image,
 355 consequently reducing the gap-filling uncertainty associated with the landscape heterogeneity and
 356 rapid land cover changes over time.

357 With the identified I_{t_n} , we then calculated the correlation coefficient for each spectral band
 358 between all the valid pixels of I_{t_0} and I_{t_n} that either belongs to the target object-class (for case 1)
 359 or the nearest adjacent object-class (for case 2). We further calculated the average correlation
 360 coefficient across all four spectral bands (\bar{r}_h) and assigned the land cover change scenarios by
 361 comparing \bar{r}_h with an empirically determined threshold (r_T) as Eq. 1.

$$362 \quad \begin{cases} \bar{r}_h \geq r_T, \text{ no rapid land cover change scenario} \\ \bar{r}_h < r_T, \text{ rapid land cover change scenario} \end{cases} \quad (1)$$

363 To assess how the threshold value (r_T) would affect our results, we varied r_T from 0.65 to 0.95
 364 with an interval of 0.05. Additionally, to assess how our method would perform in these two
 365 independent scenarios in Eq. 1, we also separately tested these two scenarios. The sensitivity
 366 analysis across four sites (Figs. S4 and S5) demonstrated that using a fixed threshold of 0.80 across
 367 all sites obtains the highest averaged accuracies, and is also quite comparable with the site-specific
 368 optimized threshold with a very narrow range of 0.75~0.80 across all four sites. Details regarding
 369 how to determine r_T as well as the recommended value of r_T for use are summarized in Table S1.

370 **3.3.2 Scenario-specific gap-filling**

371 **Scenario 1 – Single-reference-image**

372 For each object (h)-class (k), under the no rapid land cover change scenario, we used only one
 373 image I_{tn} for gap-filling. This single-image referencing method relies on a linear regression model
 374 between I_{t0} and I_{tn} :

$$375 \quad I_{t0,b,i} = \alpha_{hk,b} \times I_{tn,b,i} + \beta_{hk,b} + \varepsilon_{hk,b} \quad (2)$$

376 where $I_{t0,b,i}$ and $I_{tn,b,i}$ are spectral reflectance values of the valid pixel i in band b for I_{t0} and I_{tn} ,
 377 respectively, and $\alpha_{hk,b}$ and $\beta_{hk,b}$ are coefficients of this linear model for band b , and $\varepsilon_{hk,b}$ is the
 378 residual error. The band-specific coefficients can be solved by linearly regressing the spectral
 379 reflectance values of all the valid pixels between I_{t0} and I_{tn} .

380 Before gap-filling, we further assessed the type of any given object-class into the following two
 381 categories: 1) if only part of the pixels within the target object-class were missing, we labelled
 382 them as category 1, by which we applied Eq. 2 to the remaining valid pixels of the same object-
 383 class to first derive the linear regression model, and then applied this model to gap-fill the missing
 384 ones using Eq. 3; and 2) if all the pixels within the target object-class were missing, we labelled
 385 them as category 2, by which we applied Eq. 2 to those valid pixels of the same class available
 386 from the nearest adjacent object to derive the linear model, and then applied the derived model to
 387 gap-fill those missing ones using Eq. 4 below.

$$388 \quad \widehat{I_{t0,b,j}} = \alpha_{hk,b} \times I_{tn,b,j} + \beta_{hk,b} \quad (3)$$

389 Where $\widehat{I_{t0,b,j}}$ is the predicted reflectance value of missing pixel j in band b for I_{t0} , $I_{tn,b,j}$ is the
 390 spectral reflectance value of valid pixel j in band b for I_{tn} , and $\alpha_{hk,b}$ and $\beta_{hk,b}$ are estimated
 391 coefficients using Eq. 2 for the category 1.

$$392 \quad \widehat{I_{t0,b,j}} = \alpha_{ck,b} \times I_{tn,b,j} + \beta_{ck,b} \quad (4)$$

393 Where $\widehat{I_{t0,b,j}}$ is the predicted reflectance value of missing pixel j in band b for I_{t0} , $I_{tn,b,j}$ is the
 394 spectral reflectance value of valid pixel j in band b for I_{tn} , and $\alpha_{ck,b}$ and $\beta_{ck,b}$ are estimated
 395 coefficients using Eq. 2 for the category 2, which represents the closest object (c) to the target
 396 object (h) with valid pixels in both I_{t0} and I_{tn} for class (k).

397 **Scenario 2 – Two-reference-images**

398 For each object (h)-class(k), under the rapid land cover change scenario, we used the two images
 399 (one before I_{t_0} , I_{t_0-1} , and one after I_{t_0} , I_{t_0+1}) for gap-filling. This two-image referencing method
 400 relies on a linear regression model between I_{t_0} and the two reference images (Eqs. 5). The key
 401 assumption of this method is that the use of a pair of reference images would be more efficient to
 402 capture the rapid land cover change compared with the use of a single reference image for gap
 403 filling, especially when the target image is different from either reference image (i.e. $\bar{r}_h < r_T$).

$$404 \quad I_{t_0,b,i} - I_{t_0+1,b,i} = \alpha'_{hk,b}(I_{t_0-1,b,i} - I_{t_0+1,b,i}) + \beta_{hk,b} + \varepsilon_{hk,b} \quad (5)$$

405 where $I_{t_0,b,i}$, $I_{t_0-1,b,i}$, and $I_{t_0+1,b,i}$ are spectral reflectance values of valid pixel i in band b for I_{t_0} ,
 406 I_{t_0-1} and I_{t_0+1} , respectively, and $\alpha'_{hk,b}$ and $\beta_{hk,b}$ are coefficients of this linear model in band b ,
 407 and $\varepsilon_{hk,b}$ is the residual error. The band-specific coefficients can be solved by linearly regressing
 408 the spectral reflectance values of all the valid pixels between $I_{t_0} - I_{t_0+1}$ and $I_{t_0-1} - I_{t_0+1}$.

409 Similar to scenario 1, before gap-filling, we further assessed the type of any given object-class into
 410 the following two categories: 1) if only part of the pixels within the target object-class were missing,
 411 we labelled them as category 3, by which we applied Eq. 5 to the remaining valid pixels of the
 412 same object-class to first derive the linear regression model, and then applied this model to gap-
 413 fill those missing ones using Eq. 6 below; and 2) if all the pixels within the target object-class were
 414 missing, we labelled them as category 4, by which we applied Eq. 5 to those valid pixels of the
 415 same class available from the nearest adjacent object to derive the linear model, and then applied
 416 the derived model to gap-fill those missing ones using Eq. 7 below.

$$417 \quad \widehat{I_{t_0,b,j}} = \alpha'_{hk,b} \times (I_{t_0-1,b,j} - I_{t_0+1,b,j}) + I_{t_0+1,b,j} + \beta_{hk,b} \quad (6)$$

418 Where $\widehat{I_{t_0,b,j}}$ is the predicted reflectance value of missing pixel j in band b for I_{t_0} ,
 419 $(I_{t_0-1,b,j} - I_{t_0+1,b,j})$ is the spectral reflectance difference of valid pixel j in band b between the
 420 two reference images, and $\alpha'_{hk,b}$ and $\beta_{hk,b}$ are estimated coefficients using Eq. 5 for the category
 421 3.

$$422 \quad \widehat{I_{t_0,b,j}} = \alpha'_{ck,b} \times (I_{t_0-1,b,j} - I_{t_0+1,b,j}) + I_{t_0+1,b,j} + \beta_{ck,b} \quad (7)$$

423 Where $\widehat{I_{t_0,b,j}}$ is the predicted reflectance value of missing pixel j in band b for I_{t_0} ,
424 $(I_{t_0-1,b,j} - I_{t_0+1,b,j})$ is the spectral reflectance difference of valid pixel j in band b between the
425 two reference images, and $\alpha'_{ck,b}$ and $\beta_{ck,b}$ are estimated coefficients using Eq. 5 for the category
426 4, which represents the closest object (c) to the target object (h) with valid pixels for class (k) in
427 I_{t_0} .

428 **3.4 Post-image-processing (Task 4)**

429 To further reduce random noises (e.g. salt-and-pepper noises; Fig. S6b) while retaining high-
430 resolution details in the gap-filling results derived above, we conducted post-image-processing for
431 the target image using a commonly-used edge-preserving filter, the guided filter (He et al., 2013;
432 Fig. S6c), with a demonstrated improvement in the gap-filling results (Fig. S7). The guided filter
433 computes the filtering output by a linear transformation of a guidance image, but does not suffer
434 from the gradient reversal artefacts (He et al., 2013; Li et al., 2017). Specifically, a cloud-free
435 image or gap-filled image with the maximum number of valid pixels across the full image time
436 series is selected as the guidance image for the subsequent guided filter on the target image.

437 **3.5 Evaluation**

438 To evaluate the performance of our method, we performed three tests. The first two tests were
439 based on simulation experiments, and the third test used real cloud contaminated images.
440 Throughout all these three tests, we also cross-compared our OCBGF model performance with the
441 other three state-of-the-art methods, i.e. NSPI (Zhu et al., 2012a), ARRC (Cao et al., 2020), and
442 SAMSTS (Yan and Roy, 2018).

443 In test 1, we evaluated the effects of gap size on the gap-filling results by randomly selecting a
444 fully or near-fully clear-sky image from each study site and artificially creating seven gaps ranging
445 from 10^4 to 10^6 pixels (10^5 to 10^7 m²) on each image (Table 2). Since the percent data missing of
446 the above test remains small (e.g. within 20%), to further evaluate the boundary beyond which
447 level of percent data missing would largely affect the gap-filling results among different methods,
448 we used the Euc-plantation and Iowa-cropland sites as examples, and artificially created six circle
449 gaps sharing the same center (as the image center) but having a radius range from 500 to 3000

450 pixels with an increment of 500 pixels. These six gaps corresponded to the percent data missing
451 of 2%, 7%, 16%, 28%, 44%, and 64%, respectively, on the selected images.

452 In test 2, we evaluated the effects of rapid land cover change on the gap-filling results by selecting
453 24 fully or near-fully clear-sky images (10 from Euc-plantation, 7 from Iowa-cropland, 4 from
454 Beijing-urban, and 3 from BCI-forest sites). These images were selected based on the fact that
455 there were obvious land cover changes between the temporally adjacent images and the images
456 covered different seasons of the year 2018. With these images, we then purposely masked out the
457 areas that experienced land cover changes over time.

458 In both test 1 and 2, we used the original clear images as benchmark. The model performance was
459 assessed on a spectral band basis, using the following three metrics, i.e., the root mean square error
460 (RMSE), the correlation coefficient (CC), and the structure similarity index (SSIM) (Wang et al.,
461 2004). We used these three metrics in combination, as each of them reflects different aspects of
462 model performance assessments, with RMSE assessing the spectral band differences between the
463 predicted and reference images, CC reflecting the degree of correlation and similarity between the
464 predicted and reference images, and SSIM quantifying the structural similarities between the
465 predicted and reference images.

466 In test 3, we randomly selected one cloud contaminated image from each study site, and visually
467 assessed the structural continuity and color consistency by comparing the gap-filled areas with
468 those clear areas spatially/temporally adjacent to the gaps. For cross-method comparisons, we used
469 the default parameters of NSPI, ARRC, and SAMSTS, following Zhu et al. (2012a), Cao et al.
470 (2020), and Yan and Roy (2018), respectively.

471 **4. Results**

472 **4.1 Cross-method comparison for different simulated gap sizes**

473 To assess the effects of gap size on the gap-filled results for all four methods (NSPI, ARRC,
474 SAMSTS, and OCBGF), we conducted accuracy assessments on seven deliberately created gaps
475 of different size categories using the original clear images as benchmark. Across all four sites, we
476 found that OCBGF yielded the highest accuracy with the smallest variation across seven gap size

477 categories (RMSE=0.0065 (mean) \pm 0.0027 (standard deviation), CC=0.95 \pm 0.06, and SSIM=
478 0.94 \pm 0.06; Fig. 3 and Table 3). By contrast, the NSPI method generated the lowest model
479 accuracies with the largest variations across different gap size categories (RMSE=0.0113 (mean)
480 \pm 0.0058 (standard deviation), CC=0.87 \pm 0.15, and SSIM= 0.86 \pm 0.16; Fig. 3 and Table 3). Among
481 these four sites, OCBGF obtained much higher accuracies compared to the other three methods,
482 consistently across all gap size categories in Euc-plantation and BCI-forest (Fig. 3); OCBGF
483 achieves a comparable accuracy in Iowa-cropland and Beijing-urban as ARRC, and in Iowa-
484 cropland as SAMSTS, all of which are higher than NSPI (Fig. 3). Moreover, within each site and
485 across all three accuracy metrics, we observed that OCBGF yields more stable model accuracies
486 across all seven gap size categories than the other three models (Fig. 3). Especially in Euc-
487 plantation, Beijing-urban, and BCI-forest, the three accuracy metrics of OCBGF tend to stabilize
488 with gap size while NSPI and SAMSTS display larger accuracy fluctuations across gap sizes, with
489 large near-linear reductions in model accuracies with gaps size in Beijing-urban and BCI-forest.
490 The stability in model accuracies of ARRC with gap size is comparable with OCBGF in Euc-
491 plantation, Iowa-cropland, and Beijing-urban, but shows much larger fluctuations with a near-
492 linear reduction in model accuracies with gaps size in BCI-forest. These results altogether
493 demonstrate that OCBGF is the most accurate method showing the least sensitivity to the gap size
494 among all comparative methods, followed by SAMSTS, ARRC, and NSPI (Table 3).

495 Among the three accuracy metrics across all four methods, we observed a much lower CC and
496 SSIM at the BCI-forest site relative to the other three sites (Fig. 4). This is likely associated with
497 the low data availability at this site with a high-rainfall environment (Fig. 1c and Table 1).
498 Meanwhile, we observed that all the four methods maintain comparable CC and SSIM in the other
499 three lower-rainfall sites (i.e. Euc-plantation, Iowa-cropland, and Beijing-urban; Table 1), and the
500 accuracy reduction of CC and SSIM from these drier sites to the high-rainfall BCI-forest site is
501 much smaller in OCBGF than the other three methods, demonstrating that OCBGF could be more
502 robust than the other three methods in gap-filling, especially in the high rainfall environment.

503 To further aid the visual interpretation of these cross-method comparisons, we next presented the
504 gap-filling results (Fig. S8) and associated magnified correspondences (Fig. 5) for each of the four
505 sites. Across all three methods, we observed that the gap-filling results are overall very comparable
506 in the smaller gap size categories from 1 to 5 (Fig. S8), but tend to diverge in the remaining two

507 bigger gap size categories (Figs. S8 and 5), with the OCBGF results being the closest to original
508 images in larger gaps. Also in Fig. 5, NSPI displays some blocking artifacts (Figs. 5a, b, and d) in
509 the interior areas of Euc-plantation, Iowa-cropland, and BCI-forest, while generating some strip
510 artifacts (Figs. 5c and e) in the heterogeneous areas of Beijing-urban and BCI-forest. ARRC
511 exhibits obvious salt-and-pepper noises (Figs. 5f-h) in heterogeneous areas mixed with different
512 land covers across all four sites. SAMSTS also presents some blocking artifacts (Fig. 5i),
513 accompanied by moderate errors in gap-filling land surface areas of high reflectance values (Figs.
514 5j and k). In contrast, OCBGF achieves a more satisfying performance across all these large gaps,
515 despite small portions of fine details missing as compared to the original images (Fig. 5).
516 Collectively, both quantitative and visual cross-method comparisons show that OCBGF is more
517 accurate and stable across all land cover types and gap size categories than the other three methods.

518 Finally, to evaluate whether a large percent of data missing (e.g. >30%) would affect the gap-
519 filling results among different methods, we conducted accuracy assessments on the six artificial
520 circle gaps covering the percent data missing ranging from 2%, 7%, 16%, 28%, 44%, and 64%
521 (Fig. 6a). Our results show that OCBGF consistently yields the best and most stable model
522 performance across the full range of percent data missing (Figs. 6b and c), in contrast with the
523 other three models that display much weaker and more variable model accuracies across different
524 percent data missing. For example, when the percent data missing is relatively low (2% and 7%),
525 ARRC and SAMSTS obtain higher accuracies across all three accuracy metrics. With a further
526 increase in percent data missing (16% and 28%), SAMSTS and NSPI obtain higher accuracies in
527 CC and SSIM, while NSPI generates lower accuracy in RMSE. When the percent data missing is
528 much higher (44% and 64%), NSPI generates higher accuracies in CC and SSIM but a lower
529 accuracy in RMSE. Our proposed method of OCBGF performs best consistently and most stable
530 across all the three accuracy metrics, with only a very minor accuracy reduction with an increasing
531 percent data missing.

532 **4.2 Cross-method comparison for rapid land cover changes**

533 To examine the effects of rapid land cover changes on the gap-filled results for all four methods,
534 we turned to the simulation approach, and conducted accuracy assessments on 24 representative
535 images using original clear images as benchmarks. Across all sites and seasons, we found that

536 OCBGF outperforms the other three methods (Fig. 7), with the highest accuracies and least
537 accuracy variations ($RMSE = 0.0082$ (mean) ± 0.0023 (standard deviation), $CC = 0.95 \pm 0.031$, and
538 $SSIM = 0.94 \pm 0.035$; Table 4), followed by SAMSTS and ARRC that performs slightly better than
539 NSPI but with much larger variations (Table 4). Among these four sites and seasons, we also
540 observed that OCBGF obtains the consistently highest accuracies and lowest variations across all
541 four sites (Fig. 8). While for ARRC, SAMSTS, and NSPI, we found that these three methods have
542 mixed successes, with ARRC obtaining the lowest accuracies in the BCI-forest site, NSPI
543 obtaining the lowest accuracies in most cases of the Iowa-cropland site, and SAMSTS obtaining
544 the lowest accuracies in Beijing-urban site.

545 To visually interpret these cross-method comparisons further, we presented the two magnified
546 regions with rapid land cover changes for each site using their original clear images as benchmarks
547 (Fig. 9). Specifically, there are two types of rapid changes being selected: 1) gradual color
548 transformation (e.g. Figs. 9b, h, and i) and 2) abrupt color transformations (e.g. Figs. 9a, c, d, e, f,
549 and g). To put the gap-filling results in the time-series context, we displayed the two images (before
550 and after) temporally adjacent to each target image. Across the cases showing a gradual color
551 transformation (Figs. 9b, h, and i), our results show that the gap-filling results from OCBGF are
552 closer to the benchmarks than the other three methods, in which NSPI yields some blurred colors,
553 ARRC gives some salt-and-pepper noises, while SAMSTS exhibits some obvious errors around
554 the heterogeneous areas with complex land covers. Moreover, we also examined the cases with
555 abrupt color transformations across the image time series (Figs. 9a, c, d, e, and f). Across these
556 cases, we found that OCBGF is consistently more efficient in capturing the rapid land cover
557 changes with well-constructed spatial details compared with the other three methods. However,
558 we also found that there remains a small portion (Fig. 9g) of areas in Beijing-urban being
559 inaccurately reconstructed by all four methods. Compared with OCBGF, NSPI, ARRC, and
560 SAMSTS are found to have even larger portions of areas being inaccurately reconstructed,
561 especially for those heterogeneous areas that often come with obvious errors/noises. These results
562 altogether demonstrate that OCBGF performs the best in gap-filling those areas with rapid land
563 cover changes among all four methods examined here.

564 **4.3 Cross-method comparison in the real-world practices**

565 To assess the performance of these methods in real-world scenarios, we selected one representative
566 image mixed with cloud contaminations for each of the four sites. Our results show that OCBGF
567 achieves the best model performance with the least sensitivity to clouds/cloud shadows in general
568 (Fig. 10). Particularly, the magnified gap-filling results (highlighted in orange squares; Fig. 10)
569 show that OCBGF obtains the highest consistency with the temporally most adjacent images while
570 displaying the best spatial continuity near gap boundaries relative to the other three methods. In
571 contrast, NSPI exhibits blurred colors, especially in large gaps, ARRC exhibits obvious noises and
572 errors in heterogeneous areas, and SAMSTS displays imprecise boundaries, especially at the
573 intersections among different land cover types. These results again demonstrate that OCBGF
574 generates the best gap-filling results, with potential to best recover the signal for those areas with
575 big data gaps and rapid land cover changes.

576 **5. Discussion**

577 Over recent years, PlanetScope data has been increasingly used for monitoring rapid and fine-scale
578 land surface dynamics that can scale up to create impacts on understanding global environmental
579 change and ecosystem responses (Forzieri et al., 2021; Heinrich et al., 2021; Taubert et al., 2018;
580 Zeng et al., 2018). However, a critical challenge remains with the lack of an effective and accurate
581 method for gap-filling missing data caused by cloud and cloud shadow contaminations. For
582 example, as shown in a 50-ha plot at the Panamanian BCI-forest site (Fig. S9), the original
583 PlanetScope time-series images are not able to accurately quantify leaf phenology at the patch
584 scale of a 12 m×12 m area when they are seriously contaminated by clouds/cloud shadows, while
585 the gap-filled time-series images could help recover the considerable within-site fine-scale
586 phenology variability, suggesting the necessity of gap-filling. To address this challenge, we
587 developed an object and class based gap-filling method, OCBGF, and evaluated the effectiveness
588 of this method across four sites spanning a large variety in land cover types (i.e. plantation,
589 cropland, urban, and forest; Fig. 1b), spatial heterogeneity (e.g. homogenous and heterogeneous
590 landscapes; Fig. 1b), annual precipitation (e.g. 492-2052 mm per year; Table 1), and percent data
591 missing (e.g. ranging from 0% to 100%; Fig. 1c). When comparing our OCBGF method with the
592 other three state-of-the-art methods (NSPI, ARRC, and SAMSTS) on both simulation tests and
593 real-world cloud-contaminated cases (Figs. 3-10, Tables 3 and 4), our results consistently show

594 that OCBGF is the most accurate and stable gap-filling method for recovering cloud/cloud shadow-
595 induced data gaps in PlanetScope time-series images.

596 The effectiveness of OCBGF relies on the following two strengths. First, to minimize the
597 uncertainty associated with large data gaps and landscape heterogeneity, OCBGF adopts an object-
598 based segmentation method in conjunction with a classification method to leverage valid pixels
599 outside the gaps for gap-filling. It is widely known that large data gaps (corresponding to a large
600 pixel number contaminated by clouds/cloud shadows) often result in insufficient valid pixels with
601 spectral similarity and spatial continuity, making the gap-filling difficult with large uncertainty,
602 especially for those interior pixels of large gaps (Shen et al., 2015; Yan and Roy, 2018). Meanwhile,
603 in high spatial resolution PlanetScope images, the landscape heterogeneity can be an increasingly
604 important issue, as there are often mixed classes within an object and mixed objects within a class
605 (e.g. Fig. S1), making the conventional gap-filling methods relying on either moving window or
606 object-based segmentation approaches challenging to resolve this issue. To address these issues,
607 similar to previous approaches of NSPI (Zhu et al., 2018) and ARRC (Cao et al., 2020), we also
608 employed an unsupervised classification method to classify the composited cloud-free time series
609 into different classes based on their spectral similarity. Different from those previous approaches
610 relying on a moving window to identify the valid pixels of the same class to gap-fill each target
611 pixel (Cao et al., 2020; Zhu et al., 2018), we integrated an object-based segmentation method with
612 an unsupervised classification method to identify the object and class type of pixels, by which we
613 searched valid pixels belonging to the same object-class or the nearest object-class to help gap-fill
614 each target object-class with missing pixels. This object-class integrated approach facilitates the
615 search for spectrally and spatially similar pixels, producing results with higher accuracies when
616 gap-filling missing pixels in general, and particularly when gaps are large (Figs. 3, 4, and 6). An
617 additional analysis (Fig. S10), showing that our object-class approach can generate more stable
618 and accurate gap-filling results compared with the other two approaches respectively relying on
619 object and class alone, further suggests that it can be a more accurate way to characterize the
620 landscape heterogeneity in the real world. Although SAMSTS applied a similar integration
621 framework—segment-and-clustering to gap-fill large gaps in Landsat images (Yan and Roy, 2018),
622 it is different from our method in two aspects: 1) SAMSTS used a region-growing image
623 segmentation method based on spectral similarity, while our method adopted an object-based
624 segmentation method based on both spectral characteristics and spatial texture; 2) SAMSTS

625 performed the classification task on derived segmented objects while our classification was
626 performed on a pixel level, which makes our method more accurate as there were often more than
627 one land cover types within a segmented object. Our cross-model comparisons (Figs. 4, 6, and 8)
628 also demonstrate that our method is superior to SAMSTS for gap-filling PlanetScope image time
629 series across all sites, seasons, data gap sizes, and land cover change scenarios. This study thus
630 represents a new attempt at using the object-class integration approach to effectively recover big
631 data gaps in high spatial-resolution satellite images.

632 Second, to minimize the gap-filling uncertainty associated with rapid land cover changes, OCBGF
633 adopts a scenario-specific gap-filling approach. With a 3-m spatial resolution, PlanetScope is able
634 to detect finer-scale changes, including rapid changes in land cover and land use types, than
635 traditional satellites (Cheng et al., 2020; Rao et al., 2021; Rasanen and Virtanen, 2019). However,
636 these fine-scale rapid changes also present a challenge due to the difficulty of identifying valid
637 pixels in heterogeneous landscapes suitable for gap-filling. To address this challenge, the selection
638 of the closest and clearest reference images has been increasingly recognized as an essential step
639 to gap-fill areas with rapid land cover changes (Cao et al., 2020). However, two issues remain in
640 the existing methods that rely on either single-reference-image or multiple-reference-image. First,
641 existing gap-filling methods often require one or more cloud-free reference images for the whole
642 image or cloud patch acquired close to the target image to aid gap-filling (Chen et al., 2011; Zhu
643 et al., 2012a), resulting in an extended temporal distance to the target image and consequently
644 introducing more uncertainty during gap-filling. Second, some of these methods commonly select
645 the reference images based on their spectral similarities with the target image on the scale of a
646 whole image or cloud patch, without considering the potential diverse land cover types underneath
647 the target cloud patch (Chen et al., 2017; Lin et al., 2013). To address these two issues, we also
648 turned to the reference image(s) but focused on the scale of each object-class, through which the
649 diversity in temporal changing patterns associated with different object-classes, even within the
650 same cloud patch, is accounted for. Meanwhile, our approach also minimized the temporal distance
651 between the reference image(s) and the target image, as the reference image(s) were now selected
652 at the object-class scale, which is much smaller than the scale of a whole image or cloud patch and
653 facilitates the search of the temporally nearest reference image(s). Additionally, to account for
654 rapid land cover changes, we adopted an extra step to first use a fixed threshold (Figs. S4 and S5)
655 to determine the scenarios with/without rapid cover changes for each target object-class, and then

656 adopted the scenario-specific gap-filling procedure, through which we aimed to further minimize
657 the gap-filling uncertainty associated with rapid land cover changes. Such a scenario-specific
658 approach integrating two different scenarios also generates consistently higher accuracies across
659 all four sites than independent scenarios (Fig. S4), demonstrating the effectiveness of including
660 this approach in OCBGF for gap-filling. As landscape heterogeneity is a common feature for high
661 spatial resolution satellite imagery and is becoming more common with increasing anthropogenic
662 activities, our successful implementation of OCBGF with higher accuracies than NSPI, ARRC,
663 and SAMSTS (Figs. 7 and 8) highlights the potential to extend this approach to other high spatial-
664 resolution satellite images in the future.

665 Our study also identifies three important next steps that need to be considered for future advances.
666 First, we observed that there remained some errors in OCBGF for gap-filling areas experiencing
667 rapid fine-scale changes with varying degrees of inconsistency in temporal changing patterns when
668 compared with valid pixels of the same or the nearest object-class (Fig. 9g). This is likely because
669 OCBGF gap-filled each target object-class with missing pixels by assuming that valid pixels of
670 the same or the nearest object-class tend to have similar temporal changing patterns to the missing
671 pixels, which would not work if the temporal changing patterns are inconsistent (Fig. 9g). A
672 potential way of solving this issue is by using deep learning based approaches, as several recent
673 studies have shown that they are effective at recognizing subtle patterns and characterizing
674 nonlinear relationships between the missing and valid pixels in time-series imagery (Li et al., 2019;
675 Zhang et al., 2020; Zi et al., 2021). Second, although OCBGF has demonstrated a higher accuracy
676 than the other three methods at the high-rainfall BCI-forest site (Figs. 4 and 8), it may encounter
677 challenges with extremely low availability of valid observations, such as in certain
678 tropical/subtropical areas which are subjected to persistent cloud cover at the month-to-year scale
679 (Wang et al., 2020, 2021). To investigate such impacts, we examined the sensitivity of OCBGF
680 on the frequency of low percent valid data (e.g. $< 10\%$) and the time interval of clear-sky
681 observations, respectively. Our results in Fig. S11 demonstrated that the model accuracy is stable
682 when the low percent valid data frequency is not more than 30% and then decreases afterward.
683 Meanwhile, we also observed that OCBGF yields stable model accuracies when the average time
684 interval of clear-sky observations is not more than 19 days and accuracy decreased linearly
685 afterward (Fig. S12). This high sensitivity to the frequency of valid observations is likely because
686 OCBGF relies on time-series images temporally adjacent to the target image for gap-filling, and

687 low valid data availability might lead to insufficient or no valid pixels to build models to capture
688 temporal changing patterns. We thus recommend using as many as available time-series images as
689 input to increase the valid-observation frequency. As for certain tropical/subtropical areas with
690 extremely low valid data availability, we recommend fusing optical PlanetScope with synthetic
691 aperture radar (SAR) data (Huang et al., 2015; Li et al., 2020b; Meraner et al., 2020; Pipia et al.,
692 2019) that are unaffected by cloud interference, which could be an important next step in the future.
693 Third, although all the data used in this study came from the PS2 sensor type, there are more and
694 more sensor types available in PlanetScope constellation, and the inconsistent radiometric
695 calibration across sensor types could also introduce uncertainties into the gap-filling results. To
696 investigate such impacts, we used the data in the year 2020 from the Iowa-cropland site that are
697 mixed with three different sensor types (PS2, PS2.SD, and PSB.SD) as an example, and addressed
698 the cross-sensor radiometric inconsistency issue using a cross-calibration approach developed
699 specifically for the PlanetScope constellation (Wang et al., 2020). Using the gap-filling results
700 from the calibrated PlanetScope as benchmarks, our results (Fig. S13) show that the raw and
701 calibrated gap-filled results are almost identical. These suggest that cross-sensor radiometric
702 inconsistency may only cause very minor differences in the gap-filling results, but a more
703 comprehensive assessment is still needed.

704 **6. Conclusions**

705 Here, we developed a new object and class based gap-filling (OCBGF) method for automatic gap-
706 filling of missing pixels in PlanetScope time-series images. This method integrates object
707 segmentation and classification for automatic identification of those spectrally and spatially similar
708 pixels, and uses a scenario-specific gap-filling procedure to recover missing pixels for each target
709 object-class in the image time series. The accuracy of OCBGF was evaluated at four contrasting
710 study sites spanning large gradients in land cover types (plantation, cropland, urban, and forest; all
711 excluding water bodies), spatial heterogeneity, annual precipitation, and percent data missing.
712 Relative to three other state-of-the-art gap-filling methods, NSPI, ARRC, and SAMSTS, OCBGF
713 obtained the highest accuracy regardless of simulation tests covering various gap sizes and rapid
714 land cover changes, and also achieved better performances for gap-filling cloud/cloud shadow-
715 induced data gaps in real-world practices. With these assessments, our results suggest that OCBGF
716 is an accurate and robust approach for gap-filling PlanetScope time-series images, and has

717 potential to be extended to other high spatial resolution satellites, e.g. Gaofen-1, Skysat, and
718 SPOT-7 (Gaofen-1, 2020; SkySat, 2020; SPOT-7, 2020), in the future attempts.

719

720 **Acknowledgements:** We would like to thank Helene C. Muller-Landau for sharing the drone data
721 collected over Barro Colorado Island, Panama (Fig. S9), which was collected by a team of
722 scientists, including Helene C. Muller-Landau, Jonathan Dandois, Milton Garcia, Stephanie
723 Bohlman, Sam Grubinger, and Raquel Araujo. The work was supported by National Natural
724 Science Foundation of China (#31922090). J. Wu was in part supported by the Innovation and
725 Technology Fund (funding support to State Key Laboratories in Hong Kong of
726 Agrobiotechnology) of the HKSAR, China. J. Wang was in part supported by the Division of
727 Ecology and Biodiversity PDF research award; C. Lee was in part supported by HKU 45th round
728 PDF scheme and HKU Seed Fund for Basic Research (#202011159154). The PlanetScope data
729 used in this study were accessed through the Education and Research Program, contracted between
730 Planet Labs PBC and HKU.

731 **References:**

- 732 Calinski, T., & Harabasz, J. (1974). A Dendrite Method for Cluster Analysis. *Communications in*
733 *Statistics - Theory and Methods*, 3, 1-27.
- 734 Campoe, O.C., Stape, J.L., Laclau, J.P., Marsden, C., & Nouvellon, Y. (2012). Stand-level patterns
735 of carbon fluxes and partitioning in a Eucalyptus grandis plantation across a gradient of
736 productivity, in Sao Paulo State, Brazil. *Tree Physiology*, 32, 696-706.
- 737 Canny, J. (1986). A Computational Approach to Edge-Detection. *IEEE Transactions on Pattern*
738 *Analysis and Machine Intelligence*, 8, 679-698.
- 739 Cao, R.Y., Chen, Y., Chen, J., Zhu, X.L., & Shen, M.G. (2020). Thick cloud removal in Landsat
740 images based on autoregression of Landsat time-series data. *Remote Sensing of*
741 *Environment*, 249, <https://doi.org/10.1016/j.rse.2020.112001>.
- 742 Case, N., & Vitti, A. (2021). Reconstruction of Multi-Temporal Satellite Imagery by Coupling
743 Variational Segmentation and Radiometric Analysis. *ISPRS International Journal of Geo-*
744 *Information*, 10, <https://doi.org/10.3390/ijgi10010017>.
- 745 Chen, B., Huang, B., Chen, L.F., & Xu, B. (2017). Spatially and Temporally Weighted Regression:
746 A Novel Method to Produce Continuous Cloud-Free Landsat Imagery. *IEEE Transactions*
747 *on Geoscience and Remote Sensing*, 55, 27-37.
- 748 Chen, B., Jin, Y.F., & Brown, P. (2019). An enhanced bloom index for quantifying floral
749 phenology using multi-scale remote sensing observations. *Isprs Journal of*
750 *Photogrammetry and Remote Sensing*, 156, 108-120.
- 751 Chen, J., Jonsson, P., Tamura, M., Gu, Z.H., Matsushita, B., & Eklundh, L. (2004). A simple
752 method for reconstructing a high-quality NDVI time-series data set based on the Savitzky-
753 Golay filter. *Remote Sensing of Environment*, 91, 332-344.
- 754 Chen, J., Zhu, X.L., Vogelmann, J.E., Gao, F., & Jin, S.M. (2011). A simple and effective method
755 for filling gaps in Landsat ETM plus SLC-off images. *Remote Sensing of Environment*,
756 115, 1053-1064.
- 757 Cheng, Y., Vrieling, A., Fava, F., Meroni, M., Marshall, M., & Gachoki, S. (2020). Phenology of
758 short vegetation cycles in a Kenyan rangeland from PlanetScope and Sentinel-2. *Remote*
759 *Sensing of Environment*, 248, <https://doi.org/10.1016/j.rse.2020.112004>.
- 760 Dechant, B., Ryu, Y., Badgley, G., Kohler, P., Rascher, U., Migliavacca, M., Zhang, Y.G.,
761 Tagliabue, G., Guan, K.Y., Rossini, M., Goulas, Y., Zeng, Y.L., Frankenberg, C., & Berry,
762 J.A. (2022). NIRvP: A robust structural proxy for sun-induced chlorophyll fluorescence
763 and photosynthesis across scales. *Remote Sensing of Environment*, 268.
- 764 Detto, M., Wright, S.J., Calderon, O., & Muller-Landau, H.C. (2018). Resource acquisition and
765 reproductive strategies of tropical forest in response to the El Nino-Southern Oscillation.
766 *Nature Communications*, 9, <https://doi.org/10.1038/s41467-41018-03306-41469>.
- 767 Feng, Y., Zeng, Z.Z., Searchinger, T.D., Ziegler, A.D., Wu, J., Wang, D.S., He, X.Y., Elsen, P.R.,
768 Ciais, P., Xu, R.R., Guo, Z.L., Peng, L.Q., Tao, Y.H., Spracklen, D.V., Holden, J., Liu,
769 X.P., Zheng, Y., Xu, P., Chen, J., Jiang, X., Song, X.P., Lakshmi, V., Wood, E.F., & Zheng,
770 C.M. (2022). Doubling of annual forest carbon loss over the tropics during the early
771 twenty-first century. *Nature Sustainability*, 1-8.
- 772 Forzieri, G., Girardello, M., Ceccherini, G., Spinoni, J., Feyen, L., Hartmann, H., Beck, P.S.A.,
773 Camps-Valls, G., Chirici, G., Mauri, A., & Cescatti, A. (2021). Emergent vulnerability to
774 climate-driven disturbances in European forests. *Nature Communications*, 12, 1-12.

775 Gaofen-1 (2020). GF-1 (Gaofen-1) High-resolution Imaging Satellite / CHEOS series of China.
776 In: <https://directory.eoportal.org/web/eoportal/satellite-missions/g/gaofen-1> (accessed 03
777 October 2020).

778 Gonzalez, R.C., & Woods, R.E. (2006). *Digital Image Processing (3rd Edition)*. Prentice-Hall,
779 Inc.

780 Griffiths, P., Nendel, C., & Hostert, P. (2019). Intra-annual reflectance composites from Sentinel-
781 2 and Landsat for national-scale crop and land cover mapping. *Remote Sensing of*
782 *Environment*, 220, 135-151.

783 Guan, X.D., Liu, G.H., Huang, C., Liu, Q.S., Wu, C.S., Jin, Y., & Li, Y.F. (2017). An Object-
784 Based Linear Weight Assignment Fusion Scheme to Improve Classification Accuracy
785 Using Landsat and MODIS Data at the Decision Level. *Ieee Transactions on Geoscience*
786 *and Remote Sensing*, 55, 6989-7002.

787 Halls, J.N., & Magolan, J.L. (2019). A Methodology to Assess Land Use Development, Flooding,
788 and Wetland Change as Indicators of Coastal Vulnerability. *Remote Sensing*, 11,
789 <https://doi.org/10.3390/rs11192260>.

790 He, K.M., Sun, J., & Tang, X.O. (2013). Guided Image Filtering. *IEEE Transactions on Pattern*
791 *Analysis and Machine Intelligence*, 35, 1397-1409.

792 Heinrich, V.H.A., Dalagnol, R., Cassol, H.L.G., Rosan, T.M., de Almeida, C.T., Silva, C.H.L.,
793 Campanharo, W.A., House, J.I., Sitch, S., Hales, T.C., Adami, M., Anderson, L.O., &
794 Aragao, L.E.O.C. (2021). Large carbon sink potential of secondary forests in the Brazilian
795 Amazon to mitigate climate change. *Nature Communications*, 12, 1-11.

796 Hossain, M.D., & Chen, D. (2019). Segmentation for Object-Based Image Analysis (OBIA): A
797 review of algorithms and challenges from remote sensing perspective. *ISPRS Journal of*
798 *Photogrammetry and Remote Sensing*, 150, 115-134.

799 Huang, B., Li, Y., Han, X.Y., Cui, Y.Z., Li, W.B., & Li, R.R. (2015). Cloud Removal From Optical
800 Satellite Imagery With SAR Imagery Using Sparse Representation. *IEEE Geoscience and*
801 *Remote Sensing Letters*, 12, 1046-1050.

802 Huang, B., & Zhang, H.K. (2014). Spatio-temporal reflectance fusion via unmixing: accounting
803 for both phenological and land-cover changes. *International Journal of Remote Sensing*,
804 35, 6213-6233.

805 Ju, J.C., & Roy, D.P. (2008). The availability of cloud-free Landsat ETM plus data over the
806 conterminous United States and globally. *Remote Sensing of Environment*, 112, 1196-
807 1211.

808 Kimm, H., Guan, K.Y., Jiang, C.Y., Peng, B., Gentry, L.F., Wilkin, S.C., Wang, S.B., Cai, Y.P.,
809 Bernacchi, C.J., Peng, J., & Luo, Y.N. (2020). Deriving high-spatiotemporal-resolution
810 leaf area index for agroecosystems in the US Corn Belt using Planet Labs CubeSat and
811 STAIR fusion data. *Remote Sensing of Environment*, 239,
812 <https://doi.org/10.1016/j.rse.2019.111615>.

813 Kong, J., Ryu, Y., Liu, J., Dechant, B., Rey-Sanchez, C., Shortt, R., Szutu, D., Verfaillie, J.,
814 Houborg, R., & Baldocchi, D.D. (2022). Matching high resolution satellite data and flux
815 tower footprints improves their agreement in photosynthesis estimates. *Agricultural and*
816 *Forest Meteorology*, 316, 108878.

817 Leigh, E.G. (1999). *Tropical forest ecology: a view from Barro Colorado Island*. New York, NY:
818 Oxford University Press.

- 819 Li, D.R., Zhang, G.F., Wu, Z.C., & Yi, L.N. (2010). An Edge Embedded Marker-Based Watershed
820 Algorithm for High Spatial Resolution Remote Sensing Image Segmentation. *IEEE*
821 *Transactions on Image Processing*, 19, 2781-2787.
- 822 Li, L.F., Franklin, M., Girguis, M., Lurmann, F., Wu, J., Pavlovic, N., Breton, C., Gilliland, F., &
823 Habre, R. (2020a). Spatiotemporal imputation of MAIAC AOD using deep learning with
824 downscaling. *Remote Sensing of Environment*, 237,
825 <https://doi.org/10.1016/j.rse.2019.111584>.
- 826 Li, W.B., Li, Y., & Chan, J.C.W. (2020b). Thick Cloud Removal With Optical and SAR Imagery
827 via Convolutional-Mapping-Deconvolutional Network. *IEEE Transactions on Geoscience*
828 *and Remote Sensing*, 58, 2865-2879.
- 829 Li, W.B., Li, Y., Chen, D., & Chan, J.C.W. (2019). Thin cloud removal with residual symmetrical
830 concatenation network. *ISPRS Journal of Photogrammetry and Remote Sensing*, 153, 137-
831 150.
- 832 Li, Z.W., Shen, H.F., Li, H.F., Xia, G.S., Gamba, P., & Zhang, L.P. (2017). Multi-feature
833 combined cloud and cloud shadow detection in GaoFen-1 wide field of view imagery.
834 *Remote Sensing of Environment*, 191, 342-358.
- 835 Lin, C.H., Lai, K.H., Chen, Z.B., & Chen, J.Y. (2014). Patch-Based Information Reconstruction
836 of Cloud-Contaminated Multitemporal Images. *IEEE Transactions on Geoscience and*
837 *Remote Sensing*, 52, 163-174.
- 838 Lin, C.H., Tsai, P.H., Lai, K.H., & Chen, J.Y. (2013). Cloud Removal From Multitemporal
839 Satellite Images Using Information Cloning. *IEEE Transactions on Geoscience and*
840 *Remote Sensing*, 51, 232-241.
- 841 Liu, J., Zhenyao, S.Y., Yan, T.Z., & Yang, Y.C. (2018). Source identification and impact of
842 landscape pattern on riverine nitrogen pollution in a typical urbanized watershed, Beijing,
843 China. *Science of the Total Environment*, 628-629, 1296-1307.
- 844 Liu, R.G., Shang, R., Liu, Y., & Lu, X.L. (2017). Global evaluation of gap-filling approaches for
845 seasonal NDVI with considering vegetation growth trajectory, protection of key point,
846 noise resistance and curve stability. *Remote Sensing of Environment*, 189, 164-179.
- 847 Lloyd, S.P. (1982). Least-Squares Quantization in Pcm. *IEEE Transactions on Information*
848 *Theory*, 28, 129-137.
- 849 Lopez-Poma, R., Pivello, V.R., de Brito, G.S., & Bautista, S. (2020). Impact of the conversion of
850 Brazilian woodland savanna (cerradao) to pasture and Eucalyptus plantations on soil
851 nitrogen mineralization. *Science of the Total Environment*, 704,
852 <https://doi.org/10.1016/j.scitotenv.2019.135397>.
- 853 Lorenzi, L., Melgani, F., & Mercier, G. (2011). Inpainting Strategies for Reconstruction of Missing
854 Data in VHR Images. *IEEE Geoscience and Remote Sensing Letters*, 8, 914-918.
- 855 Luo, Y.N., Guanb, K.Y., & Pen, J. (2018). STAIR: A generic and fully-automated method to fuse
856 multiple sources of optical satellite data to generate a high-resolution, daily and cloud-/gap-
857 free surface reflectance product. *Remote Sensing of Environment*, 214, 87-99.
- 858 Maalouf, A., Carre, P., Augereau, B., & Fernandez-Maloigne, C. (2009). A Bandelet-Based
859 Inpainting Technique for Clouds Removal From Remotely Sensed Images. *IEEE*
860 *Transactions on Geoscience and Remote Sensing*, 47, 2363-2371.
- 861 Maxwell, S.K., Schmidt, G.L., & Storey, J.C. (2007). A multi-scale segmentation approach to
862 filling gaps in Landsat ETM+ SLC-off images. *International Journal of Remote Sensing*,
863 28, 5339-5356.

864 McCabe, M.F., Aragon, B., Houborg, R., & Mascaro, J. (2017). CubeSats in Hydrology:
865 Ultrahigh-Resolution Insights Into Vegetation Dynamics and Terrestrial Evaporation.
866 *Water Resources Research*, 53, 10017-10024.

867 Meraner, A., Ebel, P., Zhu, X.X., & Schmitt, M. (2020). Cloud removal in Sentinel-2 imagery
868 using a deep residual neural network and SAR-optical data fusion. *ISPRS Journal of*
869 *Photogrammetry and Remote Sensing*, 166, 333-346.

870 Myint, S.W., Gober, P., Brazel, A., Grossman-Clarke, S., & Weng, Q.H. (2011). Per-pixel vs.
871 object-based classification of urban land cover extraction using high spatial resolution
872 imagery. *Remote Sensing of Environment*, 115, 1145-1161.

873 Norris, J.R., Allen, R.J., Evan, A.T., Zelinka, M.D., O'Dell, C.W., & Klein, S.A. (2016). Evidence
874 for climate change in the satellite cloud record. *Nature*, 536, 72–75.

875 Park, J.Y., Muller-Landau, H.C., Lichstein, J.W., Rifai, S.W., Dandois, J.P., & Bohlman, S.A.
876 (2019). Quantifying Leaf Phenology of Individual Trees and Species in a Tropical Forest
877 Using Unmanned Aerial Vehicle (UAV) Images. *Remote Sensing*, 11,
878 <https://doi.org/10.3390/rs11131534>.

879 Pipia, L., Munoz-Mari, J., Amin, E., Belda, S., Camps-Valls, G., & Verrelst, J. (2019). Fusing
880 optical and SAR time series for LAI gap filling with multioutput Gaussian processes.
881 *Remote Sensing of Environment*, 235, <https://doi.org/10.1016/j.rse.2019.111452>.

882 Planet (2021). PLANET SURFACE REFLECTANCE VERSION 2.0 [WWW Document]. In.
883 [https://assets.planet.com/marketing/PDF/Planet_Surface_Reflectance_Technical_White_](https://assets.planet.com/marketing/PDF/Planet_Surface_Reflectance_Technical_White_Paper.pdf)
884 [Paper.pdf](https://assets.planet.com/marketing/PDF/Planet_Surface_Reflectance_Technical_White_Paper.pdf) (accessed 14 April 2021).

885 Pringle, M.J., Schmidt, M., & Muir, J.S. (2009). Geostatistical interpolation of SLC-off Landsat
886 ETM plus images. *ISPRS Journal of Photogrammetry and Remote Sensing*, 64, 654-664.

887 Qi, Z.M., Helmers, M.J., & Kaleita, A.L. (2011). Soil water dynamics under various agricultural
888 land covers on a subsurface drained field in north-central Iowa, USA. *Agricultural Water*
889 *Management*, 98, 665-674.

890 Qin, Y.W., Xiao, X.M., Dong, J.W., Zhang, Y., Wu, X.C., Shimabukuro, Y., Arai, E., Biradar, C.,
891 Wang, J., Zou, Z.H., Liu, F., Shi, Z., Doughty, R., & Moore, B. (2019). Improved estimates
892 of forest cover and loss in the Brazilian Amazon in 2000-2017. *Nature Sustainability*, 2,
893 764-772.

894 Qiu, S., Zhu, Z., & He, B.B. (2019). Fmask 4.0: Improved cloud and cloud shadow detection in
895 Landsats 4-8 and Sentinel-2 imagery. *Remote Sensing of Environment*, 231,
896 <https://doi.org/10.1016/j.rse.2019.1005.1024>.

897 Rao, P., Zhou, W.Q., Bhattarai, N., Srivastava, A.K., Singh, B., Poonia, S., Lobell, D.B., & Jain,
898 M. (2021). Using Sentinel-1, Sentinel-2, and Planet Imagery to Map Crop Type of
899 Smallholder Farms. *Remote Sensing*, 13, <https://doi.org/10.3390/rs13101870>.

900 Rasanen, A., & Virtanen, T. (2019). Data and resolution requirements in mapping vegetation in
901 spatially heterogeneous landscapes. *Remote Sensing of Environment*, 230,
902 <https://doi.org/10.1016/j.rse.2019.1005.1026>.

903 Roy, D.P., Huang, H.Y., Houborg, R., & Martins, V.S. (2021). A global analysis of the temporal
904 availability of PlanetScope high spatial resolution multi-spectral imagery. *Remote Sensing*
905 *of Environment*, 264, <https://doi.org/10.1016/j.rse.2021.112586>.

906 Roy, D.P., Ju, J., Lewis, P., Schaaf, C., Gao, F., Hansen, M., & Lindquist, E. (2008). Multi-
907 temporal MODIS-Landsat data fusion for relative radiometric normalization, gap filling,
908 and prediction of Landsat data. *Remote Sensing of Environment*, 112, 3112-3130.

- 909 Sadeh, Y., Zhu, X., Dunkerley, D., Walker, J.P., Zhang, Y.X., Rozenstein, O., Manivasagam, V.S.,
910 & Chenu, K. (2021). Fusion of Sentinel-2 and PlanetScope time-series data into daily 3 m
911 surface reflectance and wheat LAI monitoring. *International Journal of Applied Earth*
912 *Observation and Geoinformation*, 96, 102260.
- 913 Salman, N. (2006). Image segmentation based on watershed and edge detection techniques. *The*
914 *International Arab Journal of Information Technology*, 3, 104-110.
- 915 Shen, H.F., Li, X.H., Cheng, Q., Zeng, C., Yang, G., Li, H.F., & Zhang, L.P. (2015). Missing
916 Information Reconstruction of Remote Sensing Data: A Technical Review. *IEEE*
917 *Geoscience and Remote Sensing Magazine*, 3, 61-85.
- 918 SkySat (2020). SkySat constellation of Terra Bella - formerly SkySat Imaging Program of Skybox
919 Imaging. In. <https://directory.eoportal.org/web/eoportal/satellite-missions/s/skysat>
920 (accessed 03 October 2020).
- 921 Soille, P. (1999). *Morphological image analysis: Principles and applications* Springer-Verlag
922 Berlin Heidelberg.
- 923 Soille, P., & Pesaresi, M. (2002). Advances in mathematical morphology applied to geoscience
924 and remote sensing. *IEEE Transactions on Geoscience and Remote Sensing*, 40, 2042-
925 2055.
- 926 Solomon, C., & Breckon, T. (2010). *Fundamentals of Digital Image Processing, A Practical*
927 *Approach with example in Matlab*. Wiley-Blackwell, Hoboken.
- 928 SPOT-7 (2020). SPOT-6 and SPOT-7 Commercial Imaging Constellation. In.
929 <https://directory.eoportal.org/web/eoportal/satellite-missions/s/spot-6-7> (accessed 03
930 October 2019).
- 931 Taubert, F., Fischer, R., Groeneveld, J., Lehmann, S., Muller, M.S., Rodig, E., Wiegand, T., &
932 Huth, A. (2018). Global patterns of tropical forest fragmentation. *Nature*, 554, 519-522.
- 933 Wang, F.R., Fan, X.M., Yunus, A.P., Subramanian, S.S., Alonso-Rodriguez, A., Dai, L.X., Xu,
934 Q., & Huang, R.Q. (2019). Coseismic landslides triggered by the 2018 Hokkaido, Japan
935 (M-w 6.6), earthquake: spatial distribution, controlling factors, and possible failure
936 mechanism. *Landslides*, 16, 1551-1566.
- 937 Wang, J., Yang, D.D., Chen, S.L., Zhu, X.L., Wu, S.B., Bogonovich, M., Guo, Z.F., Zhu, Z., &
938 Wu, J. (2021). Automatic cloud and cloud shadow detection in tropical areas for
939 PlanetScope satellite images. *Remote Sensing of Environment*, 264,
940 <https://doi.org/10.1016/j.rse.2021.112604>.
- 941 Wang, J., Yang, D.D., Detto, M., Nelson, B.W., Chen, M., Guan, K.Y., Wu, S.B., Yan, Z.B., &
942 Wu, J. (2020). Multi-scale integration of satellite remote sensing improves characterization
943 of dry-season green-up in an Amazon tropical evergreen forest. *Remote Sensing of*
944 *Environment*, 246, <https://doi.org/10.1016/j.rse.2020.111865>.
- 945 Wang, Z., Bovik, A.C., Sheikh, H.R., & Simoncelli, E.P. (2004). Image quality assessment: From
946 error visibility to structural similarity. *IEEE Transactions on Image Processing*, 13, 600-
947 612.
- 948 Watkins, B., & Van Niekerk, A. (2019a). Automating field boundary delineation with multi-
949 temporal Sentinel-2 imagery. *Computers and Electronics in Agriculture*, 167,
950 <https://doi.org/10.1016/j.compag.2019.105078>.
- 951 Watkins, B., & Van Niekerk, A. (2019b). A comparison of object-based image analysis approaches
952 for field boundary delineation using multi-temporal Sentinel-2 imagery. *Computers and*
953 *Electronics in Agriculture*, 158, 294-302.

- 954 Wu, S.B., Wang, J., Yan, Z.B., Song, G.Q., Chen, Y., Ma, Q., Deng, M.F., Wu, Y.T., Zhao, Y.Y.,
 955 Guo, Z.F., Xu, X.T., Yang, X., Su, Y.J., Liu, L.L., & Wu, J. (2021). Monitoring tree-crown
 956 scale autumn leaf phenology in a temperate forest with an integration of PlanetScope and
 957 drone remote sensing observations. *ISPRS Journal of Photogrammetry and Remote*
 958 *Sensing*, 171, 36-48.
- 959 Wu, W., Ge, L.Q., Luo, J.C., Huan, R.H., & Yang, Y.P. (2018). A Spectral-Temporal Patch-Based
 960 Missing Area Reconstruction for Time-Series Images. *Remote Sensing*, 10,
 961 <https://doi.org/10.3390/rs10101560>.
- 962 Xi, W.Q., Du, S.H., Wang, Y.C., & Zhang, X.Y. (2019). A spatiotemporal cube model for
 963 analyzing satellite image time series: Application to land-cover mapping and change
 964 detection. *Remote Sensing of Environment*, 231.
- 965 Yan, L., & Roy, D.P. (2018). Large-Area Gap Filling of Landsat Reflectance Time Series by
 966 Spectral-Angle-Mapper Based Spatio-Temporal Similarity (SAMSTS). *Remote Sensing*,
 967 10, <https://doi.org/10.3390/rs10040609>.
- 968 Yan, L., & Roy, D.P. (2020). Spatially and temporally complete Landsat reflectance time series
 969 modelling: The fill-and-fit approach. *Remote Sensing of Environment*, 241,
 970 <https://doi.org/10.1016/j.rse.2020.111718>.
- 971 Zeng, C., Shen, H.F., & Zhang, L.P. (2013). Recovering missing pixels for Landsat ETM plus
 972 SLC-off imagery using multi-temporal regression analysis and a regularization method.
 973 *Remote Sensing of Environment*, 131, 182-194.
- 974 Zeng, Z.Z., Estes, L., Ziegler, A.D., Chen, A.P., Searchinger, T., Hua, F.Y., Guan, K.Y., Jintrawet,
 975 A., & Wood, E.F. (2018). Highland cropland expansion and forest loss in Southeast Asia
 976 in the twenty-first century. *Nature Geoscience*, 11, 556-562.
- 977 Zeng, Z.Z., Wang, D.S., Yang, L., Wu, J., Ziegler, A.D., Liu, M.F., Ciais, P., Searchinger, T.D.,
 978 Yang, Z.L., Chen, D.L., Chen, A.P., Li, L.Z.X., Piao, S.L., Taylor, D., Cai, X.T., Pan, M.,
 979 Peng, L.Q., Lin, P.R., Gower, D., Feng, Y., Zheng, C.M., Guan, K.Y., Lian, X., Wang, T.,
 980 Wang, L., Jeong, S.J., Wei, Z.W., Sheffield, J., Caylor, K., & Wood, E.F. (2021).
 981 Deforestation-induced warming over tropical mountain regions regulated by elevation.
 982 *Nature Geoscience*, 14, 23-29.
- 983 Zhang, C., Li, W., & Travis, D. (2007). Gaps-fill of SLC-off Landsat ETM plus satellite image
 984 using a geostatistical approach. *International Journal of Remote Sensing*, 28, 5103-5122.
- 985 Zhang, Q., Yuan, Q.Q., Li, J., Li, Z.W., Shen, H.F., & Zhang, L.P. (2020). Thick cloud and cloud
 986 shadow removal in multitemporal imagery using progressively spatio-temporal patch
 987 group deep learning. *ISPRS Journal of Photogrammetry and Remote Sensing*, 162, 148-
 988 160.
- 989 Zhang, Q., Yuan, Q.Q., Zeng, C., Li, X.H., & Wei, Y.C. (2018). Missing Data Reconstruction in
 990 Remote Sensing Image With a Unified Spatial-Temporal-Spectral Deep Convolutional
 991 Neural Network. *IEEE Transactions on Geoscience and Remote Sensing*, 56, 4274-4288.
- 992 Zheng, X., Gong, P., Strome, M. (1995). Characterizing spatial structure of tree canopy using
 993 colour photographs and mathematical morphology. *Canadian Journal of Remote Sensing*,
 994 21, 421-429.
- 995 Zhu, X.L., Gao, F., Liu, D.S., & Chen, J. (2012a). A Modified Neighborhood Similar Pixel
 996 Interpolator Approach for Removing Thick Clouds in Landsat Images. *IEEE Geoscience*
 997 *and Remote Sensing Letters*, 9, 521-525.

998 Zhu, X.L., & Helmer, E.H. (2018). An automatic method for screening clouds and cloud shadows
999 in optical satellite image time series in cloudy regions. *Remote Sensing of Environment*,
1000 214, 135-153.

1001 Zhu, X.L., Helmer, E.H., Chen, J., & Liu, D.S. (2018). *An Automatic System for Reconstructing*
1002 *High-Quality Seasonal Landsat Time Series. Chapter 2 (Pages 25-42), In: Qihao Weng,*
1003 *Ed. Remote Sensing: Time Series Image Processing.* Boca Raton: CRC Press.

1004 Zhu, X.L., Liu, D.S., & Chen, J. (2012b). A new geostatistical approach for filling gaps in Landsat
1005 ETM plus SLC-off images. *Remote Sensing of Environment*, 124, 49-60.

1006 Zhu, Z., Wang, S.X., & Woodcock, C.E. (2015). Improvement and expansion of the Fmask
1007 algorithm: cloud, cloud shadow, and snow detection for Landsats 4-7, 8, and Sentinel 2
1008 images. *Remote Sensing of Environment*, 159, 269-277.

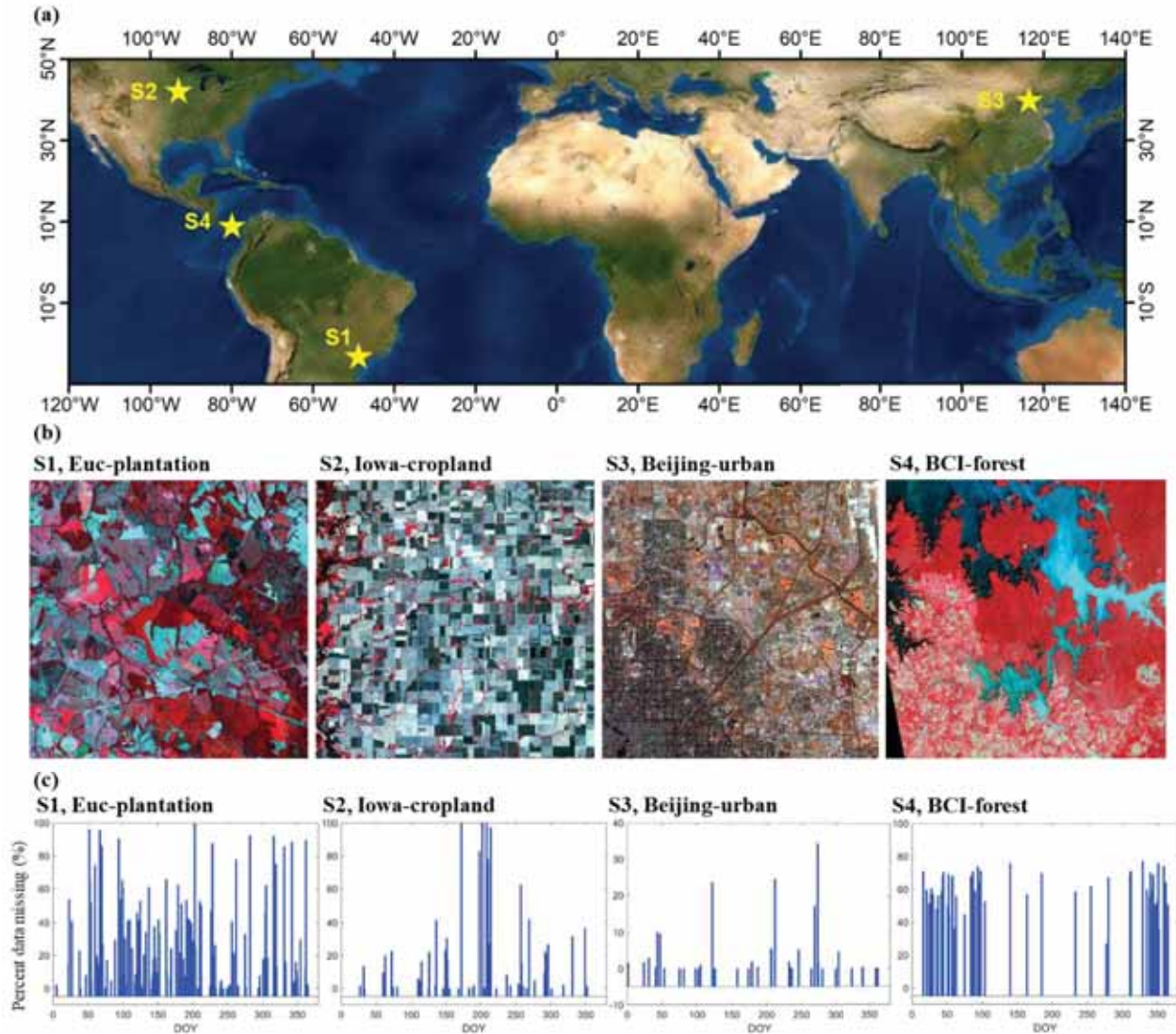
1009 Zhu, Z., & Woodcock, C.E. (2012). Object-based cloud and cloud shadow detection in Landsat
1010 imagery. *Remote Sensing of Environment*, 118, 83-94.

1011 Zi, Y., Xie, F.Y., Zhang, N., Jiang, Z.G., Zhu, W.T., & Zhang, H.P. (2021). Thin Cloud Removal
1012 for Multispectral Remote Sensing Images Using Convolutional Neural Networks
1013 Combined With an Imaging Model. *IEEE Journal of Selected Topics in Applied Earth*
1014 *Observations and Remote Sensing*, 14, 3811-3823.

1015

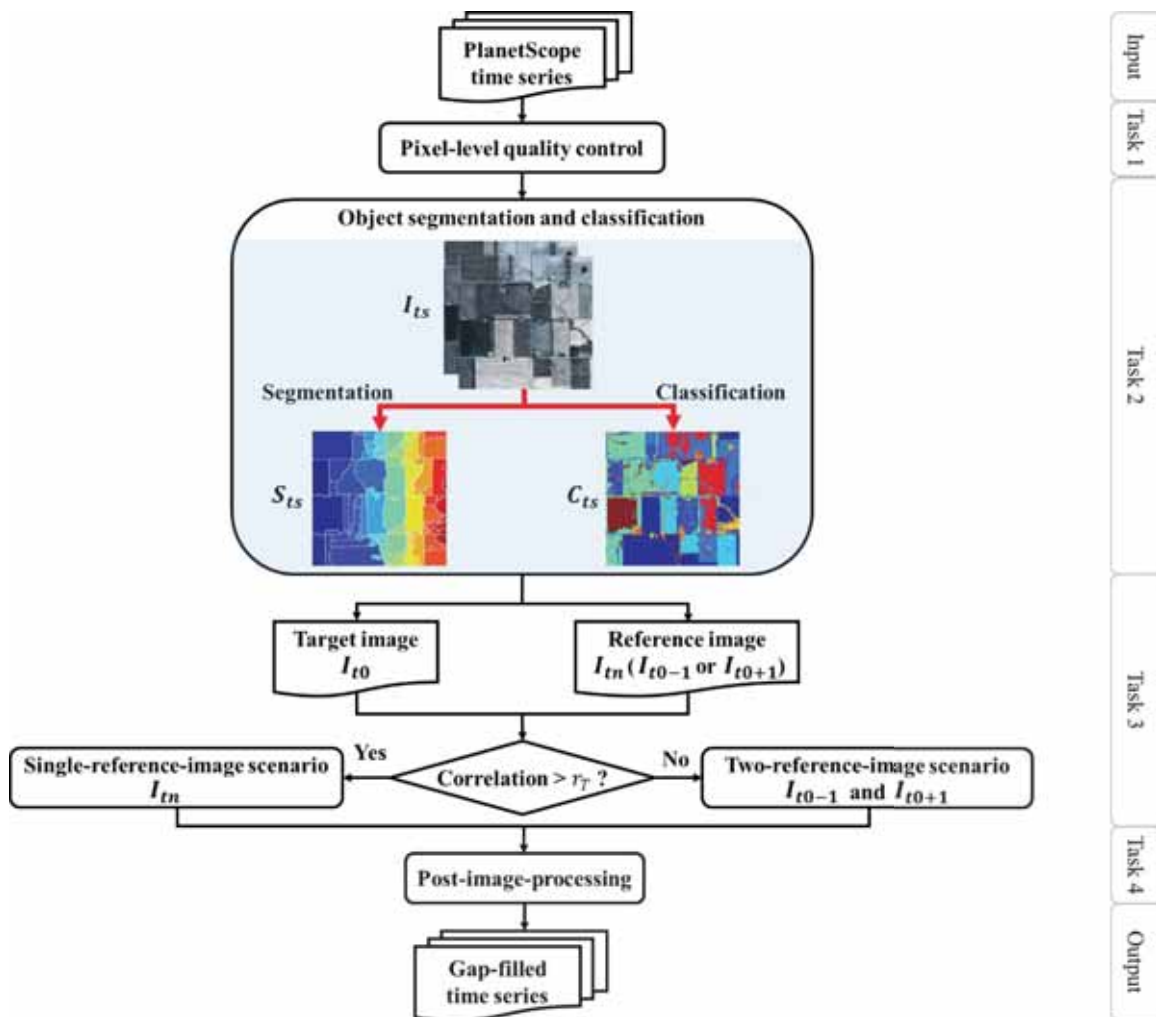
1016

1 **Figure 1.** Study sites. (a) locations (the yellow stars), (b) false color composites (RGB=NIR-Red-Green)
 2 of PlanetScope images, and (c) percent data missing of PlanetScope time-series images of the four testing
 3 sites (i.e. S1-S4; spatial extent: 20km×20km, temporal coverage: January-December, 2018), including
 4 Euc-plantation, Iowa-cropland, Beijing-urban, and Barro Colorado Island (BCI)-forest. The map in panel
 5 (a) is adapted from National Geographic, ESRI. The percent data missing in panel (c) is derived from
 6 STI-ACSS method (Wang et al., 2021).



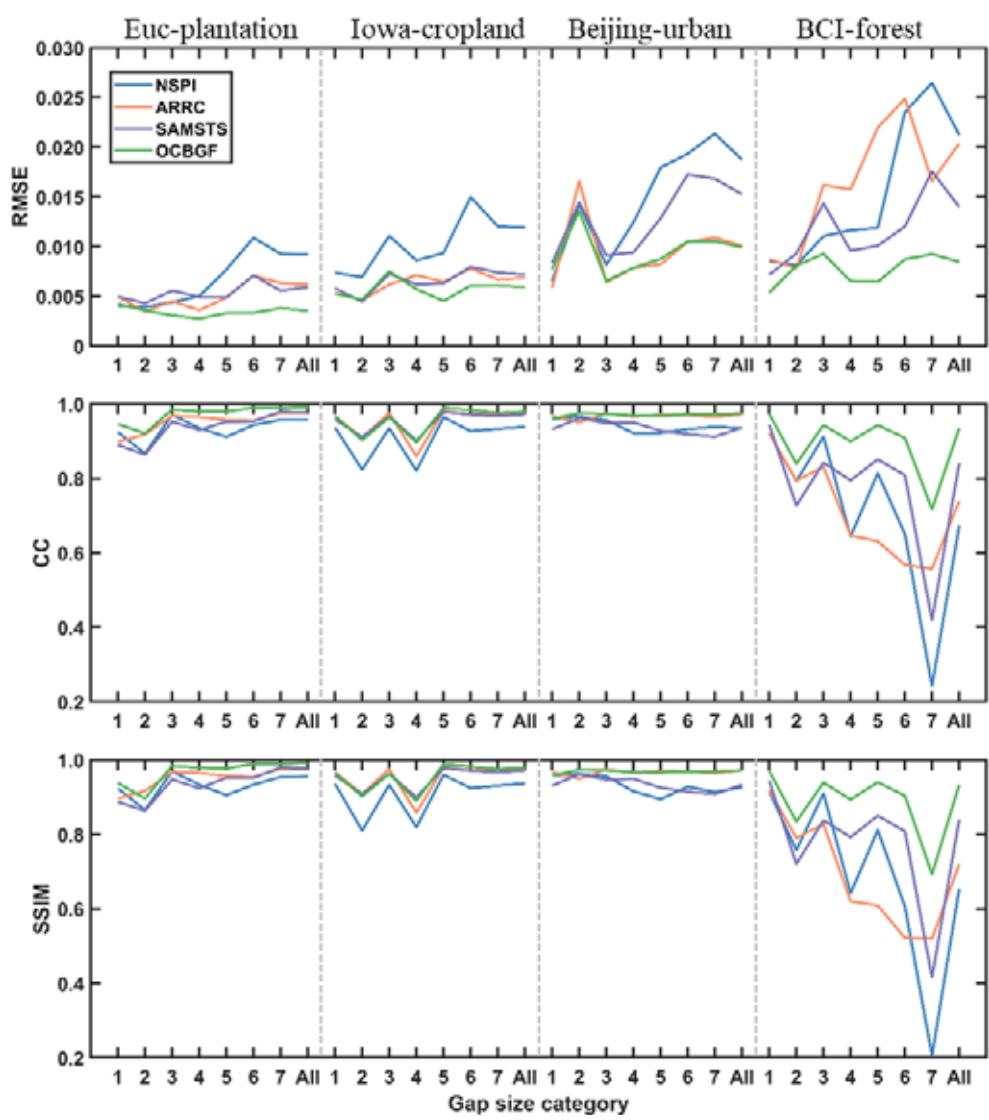
7

8 **Figure 2.** Flowchart of the four key tasks of the object and class based gap-filling (OCBGF) method. Task
 9 1: pixel-level quality control; Task 2: object segmentation and classification; Task 3: gap-filling with two
 10 scenarios: single-reference-image and two-reference-images scenarios; Task 4: post-image-processing.



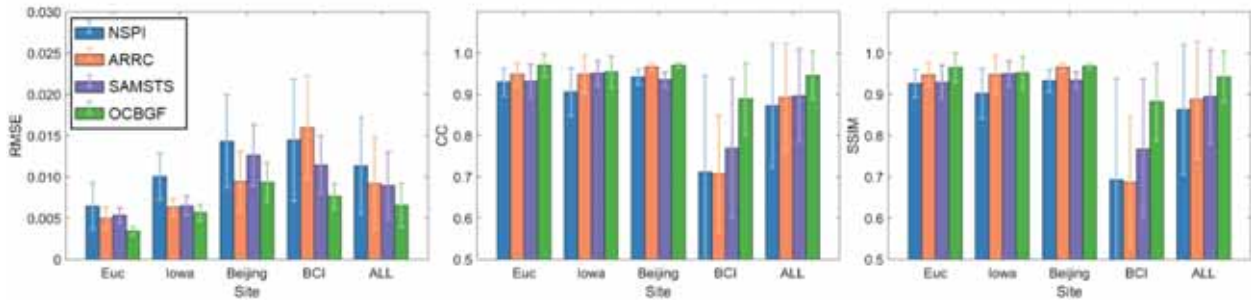
11
 12
 13

14 **Figure 3.** Cross-method comparison of gap-filled results across seven gap size categories (Table 2) at four
 15 sites using original images as benchmarks. The results of cross-method comparison include four-band
 16 average values of three indices: the root mean square error (RMSE), the correlation coefficient (CC), and
 17 the structure similarity index (SSIM) across all four sites. The four methods examined here include
 18 Neighborhood Similar Pixel Interpolator (NSPI), AutoRegression to Remove Clouds (ARRC), Spectral-
 19 Angle-Mapper Based Spatio-Temporal Similarity (SAMSTS), and OCBGF.



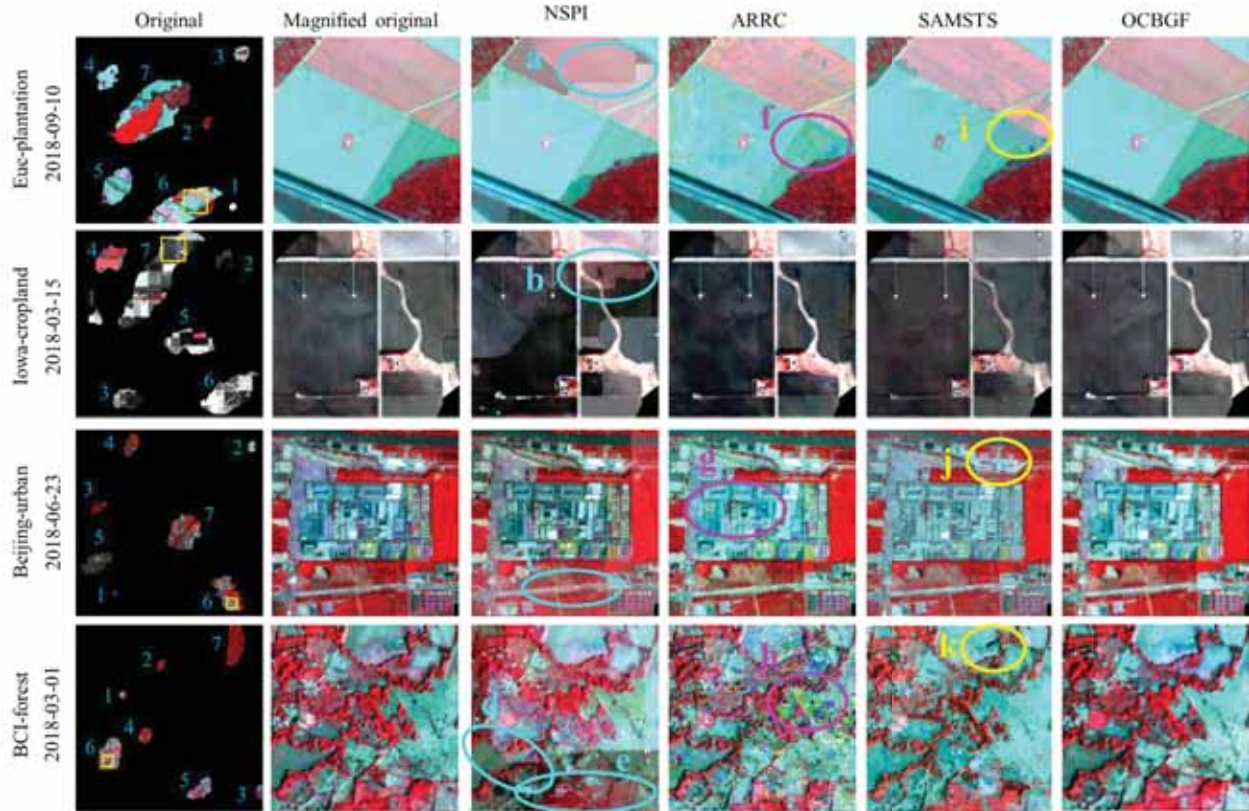
20

21 **Figure 4.** Cross-method comparison of all seven gap size categories across four sites using original images
 22 as benchmarks. Three metrics were used for the cross-method comparison, including RMSE, CC, and SSIM.
 23 The four methods examined here include NSPI, ARRC, SAMSTS, and OCBGF.



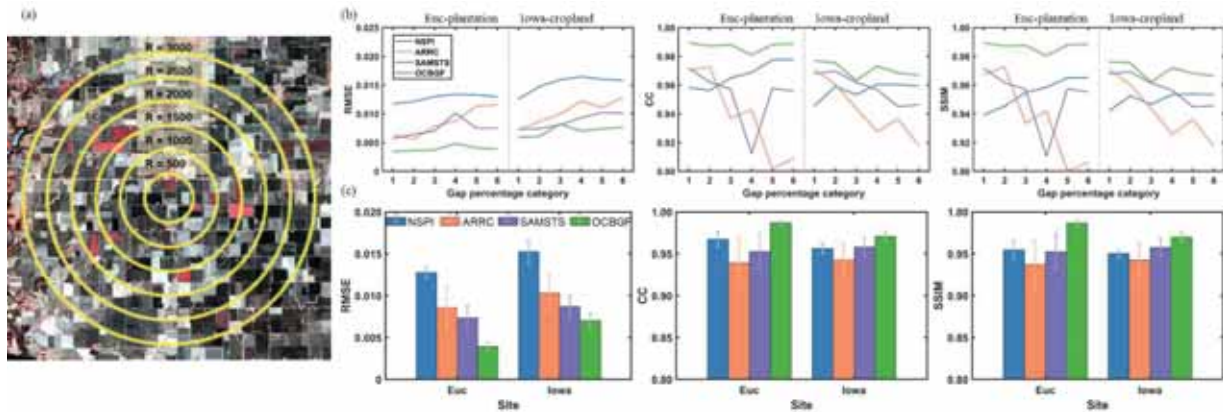
24

25 **Figure 5.** Assessing the effect of different gap sizes (Table 2) on the gap-filled results derived from NSPI,
 26 ARRC, SAMSTS, and OCBGF methods using original images as benchmarks. False color composites
 27 (RGB=NIR-Red-Green) of selected PlanetScope images and corresponding magnified areas (yellow
 28 squares in the first column) of gap-filling results using NSPI, ARRC, SAMSTS, and OCBGF are shown
 29 below. Blue circles highlight blocking (a, b, and d) and strip artifacts (c and e) in gap-filled results of NSPI,
 30 magenta circles highlight salt-and-pepper noises in gap-filled results of ARRC, and yellow circles highlight
 31 blocking artifacts (i) and errors (j and k) in gap-filled results of SAMSTS.



32

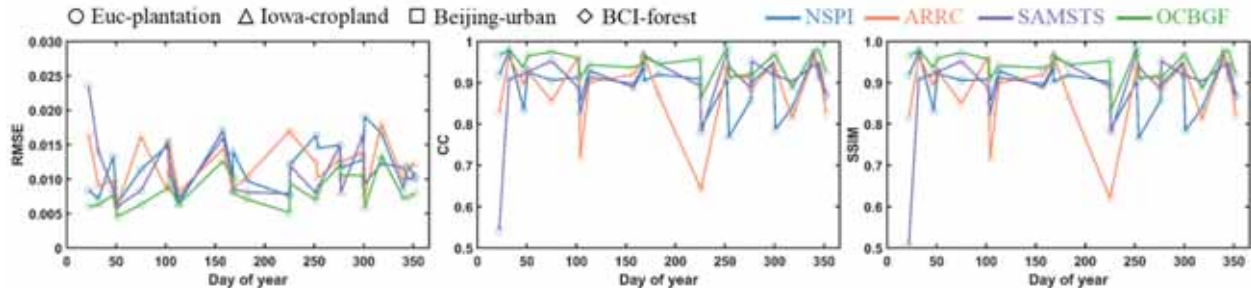
33 **Figure 6.** Cross-method comparison of gap-filled results across six gap percentage categories (a) at the
 34 Euc-plantation and Iowa-cropland sites using original images (the same as Fig. 5) as benchmarks. These
 35 categories have a circle shape with a radius (R) ranging from 500 to 3000 (pixels) with an interval of 500
 36 (pixels), resulting in a percent data missing of 2%, 7%, 16%, 28%, 44%, and 64%, respectively. The results
 37 of cross-method comparison include category-specific accuracy (b) and average accuracy of all six
 38 categories (c) across the two sites. Three metrics were used for the cross-method comparison, including
 39 RMSE, CC, and SSIM. The four methods examined here include NSPI, ARRC, SAMSTS, and OCBGF
 40 methods.



41

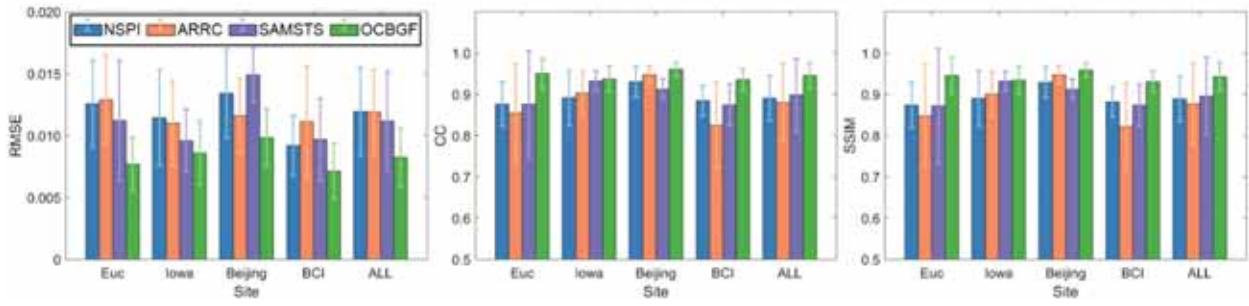
42

43 **Figure 7.** Cross-method comparison of gap-filled results across areas with land cover changes in different
 44 seasons at four sites using original images as benchmarks. The results of cross-method comparison include
 45 four-band average values of three indices: RMSE, CC, and SSIM across all four sites. The four methods
 46 examined here include NSPI, ARRC, SAMSTS, and OCBGF.



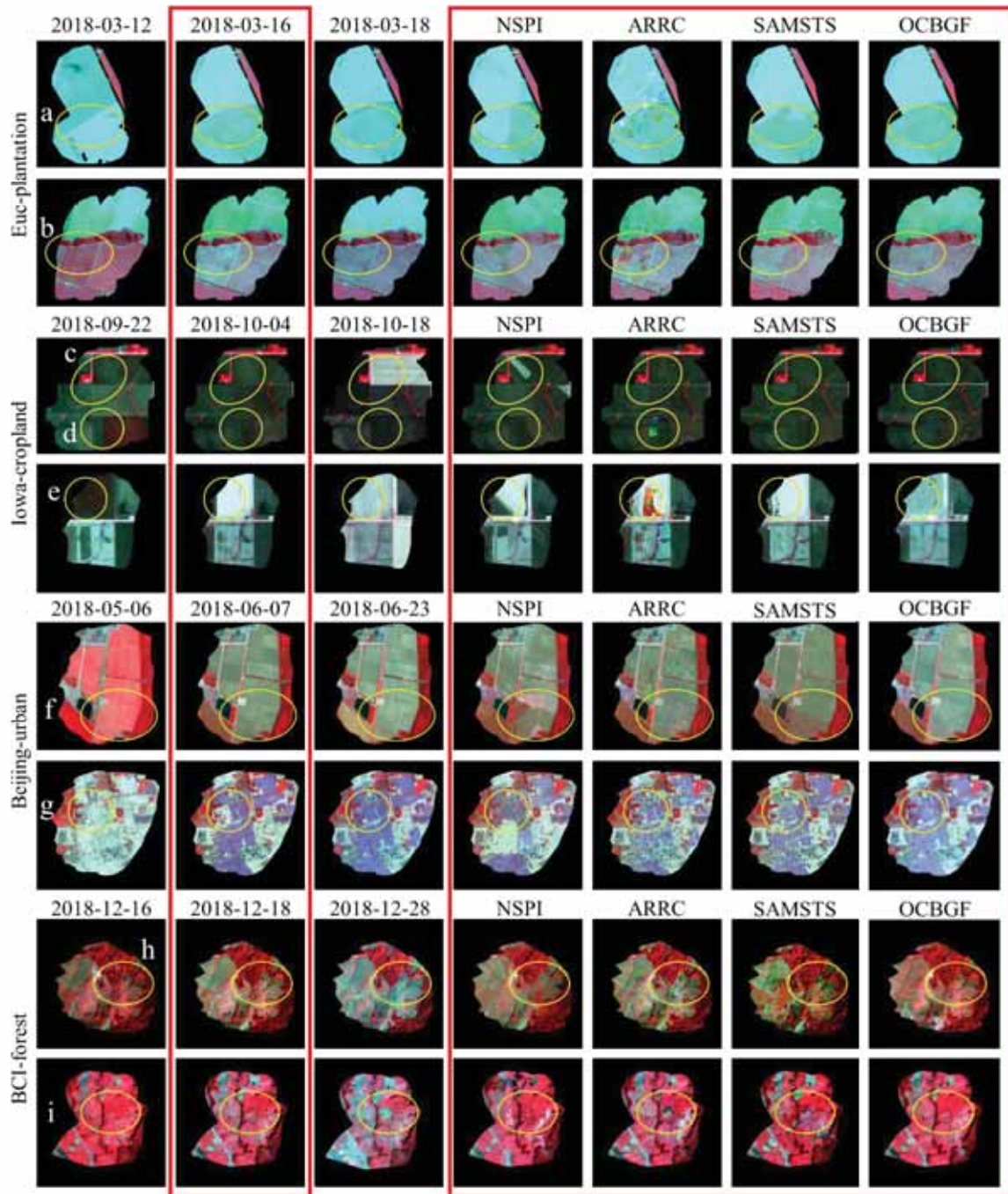
47

48 **Figure 8.** Cross-method comparison of all areas with land cover changes across four sites using original
 49 images as benchmarks. Three metrics were used for the cross-method comparison, including RMSE, CC,
 50 and SSIM. The four methods examined here include NSPI, ARRC, SAMSTS, and OCBGF.



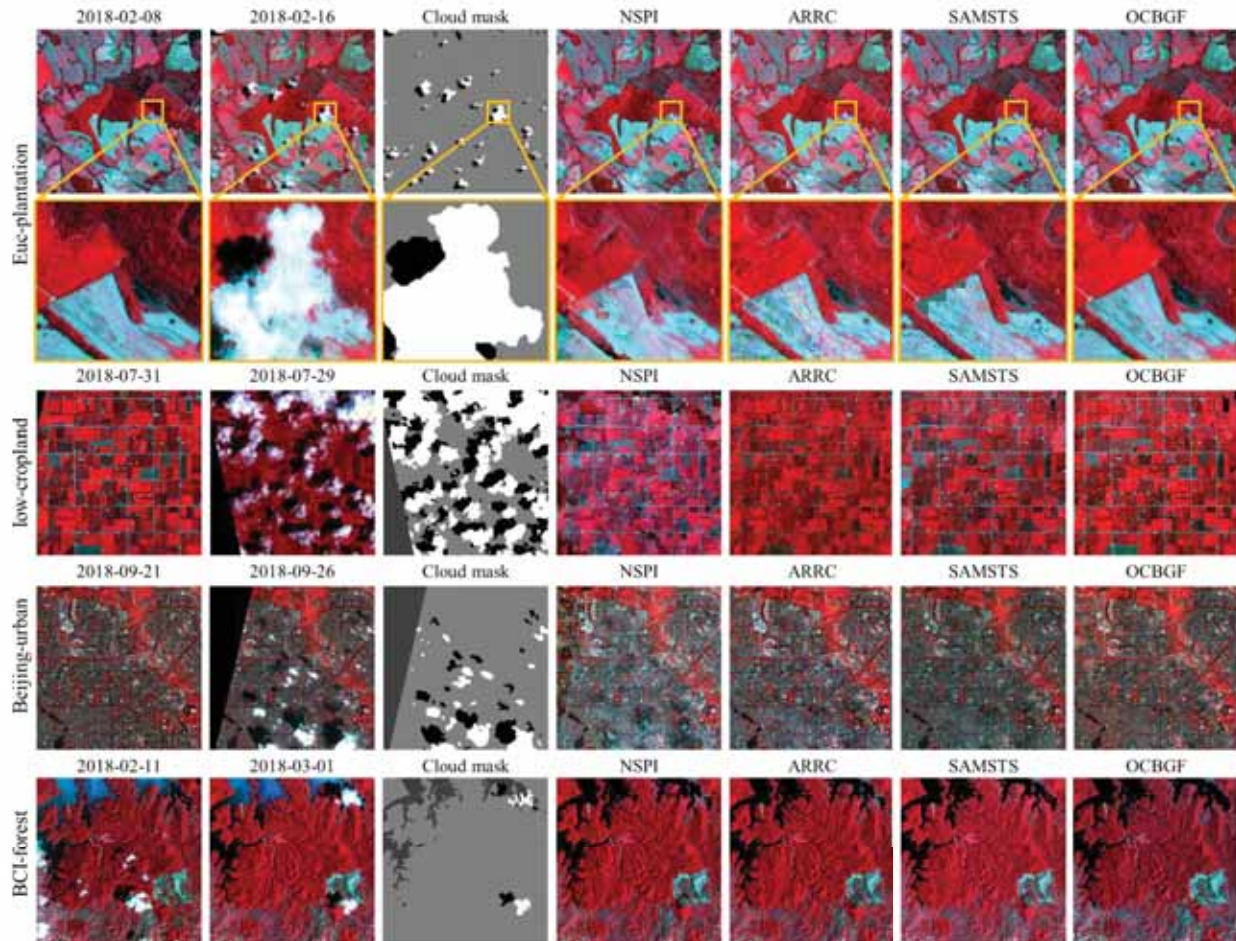
51

52 **Figure 9.** Assessing the effect of various land cover changes on the gap-filled results derived from NSPI,
 53 ARRC, SAMSTS, and OCBGF methods using original images as benchmarks. Eight representative land
 54 cover types from selected images of the four sites were used. False color composites (RGB=NIR-Red-
 55 Green) of selected PlanetScope images (including the date before the target date, the target date (the red
 56 frame), and the date after the target date in the first, second, and third column, respectively) and
 57 corresponding gap-filling results on the target date (red frame) using NSPI, ARRC, SAMSTS, and OCBGF
 58 are shown. Yellow circles highlight the areas experiencing various land cover change scenarios, with (i.e.
 59 b, h, and i) and without rapid land cover change (i.e. a, c, d, e, f, and g) over time.



60

61 **Figure 10.** Assessing the performance of gap-filling methods on four real cloud-contaminated images
 62 (target image, the second column) at all four sites. Four methods, NSPI, ARRC, SAMSTS, and OCBGF,
 63 are examined here. A (nearly) cloud-free image (the first column) close in time to the target image is used
 64 as the reference. The cloud masks of target images (the third column) are derived from STI-ACSS method
 65 (Wang et al., 2021).



66
 67

68 **Table 1.** Detailed information of the four testing sites, i.e. Euc-plantation, Iowa-cropland, Beijing-urban,
69 and BCI-forest sites, including the location, precipitation, dry season period (< 100 mm of monthly
70 precipitation), spatial coverage, temporal coverage, and the number of accessed PlanetScope images. The
71 precipitation data is assessed from the Tropical Rainfall Measuring Mission (TRMM) data from 2000 to
72 2019.

73

Site	Location	Precipitation (mm yr ⁻¹)	Dry season period	Spatial coverage (km ²)	Temporal coverage	Number of accessed PlanetScope images
Euc-plantation	22°58'S, 48°44'W	1488	Apr-Sep			129
Iowa-cropland	42°20'N, 92°57'W	1043	NA	20×20	Jan-Dec, 2018	52
Beijing-urban	40°01'N, 116°29'E	492	NA			39
BCI-forest	9°06'N, 79°50'W	2052	Jan- Apr			51

74

75 **Table 2.** The number of pixels within each of the seven gap size categories over the four study sites.

Gap size category	Euc-plantation	Iowa-cropland	Beijing-urban	BCI-forest
#1	12,798	43,563	37,565	47,504
#2	27,131	72,346	85,952	78,936
#3	57,988	140,179	169,222	158,086
#4	114,959	181,624	307,798	214,020
#5	251,583	286,482	537,434	381,917
#6	544,319	443,563	815,649	744,642
#7	938,344	937,216	1,101,519	918,274

76

77

78 **Table 3.** Accuracy assessments (mean and standard deviation) across all seven gap size categories and all
 79 four sites using original images as benchmarks. RMSE, CC, and SSIM are used to evaluate the accuracy of
 80 gap-filling results respectively derived from NSPI, ARRC, SAMSTS, and OCBGF.

Method	NSPI		ARRC		SAMSTS		OCBGF	
	Mean	SD	Mean	SD	Mean	SD	Mean	SD
RMSE	0.0113	0.0058	0.0092	0.0056	0.0090	0.0041	0.0065	0.0027
CC	0.87	0.15	0.89	0.13	0.90	0.11	0.95	0.06
SSIM	0.86	0.16	0.89	0.14	0.89	0.11	0.94	0.06

81

82

83

84 **Table 4.** Accuracy assessments (mean and standard deviation) across all the areas with land cover changes
 85 and all four sites using original images as benchmarks. RMSE, CC, and SSIM are used to evaluate the
 86 accuracy of gap-filling results respectively derived from NSPI, ARRC, SAMSTS, and OCBGF.

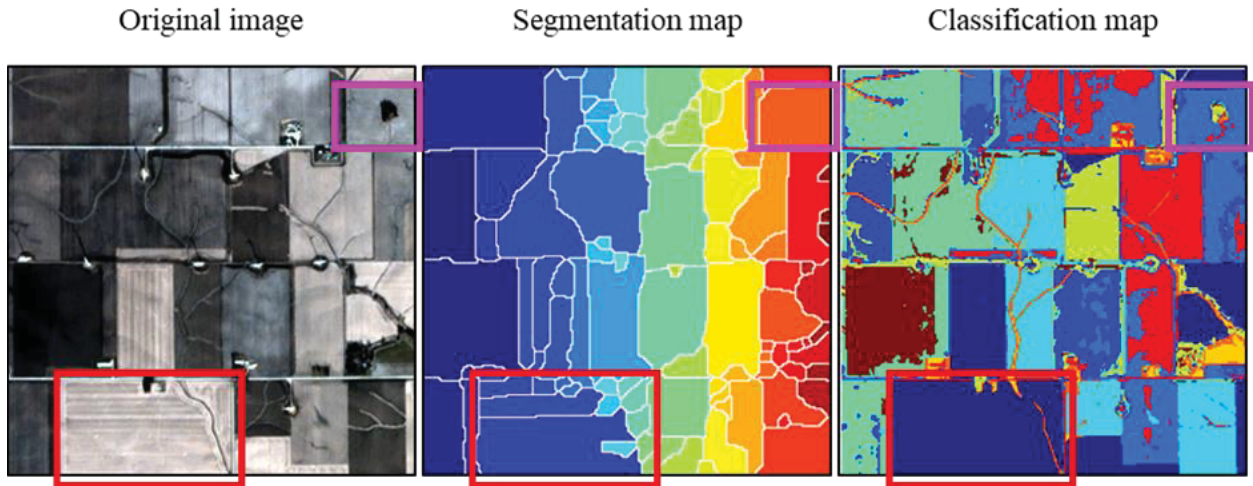
Method	NSPI		ARRC		SAMSTS		OCBGF	
	Mean	SD	Mean	SD	Mean	SD	Mean	SD
RMSE	0.0120	0.0035	0.0119	0.0034	0.112	0.0040	0.0082	0.0023
CC	0.89	0.055	0.88	0.094	0.90	0.087	0.95	0.031
SSIM	0.89	0.055	0.88	0.098	0.90	0.094	0.94	0.035

87

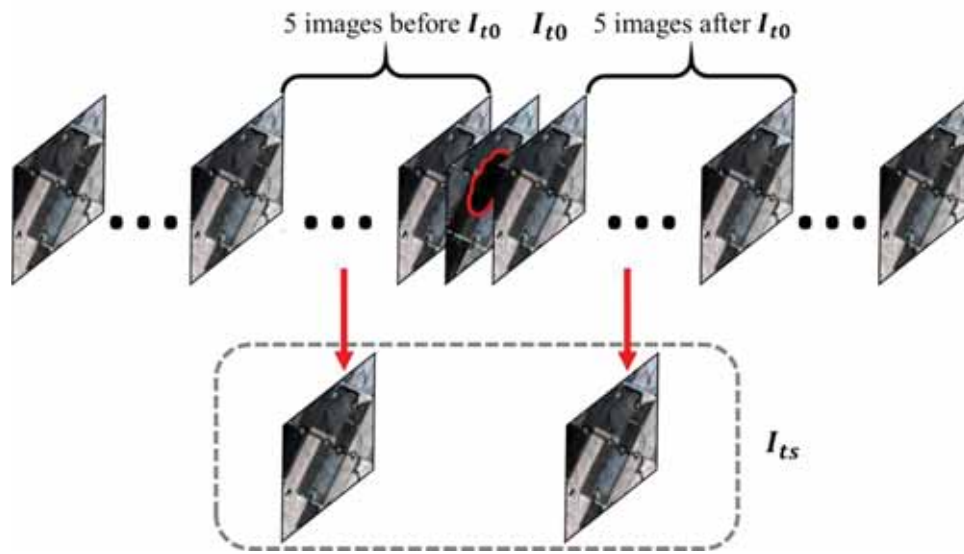
88

89

90 **Figure S1.** Example demonstration of the spatial heterogeneity of land cover in nature, including the
91 original image (November 27, 2018 at the Iowa-cropland site), segmentation map, and classification map.
92 The red squares highlight the situation of multiple objects within the same class and the magenta squares
93 highlight multiple classes within the same object.



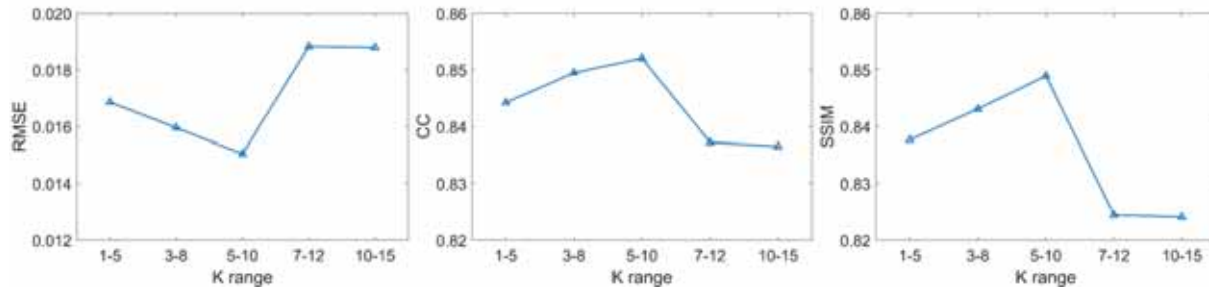
95 **Figure S2.** Example demonstration of how the composited cloud-free time series I_{ts} is generated for the
96 automatic object segmentation and classification, which represents Step 1 of Task 2 as shown in Fig. 2. I_{ts}
97 is automatically determined by searching the clearest images ($n=2$) from each 5-image group before/after
98 the target image I_{t0} .



99

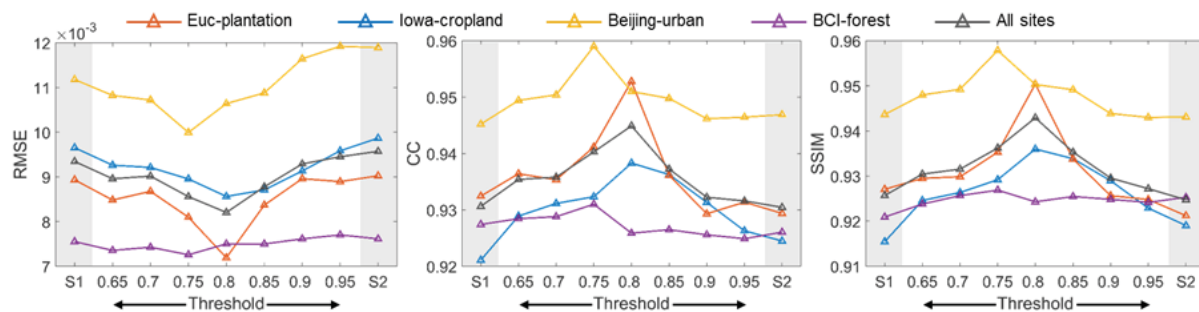
100

101 **Figure S3** Sensitivity analysis for determining the optimal range of the class number (K) using the Iowa-
102 cropland site as an example. Specifically, the five different K ranges were set, including 1-5, 3-8, 5-10, 7-
103 12, and 10-15. We used seven testing images acquired on the same dates as Fig. 7 and masked out areas
104 using the simulated mask with a percent data missing of 64% as shown in Fig. 6a.



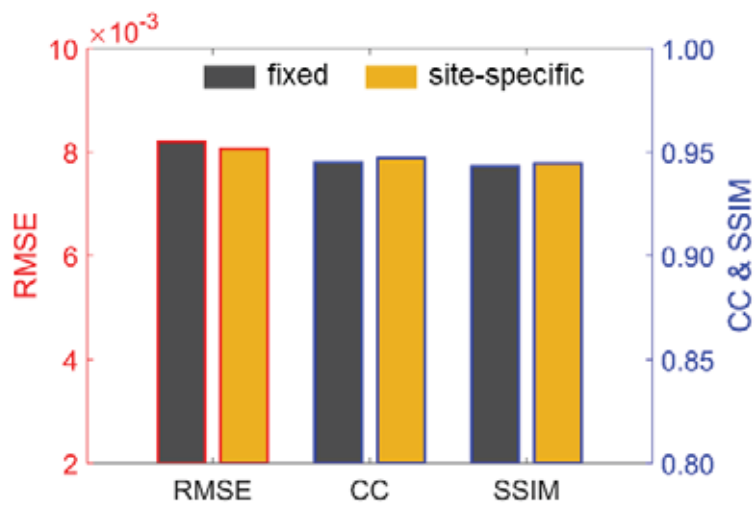
105

106 **Figure S4.** Sensitivity analysis for determining the optimal threshold value that helps best separate the two
 107 gap-filling scenarios as shown in Task 3 of Fig. 2. Specifically, the threshold value was set to vary from
 108 0.65 to 0.95, and accuracies were assessed respectively on a site basis (color lines) and across all sites (grey
 109 color line). And the two independent scenarios (i.e. S1 and S2 shown in shading areas) were also separately
 110 tested. Three metrics for accuracy assessments include RMSE, CC, and SSIM. The data used for this test
 111 is the same as the data shown in Fig. 7 in the main text.



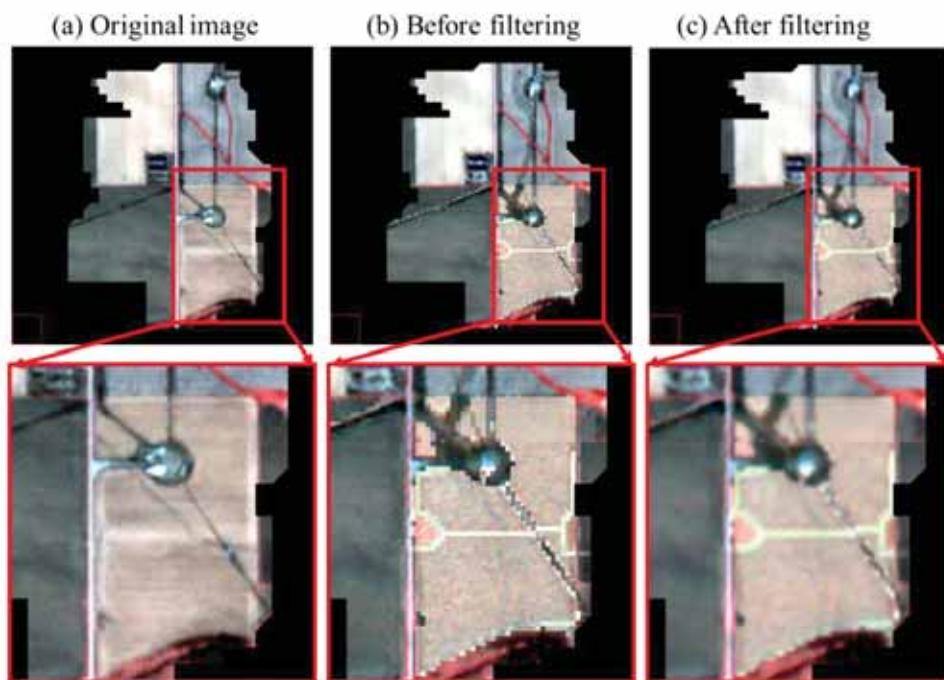
112

113 **Figure S5.** Comparison of gap-filled results across all sites using a fixed (i.e. 0.8) and site-specific
114 optimal threshold value. The overall accuracy is very comparable across three metrics, including RMSE
115 (0.0082 vs. 0.0081), CC (0.94 vs. 0.95), and SSIM (0.94 vs. 0.94).



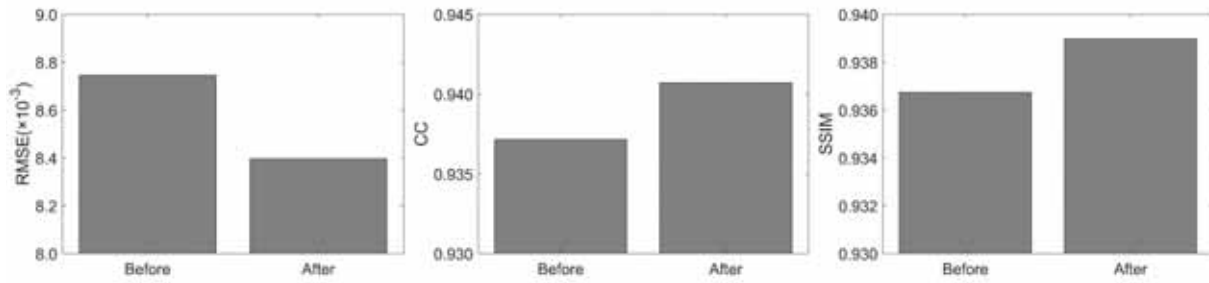
116

117 **Figure S6.** Example demonstrating the effectiveness of the post-image-processing (Task 4; Fig. 2). The
118 original image (a) and the gap-filled results derived before (b) and after applying the guided filter (c) on
119 November 14, 2018 at the Iowa-cropland site are shown in the upper panel. The corresponding magnified
120 areas in the red squares of the upper panel are shown in the lower panel.



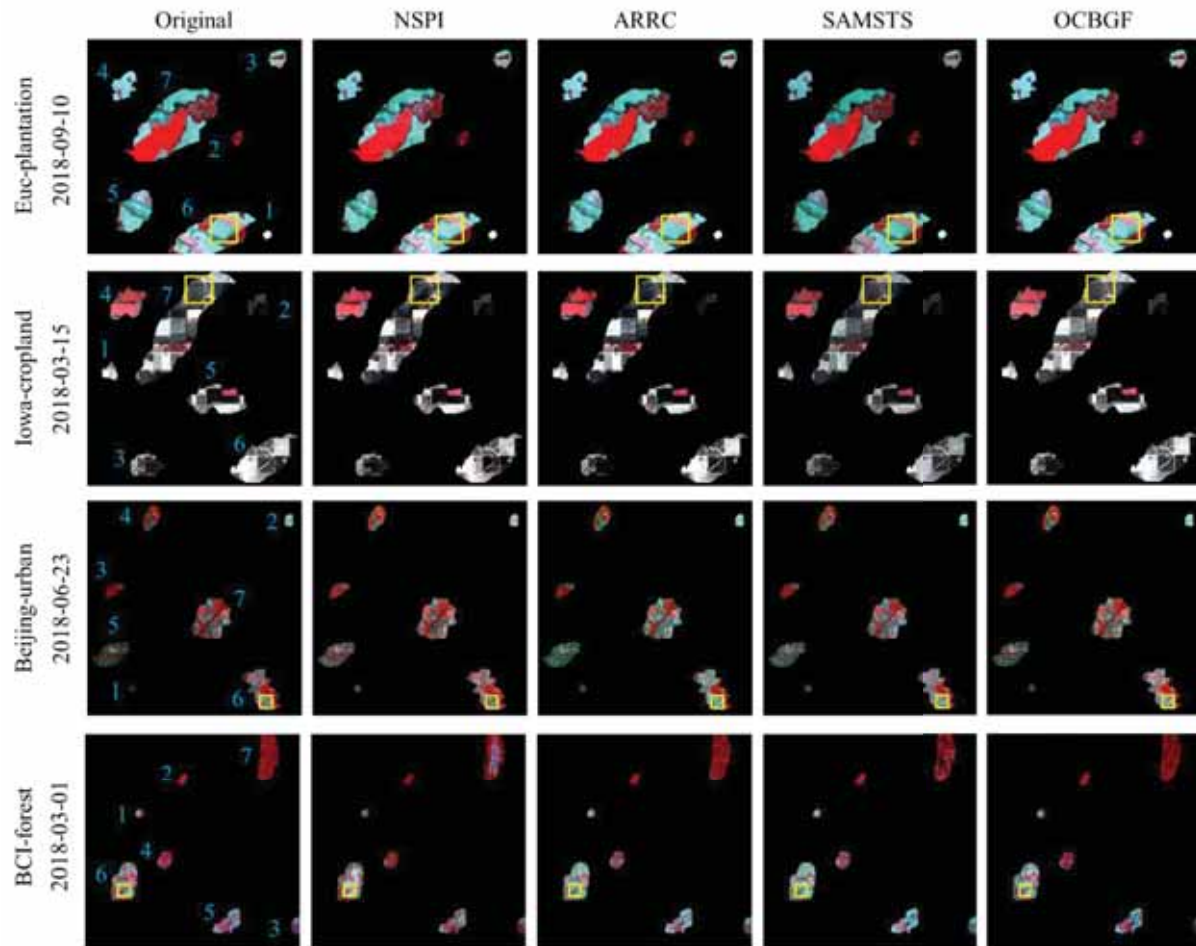
121
122

123 **Figure S7.** Results assessing the accuracy of gap-filled results derived before and after applying the guided
124 filter, using the Iowa-cropland site as an example. All seven testing images covering areas with land cover
125 changes and across different seasons in 2018 were used here.



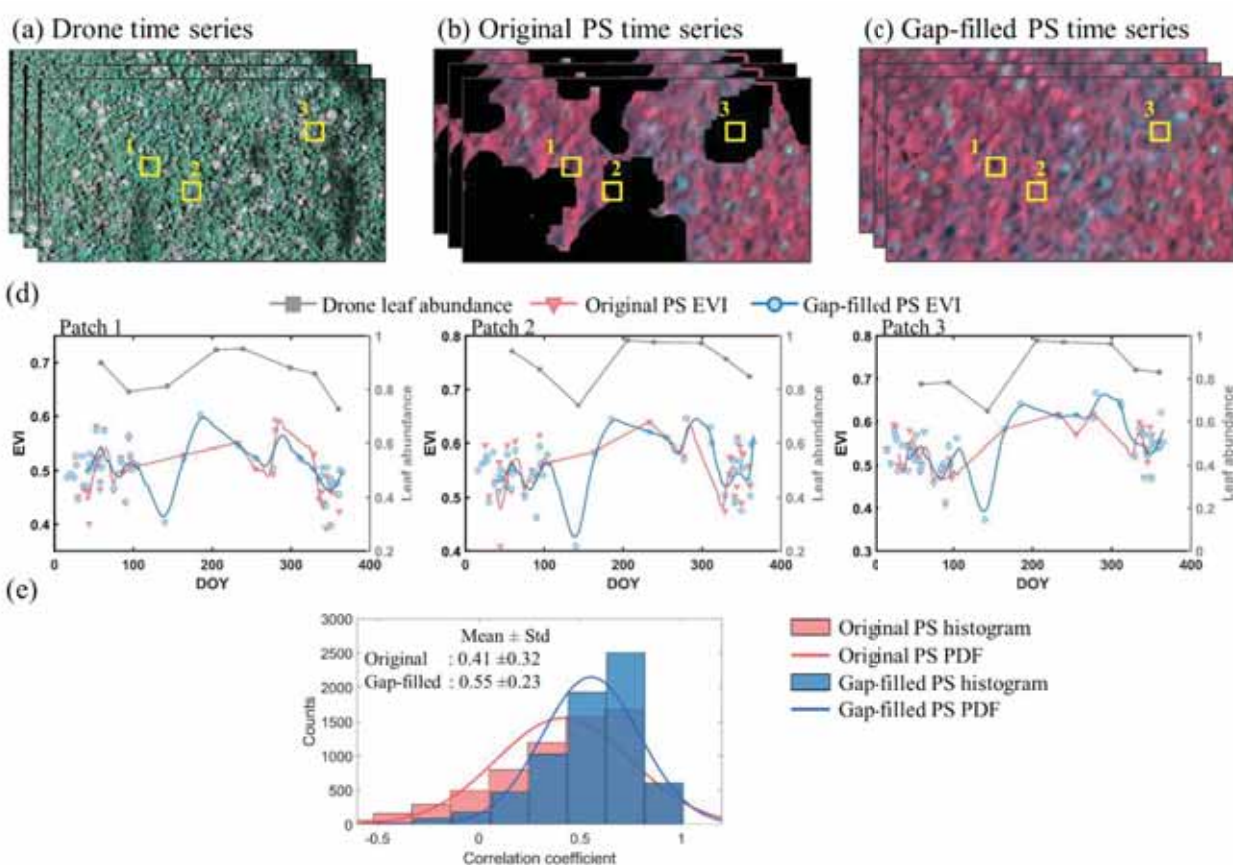
126

127 **Figure S8.** Assessing the effect of different gap sizes (Table 2) on the gap-filled results derived from
 128 NSPI, ARRC, SAMSTS, and OCBGF methods using original images as benchmarks. False color
 129 composites (RGB=NIR-Red-Green) of selected PlanetScope images and corresponding gap-filling results
 130 using NSPI, ARRC, SAMSTS, and OCBGF are shown below. Yellow squares are magnified areas shown
 131 in Fig. 5 in the main text.



132

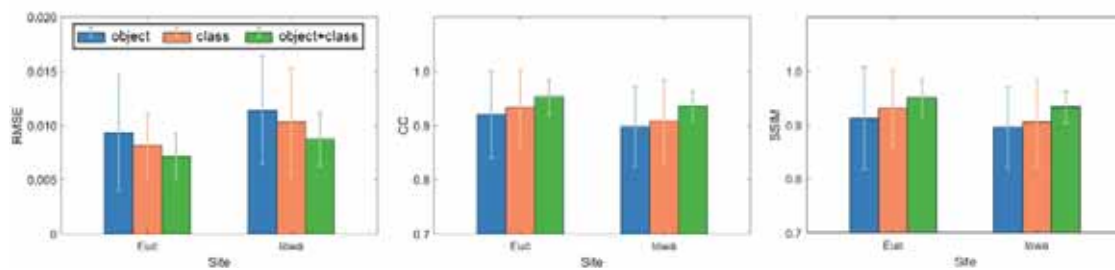
133 **Figure S9.** An example demonstration of whether gap-filled PlanetScope time series could boost the patch
 134 scale (12m×12m) leaf phenology monitoring in a 50-ha plot of a moist forest landscape at the BCI-forest
 135 site. (a) True color composites (RGB=Red-Green-Blue) of drone time-series images ($n=8$; Araujo et al.,
 136 2021), (b) false color composites (RGB=NIR-Red-Green) of original PlanetScope (PS) time-series images
 137 ($n=51$) with masked clouds/cloud shadows (black areas), and (c) gap-filled PS time-series images using
 138 OCBGF. Yellow squares are three representative patches that are affected by serious cloud contaminations.
 139 (d) Seasonal variability in patch-scale leaf abundance (gray) derived from the BCI drone images (Park et
 140 al. 2019), and enhanced vegetation index (EVI) derived respectively from original (pink) and gap-filled
 141 (blue) PS. (e) The histogram and probability distribution function (PDF) of the correlation coefficients (CCs)
 142 between the drone-derived leaf abundance seasonality and the two versions of EVI seasonality derived
 143 respectively from original and gap-filled PS time-series images, across all patches. The considerably higher
 144 average CCs value of gap-filled PS suggests that gap-filled PS could better track leaf phenology at this site
 145 compared with drone image time series.



146

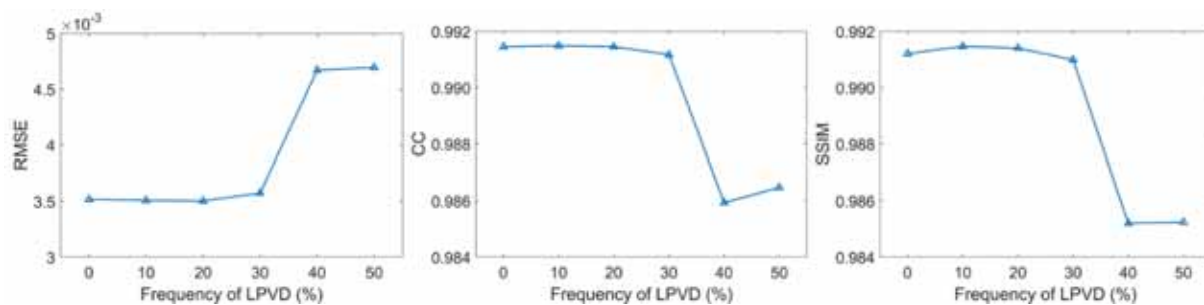
147

148 **Figure S10.** Effectiveness assessments on the three approaches (i.e. object alone, class alone, object-class)
149 for gap-filling at Euc-plantation and Iowa-cropland sites, with three accuracy metrics of RMSE, CC, and
150 SSIM.



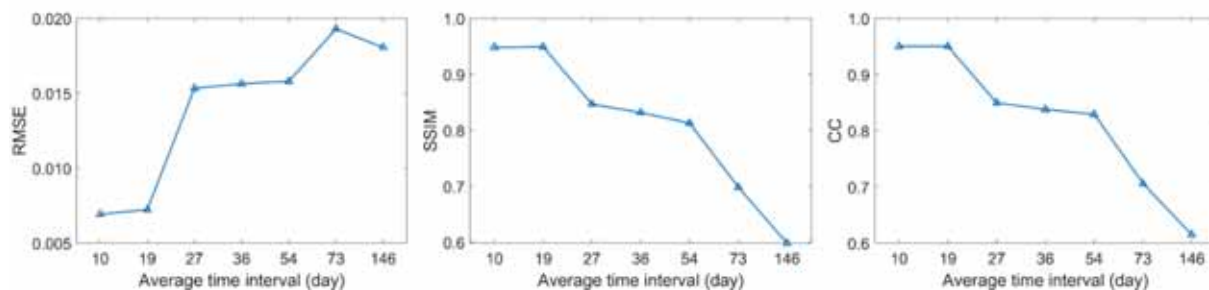
151

152 **Figure S11.** Sensitivity analysis on examining the effect of the frequency of low percent valid data (LPVD;
153 with percent valid data less than 10%) on the gap-filled results, using the Euc-plantation site as an example.
154 The image acquired on September 10, 2018 (the same as the first image in Fig. 5) was selected as the target
155 image for gap-filling, and nine temporally-adjacent images (i.e. five before and four after) were selected
156 from the whole image time series. The sensitivity analysis was performed in the six cases with the frequency
157 of LPVD ranging from 0% to 50% with an interval of 10%.



158

159 **Figure S12.** Sensitivity analysis on examining the effect of clear-sky observation frequency on the OCBGF
160 gap-filled results, using the Iowa-cropland site as an example. A total of 35 clear-sky images in 2018 were
161 identified, and the middle (acquired on July 02, 2018) of this time series was selected for gap-filling. The
162 sensitivity analysis was performed in the following 7 cases, which respectively selected once every N
163 images throughout the entire image time series, where $N=1, 2, 3, 4, 6, 8,$ and $16,$ resulting in an average
164 time interval of 10, 19, 27, 36, 54, 73, and 146 days, respectively. Our results show that the gap-filled
165 accuracies remain high when the average time interval is less than 19 days, but rapidly decrease afterward
166 with a reduction in clear-sky observation frequency.



167