

Characterization of a Smog Chamber for Studying Formation of Gas-phase Products and Secondary Organic Aerosol

Qi Yuan¹, Zhuozhi Zhang¹, Meng Wang¹, Kin Fai Ho², Tao Wang¹, Shuncheng Lee^{1*}

¹ *Department of Civil and Environmental Engineering, The Hong Kong Polytechnic University*

² *School of Public Health and Primary Care, The Chinese University of Hong Kong, Shatin, Hong Kong, China*

*To Whom Correspondence Should be addressed

E-mail: shun-cheng.lee@polyu.edu.hk

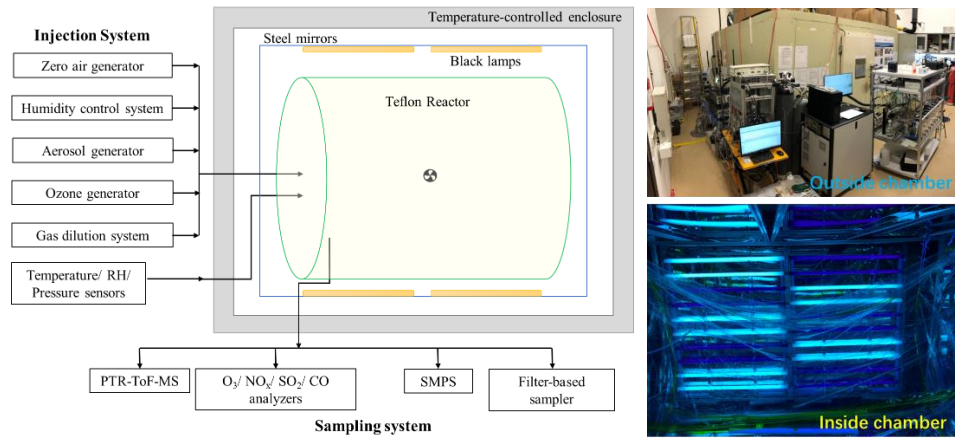
Tel: +852 27666011

Highlights

A smog chamber was designed and comprehensively characterized.

The wall loss of gases and particles are smaller than other chambers.

The chamber can provide high quality data for gas-phase oxidation and SOA formation.



Conflict of Interest

The authors confirm that there is no conflict of interest to be declared.

1 **Characterization of a Smog Chamber for Studying Formation of Gas-phase**
2 **Products and Secondary Organic Aerosol**

3

4 Qi Yuan¹, Zhuozhi Zhang¹, Meng Wang¹, Kin Fai Ho², Tao Wang¹, Shuncheng Lee^{1,*}

5

6 1. Department of Civil and Environmental Engineering, The Hong Kong Polytechnic
7 University

8 2. School of Public Health and Primary Care, The Chinese University of Hong Kong,
9 Shatin, Hong Kong, China

10

11

12 **Abstract:** Smog chambers provide a potent approach to explore the secondary organic
13 aerosol formation under varied conditions. This study describes the construction and
14 characterization of a new smog chamber facility for studying the formation mechanisms
15 of gas-phase products and secondary organic aerosol from the photooxidation of
16 volatile organic compounds. The chamber is a 5.4 m³ Fluorinated Ethylene Propylene
17 (FEP) Teflon reactor with the potential to perform photooxidation experiments at
18 controlled temperature and relative humidity. Detailed characterizations were
19 conducted for evaluation of stability of environmental parameters, mixing time,
20 background contamination, light intensity, and wall losses of gases and particles. The
21 photolysis rate of NO₂ (J_{NO_2}) ranged from $(1.02\text{--}3.32) \times 10^{-3} \text{ sec}^{-1}$, comparable to the
22 average J_{NO_2} in ambient environment. The wall loss rates for NO, NO₂, and O₃ were
23 0.47×10^{-4} , 0.37×10^{-4} , and $1.17 \times 10^{-4} \text{ min}^{-1}$, while wall loss of toluene was obsolete
24 found in a 6-hours test. The particle number wall loss rates are $(0.01\text{--}2.46) \times 10^{-3} \text{ min}^{-1}$
25 for 40–350 nm with an average lifetime of more than one day. A series of toluene
26 photooxidation experiments were carried out in absence of NO_x under dry conditions.
27 The results of the simulation experiments demonstrated that the chamber is well
28 designed to simulate photolysis progress in the atmosphere.

29 **Keywords:**

30 Smog chamber

31 Gas-phase oxidation

32 Chamber characterization

33 Chemical mechanism

34

35 -----

36 *Corresponding author. E-mail: shun-cheng.lee@polyu.edu.hk (S.C. Lee)

37

38 **Introduction**

39 Atmospheric chemical processes of volatile organic compounds (VOCs) lead to a
40 variety of secondary oxidized products in the gas phase and particle phase. Such
41 pollutants are often harmful to human health and play important roles in climate change
42 and global atmospheric chemistry (Hallquist et al., 2009; Harrison and Yin, 2000). In
43 real atmosphere, the chemical processes are not easily studied since atmospheric
44 pollutants would undergo not only chemical production and loss, but also emission,
45 transportation, and deposition (Brasseur, 2017). To isolate the chemical processes from
46 other atmospheric processes, smog chambers have been developed to better investigate
47 the near-ground ozone (O₃) formation, mechanisms of secondary organic aerosol (SOA)
48 formation, kinetics and mechanisms of multiphase reactions, and ageing processes of
49 primarily emitted pollutants (Babar et al., 2017; Carter et al., 2005; Cocker et al., 2001;
50 Karl et al., 2004; Paulsen et al., 2005; Saathoff et al., 2003; Wang et al., 2011).

51 With the aid of smog chamber experiments, the VOC oxidation mechanisms, the
52 key factors affecting oxidation, and the SOA yield were revealed. In earlier stages, the
53 chambers were aimed to simulate atmospheric processes under near-real atmospheric
54 conditions to promote the development of air quality models. Based on the results from
55 chamber studies on the oxidation of large quantities of VOCs, several atmospheric
56 chemical mechanisms were developed, including Statewide Air Pollution Research

57 Center Mechanism (SAPRC), Regional Atmospheric Chemical Mechanism (RACM),
58 and Master Chemical Mechanism (MCM) (Chu et al., 2021). Among those mechanisms,
59 reactions of hydroperoxyl radical (OH) with VOCs were dominant removal processes,
60 determining the atmospheric lifetime of VOCs. The formation of alkyl radicals, alkoxy
61 radicals, alkyl peroxy radicals (RO₂) and Criegee intermediates were considered as the
62 main transformation pathways of VOCs in the atmosphere (Hallquist et al., 2009). With
63 the innovation of detection techniques, more generations of oxidation products with
64 different volatility were identified in molecular levels, which facilitates the
65 identification of new mechanisms and formation pathways (e.g., new particle formation
66 and autooxidation) (Guo et al., 2020; Pye et al., 2019). Despite the extensive chamber
67 studies on oxidation of VOCs, there are great divergences in the reaction pathways and
68 SOA yields. Taking aromatic hydrocarbons as an example, many new oxidation
69 mechanisms were proposed, e.g., formation of polyhydroxyl compounds from phenolic
70 compounds, autooxidation of highly oxygenated compounds, and hydrogen transfer of
71 aldehydes (Li et al., 2021b; Srivastava et al., 2022). These uncertainties of mechanisms
72 causes the yield of intermediates (e.g., phenol, glyoxal, and methylglyoxal) differs up
73 to 10 times in different studies, which leads to 25%~50% loss of organic carbon in
74 aromatic SOA (Song, 2021). Furthermore, the oxidation durations in past chamber
75 studies are less than the equivalent of 1 day at typical atmospheric oxidation
76 concentrations, which may not fully characterize the production and ageing of
77 secondary organic compounds associated with multiple generations of oxidation
78 (Garmash et al., 2020). Thus, a better understanding of gaseous oxidation of VOCs and
79 SOA chemical properties is required to improve simulation of missing SOA in
80 atmospheric models through laboratory studies.

81 The construction of smog chambers varied widely in terms of the operating
82 environments, chamber sizes, light sources, and reactor materials to reach different
83 research purposes. Regarding the light sources used and chamber size needed, the smog
84 chambers were usually constructed in indoor and outdoor environment. Taking

85 advantage of the natural solar radiation and with no size limitation from indoor
86 environment, outdoor chambers can be as large as 270 m³ (Rohrer et al., 2005), while
87 indoor chambers are built with a size range of 1–90 m³ (Wagner et al., 2011; Wu et al.,
88 2007). Comparing with that of outdoor chambers, the operating conditions e.g.
89 (temperature, relative humidity (RH), pressure and light intensity) of indoor chambers
90 can be rapidly, precisely, and stably controlled. This would lead to more atmospheric
91 relevant conditions for simulation experiments and less uncertainty of experiment
92 results (Li et al., 2021a; Wagner et al., 2011). Meanwhile, the larger surface to volume
93 ratios of small chambers enhance the wall effects to some extent. Light sources used in
94 indoor chambers are normally solar, arc and black lights (White et al., 2018). Solar and
95 arc lights are preferred due to their similarity to sunlight, while blacklights are often
96 used for their price advantage. Recent-built chambers combine different types of black
97 lamps with varied wavelength range (280–400 nm) as light sources (Chu et al., 2021).
98 Although these ultraviolet (UV) lights have different spectra with solar radiation, their
99 photolysis rates can be characterized to well simulate the relevant atmospheric
100 conditions. It is worth noting that in some cases the reaction rates would be affected by
101 light sources with different wavelength ranges with solar light (Carter et al., 2005).
102 Common chamber walls or reactors are made of stainless steel, aluminum, Pyrex, quartz,
103 PFA (perfluoroalkoxy alkanes) Teflon, and FEP (fluorinated ethylene propylene) Teflon.
104 Due to the good chemical resistance, high light transmission rate, and adjustability of
105 volume, FEP Teflon film is most used both as reactor material and as coating material
106 inside metallic reactors (Bloss et al., 2005a; Chu et al., 2021; Shao et al., 2022; Wang
107 et al., 2014).

108 In this study, a new indoor photochemical chamber was built in the air laboratory
109 at the Hong Kong Polytechnic University (PolyU chamber) to study the formation
110 mechanisms of gas-phase products and SOA from photooxidation of aromatic VOCs.
111 Characterization experiments on stability of environmental conditions, mixing time,
112 purity of zero air, photolysis rates of nitrogen dioxide (NO₂), and wall loss of gases and

113 particles were conducted. In addition, photooxidation experiments on a typical aromatic
114 hydrocarbon, toluene, in the absence of nitrogen oxides (NO_x) were conducted to
115 examine the performance of the chamber.

116

117 **1 Description of the PolyU chamber**

118 The PolyU chamber is operated in batch mode, in which the composition of the gaseous
119 precursors, aerosol seed, UV lights, temperature, and RH are controlled throughout the
120 duration of several hours. It is composed of three parts: the inner Teflon reactor with
121 frame, the in-between light sources on stainless steel mirrors, and the outer temperature-
122 controlled enclosure. It is equipped with a set of air purification system, gas/aerosol
123 generators, analytical instruments for gases and VOCs, and particle analysis
124 instruments. The schematic of the chamber along with the instruments is shown in Fig.
125 1.

126

127 **1.1 Enclosure and environmental condition control**

128 There are two nested enclosures outside the Teflon reactor. A 3.2 m × 3.2 m × 2.5 m
129 rectangular enclosure comprises a stainless-steel framework with air conditioning ducts
130 is the outermost shell of the chamber. Between the reactor and the walls of the chamber,
131 six polished stainless-steel mirrors surrounding the reactor enable a maximum and
132 homogeneous light intensity. The air conditioning system on the enclosure controls the
133 temperature in a range of 10–40°C, with an accuracy of ±1°C (Tan et al., 2021). A
134 humidity control system with zero air generator and purified water in a water tank is
135 connected into the reactor with Teflon tubing, providing a wide humidity range of 5%–
136 80%. Three sensors from Beijing Convenient Environmental Tech Co. Ltd are inserted
137 into the Teflon reactor to monitor the temperature, RH, and air pressure.

138

139 **1.2 Teflon reactor**

140 The reactor in this study was made of 75 μm FEP Teflon material which is chemically
141 inert, resistant to most chemicals, and with high transmittance ($> 95\%$) of UV light. It
142 has a volume of 5.4 m^3 ($\text{Ø } 1.7 \text{ m} \times 2.4 \text{ m}$) and a surface to volume ratio of 1.59 m^{-1} .
143 Compared with stainless steel and quartz, the weakness of FEP film is the accumulation
144 of electrostatic charge which can increase the wall loss rates for particles with a
145 diameter smaller than 500 nm (McMurry and Grosjean, 1985). To reduce the effect of
146 electrostatic charge on particle wall loss, two ionizing air blowers are installed outside
147 the reactor at diagonal positions to keep the air flow evenly across the surface of the
148 reactor. The effects of ionizing air blowers were discussed in Section 3.4. The reactor
149 is compression-sealed and suspended in the enclosure and mounted within the rectangle
150 frame. It is flexible to be moved horizontally to maintain a differential positive pressure,
151 minimizing contamination and dilution from air outside. A pressure sensor is used to
152 measure the pressure inside the reactor with an accuracy of $\pm 3 \text{ Pa}$. The experiments
153 lasted for 4 to 6 hours before the reactor volume decreases to 2/5 of its maximum value.
154 There are seven Teflon ports for injection of purified air, VOCs, oxidants (H_2O_2 and
155 O_3), other gases (e.g., NO_x), sampling of gases for online instruments, and vacuuming
156 the chamber for cleaning. Two additional stainless-steel ports connected with
157 conductive silicone tube are for injection and sampling of particles. One stainless steel
158 fan coated with Teflon is installed at the bottom to accelerate the mixing of reactants.
159 No detectable VOCs are emitted from the fans.

160

161 **1.3 Light sources**

162 To simulate photochemical processes, a total of 42 UV lights are arranged on two
163 opposite sides of the stainless-steel mirrors as the light source to approximate the
164 atmospheric actinic spectrum. Forty of the black lights (36 W, Philips and 36 W,
165 Huaqiang) provide UV-A type light with wavelength ranges of 320–400 nm (center at
166 340 nm) and 340–400 nm (center at 350 nm). Two UV-C type lamps (40 W, Philips)

167 provide lights with a center wavelength of 254 nm. The forty UV-A lamps are controlled
168 in 20 groups, and the two UV-C lamps can be controlled independently. Therefore, the
169 light intensity can be regulated into several levels. Due to the space limit of the outer
170 enclosure, the distance between the lights and the reactor is ~ 50 cm. The unwanted
171 heat generated from the irradiation source is removed by the air conditioning system on
172 the enclosure.

173

174 **1.4 Chamber air purification, conditioning, and injection system**

175 A zero-air generator (T701H, Teledyne API) provides clean and dry air at a flow of 20
176 L/min by passing compressed ambient air at ~ 30 psi. The zero-air generator contains a
177 dryer, an activated charcoal filter, and a chemical scrubber to provide air with free of
178 NO_x, O₃, CO, SO₂, VOCs and particles. The clean air can be conditioned by passing
179 through the humidifier, ozone generator, VOC liquid injection unit and aerosol mixing
180 tank before entering the chamber.

181 Gaseous reactants (e.g., NO_x and VOCs) are injected from commercially prepared
182 gas cylinders with target gases in nitrogen via a FEP Teflon line into the reactor, and
183 then flushed by purified dry air or nitrogen. For the injection of liquid reactants, certain
184 amount of the liquids with known concentrations are introduced into a U-type three-
185 way injection tube with microliter syringes, and then the liquids are heated and blew
186 into the chamber by pure nitrogen. Ozone is generated by a commercial ozone generator
187 (Model 2001; Jelight Company, USA) with purified air from the zero-air generator.
188 Seed particles are generated by an atomizer (Model 3079; TSI Incorporated, USA) with
189 (NH₄)₂SO₄ solution through conductive silicon tubing into the reactor. The flow rate
190 and gas volume of injection are controlled with mass flow controllers.

191

192 **1.5 Instrumentation**

193 The chamber is equipped with advanced instrumentations to measure the gaseous and
194 particulate chemical composition, including trace gases (O₃, NO_x and NO_y), VOCs,

195 oxygenated VOCs (OVOCs), particle number and size distribution, and particulate
196 oxygenated products. The species/ parameters to be measured and details on the
197 instrumentation are summarized in **Table 1**. Briefly, NO_x and O₃ are measured by
198 Teledyne T series with the accuracy of 0.5%. A large variety of VOCs and OVOCs are
199 quantified in real-time by a PTR-TOF-MS (QiTOF, IONICON, Austria) with a high
200 sensitivity (< 10 pptv) and fast response time (1 sec). With the selective reagent
201 ionization function, reagent ions in PTR-TOF-MS can be switched among H₃O⁺, O₂⁺
202 and NO⁺, thus enhancing the analyzable compound classes and increasing selectivity.
203 OVOCs with more functional groups can be measured by a high-resolution time-of
204 flight chemical ionization mass spectrometer (HR-ToF-CIMS, Aerodyne) with nitrate
205 or iodide as reagent ions. In addition to the OVOCs, oxygenated products in SOA are
206 measured with online and offline approaches. A filter-inlet for gases and aerosols
207 (FIGAERO) is optionally coupled to the HR-ToF-CIMS to measure the semi- and low-
208 volatile organic compounds in particulate matters. Offline samples are collected with a
209 custom-built sampler with 47 mm filters at 20 L/min, and then analyzed with an
210 Orbitrap Fusion Lumos mass spectrometer coupled with a liquid chromatography (LC-
211 Orbitrap MS, Thermo Scientific). Aerosol size and number distribution are monitored
212 by a scanning mobility particle sizer (SMPS) which consists of a differential mobility
213 analyzer (3082 DMA, TSI, USA) and a condensation particle counter (3776 CPC, TSI,
214 USA). Environmental parameters, i.e., temperature, RH, and pressure, are monitored
215 with sensors from Beijing Convenient Environmental Tech Co.Ltd.

216

217 **2 Characterization of the chamber**

218 **2.1 Stability of RH, temperature, and control of mixing**

219 Temperature and humidity are two important environmental parameters during
220 chemical processes. The temperature inside the reactor is controlled by the air
221 conditioning system of the chamber enclosure by presetting a targeted value to keep a
222 stable experimental condition and avoid overheating by UV lights. Water vapor is

223 flushed into the chamber by bubbling deionized water with a certain flow of purified
224 air until RH reaches set point. A temperature of 25°C and a RH of 50% were set to
225 examine the stability of the system with and without UV light irradiation. **Fig. 2(a)**
226 shows the evolution of temperature and RH during the test. The temperature and
227 humidity reached the set point during the first 20 minutes and then remained stable
228 under dark conditions. Once all the 42 lights were turned on, temperature increased
229 from 25°C to ~27°C in 10 minutes and then stayed at $26.0 \pm 0.4^\circ\text{C}$. The increase of
230 temperature caused the slight drop (1%) of RH. The results showed that although there
231 is unwanted heating when lights are turned on, temperature and RH can be well
232 controlled for stable and repeatable experiments.

233 Mixing time of gases were tested with NO_2 and toluene as tracers. Before the
234 injection of gases, the mixing fan was turned on with full speed. Considering the
235 volume of the chamber and the mixing ratios of gases, 177 ppm NO_2 and 32 ppm
236 toluene were injected into the chamber at 1.22 L/min and 5.6 L/min for 3 minutes to
237 reach ~120 ppb and ~100 ppb, respectively. **Fig. 2(b)** shows that the mixing of gases is
238 within 5 minutes from the beginning of the injection, comparable to that of few minutes
239 in other smog chambers (Shao et al., 2022; Wang et al., 2014).

240

241 **2.2 Background contamination of the chamber**

242 After cleaning procedure of the chamber, the background level of contaminants in the
243 chamber should be examined and kept as low as possible, especially for the tests carried
244 out with atmospherically relevant concentrations. Hydrocarbons and OVOCs in the
245 chamber were below detection limit of PTR-TOF-MS. Results of air purity test in
246 chamber are shown in **Fig. 3**. NO_x and O_3 are detected to be lower than 1 ppb after
247 cleaning. And particle number concentration inside the chamber is less than 10 cm^{-3} in
248 number and $\sim 0 \mu\text{g}/\text{m}^3$. After seven hours photooxidation of zero air, minimal particle
249 formation of number concentration at $< 10^3 \text{ cm}^{-3}$ and mass concentration $< 0.6 \mu\text{g}/\text{m}^3$
250 was observed, respectively. Approximately 4 ppb of O_3 was generated after seven hours

251 irradiation, with a formation rate of 0.63 ppb/hr. The formation rates of SOA and O₃
252 were comparable to previous studies on characterization of chambers (Babar et al., 2017;
253 Hu et al., 2014; Ma et al., 2022; Wu et al., 2007). The formation of particle
254 concentration and O₃ from photolysis of zero air is minimal compared with SOA formed
255 in normal experimental conditions (> 20 µg/m³), proving that pollutants levels in zero
256 air would cause negligible effects on experimental results in our study.

257

258 **2.3 Light intensity**

259 The artificial radiation has a broad radiation distribution owing to the chosen
260 combination of illumination sources, producing irradiation over the wavelength range
261 200-400 nm. Light intensity is represented by the photolysis rate of NO₂, which is
262 estimated through injecting NO₂ standard gas into the reactor, then irradiating the NO₂
263 in chamber with varied light intensities, and measuring the concentrations of NO, NO₂
264 and O₃. The photolysis rate of NO₂ (J_{NO_2}) is estimated with the following equation

$$265 \quad J_{NO_2} = k_{NO+O_3} [NO][O_3]/[NO_2] \quad (1)$$

266 where $[NO]$, $[O_3]$ and $[NO_2]$ represent concentrations (molecule/cm³) of the gases, and
267 $k_{NO+O_3} = 1.95 \times 10^{-14}$ cm³/(mol·sec) (Atkinson et al., 2004) is the rate constant of O₃
268 and NO reaction. A series of NO₂ actinometry experiments were carried out with
269 different number of UV lights turned on. With the setting ranging from 25% (6 of 340
270 nm and 4 of 350 nm wavelength lamps) to 100% (all lamps) of the lights turned on, the
271 J values ranged from 1.02×10^{-3} sec⁻¹ to 3.32×10^{-3} sec⁻¹ (**Fig. 4**). The maximum NO₂
272 photolysis rate (3.32×10^{-3} sec⁻¹) in our chamber is comparable to those reported in
273 other chambers provided in **Table 2**, e.g., $(2.32-2.55) \times 10^{-3}$ sec⁻¹ in the 5 m³ chamber
274 at the Institute of Chemistry, Chinese Academy of Sciences dual reactor chamber
275 (ICCAS-DRC) (Wang et al., 2015), and 2.83×10^{-3} sec⁻¹ in the 7 m³ chamber at the
276 Kyungpook National University (Babar et al., 2017). The maximum J_{NO_2} value of our
277 chamber is comparable to the average ambient values during the late summer and
278 autumn, i.e., $(2.28 \pm 2.80) \times 10^{-3}$ sec⁻¹ and $(2.12 \pm 2.78) \times 10^{-3}$ sec⁻¹, and the lower J_{NO_2}

279 values with half lights turned on is comparable to the ambient values in winter ($(1.56 \pm$
280 $2.35) \times 10^{-3} \text{ sec}^{-1}$) in Hong Kong (Li et al., 2018), which shows that the light system is
281 capable of simulating the ambient photochemistry in different seasons.

282 The heat generated by the UV lights in indoor chambers and trapped in Teflon
283 reactors may affect the reaction rates of VOCs and the gases. With all the lamps on, we
284 carefully monitored the variation of temperature and used the control system to slightly
285 adjust it. The typical change of temperature in the reactor is less than 2°C after 6 hours
286 of use.

287

288 **2.4 Wall loss of gases and particles**

289 Adsorption of gases had been shown to be a substantial source of gas losses inside
290 Teflon bags which influences the gas-phase reactivity and SOA formation (Grosjean,
291 1985). In the PolyU chamber, the wall loss rates of NO, NO₂, O₃ and toluene were
292 investigated by injecting known concentrations of gases into the chamber and
293 measuring their concentration decay for an extended period under dark conditions (**Fig.**
294 **5**). Approximately 450 ppb NO, 120 ppb NO₂ and 500 ppb O₃, and 100 ppb toluene
295 were separately injected into the chamber and monitored for 6 hours allowing sufficient
296 time for a measurable decay. All the wall loss experiments of gaseous species were
297 conducted under the temperature and RH of ~25°C and < 10% respectively. The wall
298 loss rates of NO, NO₂, O₃ in our chamber are 0.47×10^{-4} , 0.37×10^{-4} , and 1.17×10^{-4}
299 min^{-1} , respectively. The concentration of toluene remained 96.491 ± 0.291 ppb during
300 the wall loss test. It suggested that negligible wall loss for toluene was observed in
301 PolyU chamber, which is similar to that reported by Wang et al. (2011), Hu et al. (2014),
302 and Babar et al. (2017). The wall loss rates of these gases in our chamber are 2–10 times
303 lower than those of other chambers (**Table 2**).

304 Particle deposition onto the surface of the reactor is recognized to be influenced by
305 diffusion, coagulation of the particles as a function of particle size, and the charged wall

306 materials (Grosjean, 1985). The particle number wall loss rate is described by first-
307 order kinetics as

$$308 \quad dN(d_p, t)/dt = -K_N(d_p)N(d_p, t) \quad (2)$$

309 where $N(d_p, t)$ is the particle number concentration, d_p is the diameter of the particle and
310 $K_N(d_p)$ is the particle number loss coefficient (Cocker et al., 2001). The $K_N(d_p)$ values
311 can be estimated from the decrease of particle number concentration versus time when
312 no new particles are formed or injected into the chamber. In this study, 0.5 mol/L
313 ammonium sulfate ((NH₄)₂SO₄, AS) solution was atomized to generate seed particles.
314 The AS particles were introduced into the reactor with a flow rate of 1.5 L/min for 10
315 minutes and diluted to ~8000 cm⁻³ to reduce the coagulation of smaller particles. The
316 changing of particle distribution is shown in **Fig. 6**. In general, the particle wall loss
317 rates of the PolyU chamber were $(0.01\text{--}2.46) \times 10^{-3} \text{ min}^{-1}$ within 40–350 nm,
318 corresponding to an average lifetime of ~25 hr. For particle diameters of 100 nm, the
319 wall loss rate is $0.2 \times 10^{-3} \text{ min}^{-1}$ (**Table 2**), comparable to that of the CESAM chamber
320 in France (Wang et al., 2011), but smaller than the value of $(1.33\text{--}4.3) \times 10^{-3} \text{ min}^{-1}$
321 reported in other chambers with similar volumes (Babar et al., 2017; Bae et al., 2003;
322 Ma et al., 2022; Smith et al., 2019; Wang et al., 2015; Wu et al., 2007). The lower wall
323 losses of gases and particles in this study is probably due to smaller surface-to-volume
324 area (1.59 m^{-1}) and the removal of static electricity from ionizing air blowers.

325

326 **2.5 Experiment of toluene photooxidation**

327 A sequence of experiments on OVOC and SOA formation from toluene photooxidation
328 has been performed to validate the utility of PolyU chamber for studying SOA
329 formation and atmospheric chemistry. This reaction system has been widely studied for
330 the yields of gas-phase products and SOA, and the impact of some factors SOA
331 formation (Chen et al., 2019; Hildebrandt Ruiz et al., 2015; Hinks et al., 2018; Ng et
332 al., 2007; Qi et al., 2020; Sato et al., 2007). The initial conditions and yield for each
333 experiment are presented in **Table 3**. All the experiments were conducted without

334 aerosol seed at a temperature of $25 \pm 0.5^\circ\text{C}$ and dry conditions ($\text{RH} < 10\%$). In general,
335 200 μL of H_2O_2 solution was injected into a U-type glass tube and blew into the chamber
336 by pure nitrogen, afterwards 73–155 ppb of toluene was injected into the chamber. After
337 turning the UV lights on, H_2O_2 was rapidly converted into OH radical, leading to the
338 decay of toluene and formation of OVOCs and SOA. Mass concentrations of SOA were
339 calculated from the corrected volume concentrations by setting the particle density as
340 1.4 g/cm^3 based on literature data (Nakao et al., 2011; Ng et al., 2007; Sato et al., 2007).

341 The concentrations of toluene and gas-phase oxidation products were tracked with
342 PTR-TOF-MS. **Fig. 7** shows a comparison of mass spectra from one typical test before
343 UV lights on and after at least 5 hours of irradiation. The mass peaks at m/z 31.017,
344 47.012, 61.028, 73.028, 77.022, 85.028, 99.007, 107.049, 109.065, and 123.044 were
345 assigned to CH_2O (formaldehyde), CH_2O_2 (formic acid), $\text{C}_2\text{H}_4\text{O}_2$ (acetic acid), $\text{C}_3\text{H}_4\text{O}_2$
346 (methylglyoxal), $\text{C}_2\text{H}_4\text{O}_3$ (peracetic acid), $\text{C}_4\text{H}_4\text{O}_2$ (butenedial/furanone), $\text{C}_4\text{H}_2\text{O}_3$
347 (maleic anhydride), $\text{C}_7\text{H}_6\text{O}$ (benzaldehyde), $\text{C}_7\text{H}_8\text{O}$ (cresol), and $\text{C}_7\text{H}_6\text{O}_2$ (*p*-methyl
348 benzoquinone). The simplified formation pathways of the products from toluene
349 photooxidation implemented in Master Chemical Mechanism (MCM) v3.3.1 are listed
350 in **Fig. 8**. Benzaldehyde was formed as a first-generation product from OH-initiated H-
351 abstraction from the methyl group of toluene, while the formation of products from
352 phenol, peroxide-bicyclic ring-opening, and the epoxy-oxy channels are initiated by
353 OH addition on the aromatic ring. Cresols are formed by elimination of hydroperoxyl
354 radical from phenol route, and the OH addition of it further produces *p*-methyl-
355 benzoquinone as a second-generation product. Peroxide-bicyclic routes generates
356 methyl substituted peroxide radicals, resulting in the direct formation of ring-opening
357 products, i.e., butenedial, furanone, and methylglyoxal, following the formation of
358 maleic anhydride as the second-generation products. Intermediate of epoxy route, 2,3-
359 Epoxy-6-oxo-4-heptenal (not detected), further reacts and for methylglyoxal and
360 formaldehyde as the second-generation products. Multi-generation oxidation of the
361 first- and second-generation of products sequentially forms small molecule carbonyls,

362 i.e., formaldehyde, formic acid, acetic acid, methylglyoxal, and peracetic acid.

363 The proposed yields of products from the four routes as implemented in MCM are
364 higher than that in this study. For instance, the yields of benzaldehyde and cresol are
365 0.5%–2% and 1%–6% in this study versus 7% and 18% in MCM, respectively. It should
366 be noted that MCM provides yields of stable, volatile early-generation products, which
367 means the yields of these OVOCs would be different in ageing processes (Bloss et al.,
368 2005b). To be specific, reactions of reactive OVOCs can be further oxidized into other
369 products, leading to lower gas-phase product yield. High yields of later generation
370 products, formic acid (17%–27%), acetic acid (12%–17%) and peracetic acid (2%–5%)
371 were observed in this study (**Fig. 9**). It agrees well with the fact that excess H₂O₂
372 injection results in the ageing processes in our tests and prompted the production of
373 multi-generation products. On the other hand, yield of intermediate products varied
374 under different NO_x conditions (Arey et al., 2009; Chen et al., 2019; Gómez Alvarez et
375 al., 2007). In this study, we performed the toluene photooxidation experiments in
376 absence of NO_x. It's shown by Song et al. (2021) that yield of methylglyoxal and
377 benzaldehyde could be 10–20 times lower under no NO_x conditions than that under
378 high NO_x conditions, which is in good agreement with the current study. Furthermore,
379 a near-explicit mechanism of toluene from MCM version 3.3.1 using the Frame-work
380 for 0-D Atmospheric Modeling (F0AM) was applied to simulate the toluene
381 photooxidation. The description of F0AM in detail can be seen in the supporting
382 information. The time series of the selected compounds from model and experimental
383 results were compared in **Appendix A Fig. S1** and **Fig. S2**. Toluene was consumed faster
384 in experimental measurements than in model results. As first-generation products,
385 benzaldehyde and cresols are predicted to be 4 times higher than in experimental results,
386 whilst the further generation products are much higher in experimental results than that in
387 model results. These trends indicate that the oxidation was accelerated and aged in our
388 experiment. It should be noted that the J_{NO₂} input in F0AM was acquired from experiments
389 on photolysis rate of NO₂ in chamber. Different from other reactions, NO₂ cannot be

390 photolyzed by UV light under 300 nm. The two 254 nm wavelength UV lights would be an
391 important cause of the discrepancy between model and experimental results.

392 As shown in **Fig. 9**, slope of the growth curve of OVOC yields as a function of
393 OVOC mass concentration (M_0) showed dependence on initial toluene concentration.
394 With higher initial toluene concentration, the first-generation products (benzaldehyde
395 and cresols) showed higher yields with a much faster rate of growth. After some
396 irradiated time, the yield of benzaldehyde and cresols decreased with the increment of
397 M_0 , which refers to the consumption of these OVOCs and potential production of
398 further generation products and SOA (Chen et al., 2019). As later-generation products,
399 acids (formic, acetic and peracetic acids) grow with higher slopes when initial
400 concentration of toluene is lower. Meanwhile, the peak of SOA number concentration
401 appeared earlier and stronger along with lower initial concentration of toluene (**Fig. 10**).
402 These phenomena are in line with the findings that the first-generation products
403 compete with toluene on consumption of OH radical, leading to a faster formation of
404 intermediate VOCs, and promoting the formation of further generation products in gas-
405 and particle-phase (Atkinson and Arey, 2003; Chen et al., 2019; Ng et al., 2007).

406 In the photooxidation of toluene under dry conditions with no seed and NO_x , the
407 SOA yield varied from 24.7%–34.3%, comparable to those of Hinks et al. (2018),
408 Hildebrandt Ruiz et al. (2015), and Ng et al. (2007) (**Table 4**). The high SOA yield
409 under no NO_x conditions is likely due to no competition between NO_x and HO_2 to
410 reaction with RO_2 , which is indicated by other studies under low NO_x conditions with
411 much lower yield with a range of 1.9%–17% (Chen et al., 2019; Qi et al., 2020; Sato et
412 al., 2007). To be noted, the excess OH would cause ageing of SOA, leading to a higher
413 SOA yield (Hildebrandt Ruiz et al., 2015). With higher precursor concentrations,
414 intermediates are quickly oxidized into SOA, following the size increasing in ageing
415 processes (**Fig. 10**).

416

417 **3 Conclusions**

418 We built an indoor smog chamber system to study VOC photooxidation and SOA
419 formation progress close to the real atmospheric environment. State-of-the-art on-line
420 instruments for monitoring gas-phase and particle-phase oxidation products have been
421 equipped in our system. A series of characterization experiments have been conducted,
422 showing that our system is with good temperature and humidity stability, high mixing
423 efficiency, low contamination of background air, small wall loss of gases and particles,
424 and moderate light intensity compared with other chambers and ambient environment.
425 The yields of photooxidation of toluene under dry conditions in absence of NO_x showed
426 good agreement with previous studies under similar experimental conditions. OVOCs
427 and SOA formation depends on the initial concentrations of toluene and OH radical.
428 With higher concentrations of toluene, faster consumption of intermediates and
429 formation of later generation products and SOA were observed. The overdose H₂O₂
430 results in excess OH, which aggravated the ageing processes of toluene oxidation
431 products. These results illustrate that our chamber can be used for investigation of
432 transformation of gas-phase pollutants and SOA formation and ageing processes.

433

434 **Authorship Contribution Statement**

435 **Qi Yuan:** Data curation, Formal analysis, Investigation, Methodology, Writing-original
436 draft, Writing-review & editing

437 **Zhuozhi Zhang:** Formal analysis, Software, Writing-review & editing

438 **Meng Wang:** Methodology, Writing-review & editing

439 **Kin Fai Ho:** Funding acquisition, Methodology, Supervision

440 **Tao Wang:** Funding acquisition, Methodology

441 **Shun-cheng Lee:** Funding acquisition, Methodology, Supervision, Writing-review &
442 editing

443

444

445

446 **Acknowledgement**

447 This research was financially supported by the Research Grants Council (RGC) of
448 Hong Kong Special Administrative Region, China (T24-504/17-N). The authors would
449 like to acknowledge the HKPolyU University Research Facility in Chemical and
450 Environmental Analysis (UCEA) for the equipment support. The authors are grateful
451 to Dr. Yi Chen from Hong Kong University of Science and Technology for the
452 inspiration and discussion on the characterization experiments.

453

454 **Reference**

- 455 Arey, J., Obermeyer, G., Aschmann, S.M., Chattopadhyay, S., Cusick, R.D., and Atkinson, R., 2009.
 456 Dicarbonyl products of the OH radical-initiated reaction of a series of aromatic hydrocarbons, *Environ.*
 457 *Sci. Technol.*, 43, 683-689.
- 458 Atkinson, R., and Arey, J., 2003. Atmospheric Degradation of Volatile Organic Compounds, *Chem. Rev.*,
 459 103, 4605-4638.
- 460 Atkinson, R., Baulch, D.L., Cox, R.A., Crowley, J.N., Hampson, R.F., Hynes, R.G., Jenkin, M.E., Rossi,
 461 M.J., and Troe, J., 2004. Evaluated kinetic and photochemical data for atmospheric chemistry: Volume I
 462 - gas phase reactions of O_x, HO_x, NO_x and SO_x species, *Atmos. Chem. Phys.*, 4, 1461-1738.
- 463 Babar, Z.B., Park, J.-H., Kang, J., and Lim, H.-J., 2017. Characterization of a Smog Chamber for
 464 Studying Formation and Physicochemical Properties of Secondary Organic Aerosol, *Aerosol Air Qual.*
 465 *Res.*, 16, 3102-3113.
- 466 Bae, G., Kim, M., Lee, S., Song, K., Jin, H., and Moon, K., 2003. Design and performance evaluation of
 467 the KIST indoor smog chamber, *J. Korean Soc. Atmos. Environ.*, 19, 437-449.
- 468 Bloss, C., Wagner, V., Bonzanini, A., Jenkin, M.E., Wirtz, K., Martin-Reviejo, M., and Pilling, M.J.,
 469 2005a. Evaluation of detailed aromatic mechanisms (MCMv3 and MCMv3.1) against environmental
 470 chamber data, *Atmos. Chem. Phys.*, 5, 623-639.
- 471 Bloss, C., Wagner, V., Jenkin, M.E., Volkamer, R., Bloss, W.J., Lee, J.D., Heard, D.E., Wirtz, K., Martin-
 472 Reviejo, M., Rea, G., Wenger, J.C., and Pilling, M.J., 2005b. Development of a detailed chemical
 473 mechanism (MCMv3.1) for the atmospheric oxidation of aromatic hydrocarbons, *Atmos. Chem. Phys.*,
 474 5, 641-664.
- 475 Brasseur, G., & Jacob, D., 2017. Chemical Processes in the Atmosphere, in: *Modeling of Atmospheric*
 476 *Chemistry*, Cambridge University Press, Cambridge, 54-83.
- 477 Carter, W.P.L., Cocker, D.R., Fitz, D.R., Malkina, I.L., Bumiller, K., Sauer, C.G., Pisano, J.T., Bufalino,
 478 C., and Song, C., 2005. A new environmental chamber for evaluation of gas-phase chemical mechanisms
 479 and secondary aerosol formation, *Atmos. Environ.*, 39, 7768-7788.
- 480 Chen, T., Liu, Y., Chu, B., Liu, C., Liu, J., Ge, Y., Ma, Q., Ma, J., and He, H., 2019. Differences of the
 481 oxidation process and secondary organic aerosol formation at low and high precursor concentrations, *J.*
 482 *Environ. Sci. (China)*, 79, 256-263.
- 483 Chu, B., Chen, T., Liu, Y., Ma, Q., Mu, Y., Wang, Y., Ma, J., Zhang, P., Liu, J., Liu, C., Gui, H., Hu, R.,
 484 Hu, B., Wang, X., Wang, Y., Liu, J., Xie, P., Chen, J., Liu, Q., Jiang, J., Li, J., He, K., Liu, W., Jiang, G.,
 485 Hao, J., and He, H., 2021. Application of smog chambers in atmospheric process studies, *Natl. Sci. Rev.*,
 486 9.
- 487 Cocker, D.R., Flagan, R.C., and Seinfeld, J.H., 2001. State-of-the-Art Chamber Facility for Studying
 488 Atmospheric Aerosol Chemistry, *Environ. Sci. Technol.*, 35, 2594-2601.
- 489 Garmash, O., Rissanen, M.P., Pullinen, I., Schmitt, S., Kausiala, O., Tillmann, R., Zhao, D., Percival, C.,
 490 Bannan, T.J., Priestley, M., Hallquist, Å.M., Kleist, E., Kiendler-Scharr, A., Hallquist, M., Berndt, T.,
 491 McFiggans, G., Wildt, J., Mentel, T.F., and Ehn, M., 2020. Multi-generation OH oxidation as a source
 492 for highly oxygenated organic molecules from aromatics, *Atmos. Chem. Phys.*, 20, 515-537.
- 493 Gómez Alvarez, E., Viidanoja, J., Muñoz, A., Wirtz, K., and Hjorth, J., 2007. Experimental confirmation

494 of the dicarbonyl route in the photo-oxidation of toluene and benzene, *Environ. Sci. Technol.*, 41, 8362-
495 8369.

496 Grosjean, D., 1985. Wall loss of gaseous pollutants in outdoor Teflon chambers, *Environ. Sci. Technol.*,
497 19, 1059-1065.

498 Guo, S., Hu, M., Peng, J., Wu, Z., Zamora, M.L., Shang, D., Du, Z., Zheng, J., Fang, X., Tang, R., Wu,
499 Y., Zeng, L., Shuai, S., Zhang, W., Wang, Y., Ji, Y., Li, Y., Zhang, A.L., Wang, W., Zhang, F., Zhao, J.,
500 Gong, X., Wang, C., Molina, M.J., and Zhang, R., 2020. Remarkable nucleation and growth of ultrafine
501 particles from vehicular exhaust, 117, 3427-3432.

502 Hallquist, M., Wenger, J.C., Baltensperger, U., Rudich, Y., Simpson, D., Claeys, M., Dommen, J.,
503 Donahue, N.M., George, C., Goldstein, A.H., Hamilton, J.F., Herrmann, H., Hoffmann, T., Iinuma, Y.,
504 Jang, M., Jenkin, M.E., Jimenez, J.L., Kiendler-Scharr, A., Maenhaut, W., McFiggans, G., Mentel, T.F.,
505 Monod, A., Prévôt, A.S.H., Seinfeld, J.H., Surratt, J.D., Szmigielski, R., and Wildt, J., 2009. The
506 formation, properties and impact of secondary organic aerosol: current and emerging issues, *Atmos.*
507 *Chem. Phys.*, 9, 5155-5236.

508 Harrison, R.M., and Yin, J., 2000. Particulate matter in the atmosphere: which particle properties are
509 important for its effects on health?, *Sci. Total. Environ.*, 249, 85-101.

510 Hildebrandt Ruiz, L., Paciga, A.L., Cerully, K.M., Nenes, A., Donahue, N.M., and Pandis, S.N., 2015.
511 Formation and aging of secondary organic aerosol from toluene: changes in chemical composition,
512 volatility, and hygroscopicity, *Atmos. Chem. Phys.*, 15, 8301-8313.

513 Hinks, M.L., Montoya-Aguilera, J., Ellison, L., Lin, P., Laskin, A., Laskin, J., Shiraiwa, M., Dabdub, D.,
514 and Nizkorodov, S.A., 2018. Effect of relative humidity on the composition of secondary organic aerosol
515 from the oxidation of toluene, *Atmos. Chem. Phys.*, 18, 1643-1652.

516 Hu, C., Cheng, Y., Pan, G., Gai, Y., Gu, X., Zhao, W., Wang, Z., Zhang, W., Chen, J., Liu, F., Shan, X.,
517 and Sheng, L., 2014. A Smog Chamber Facility for Qualitative and Quantitative Study on Atmospheric
518 Chemistry and Secondary Organic Aerosol, *Chinese J. Chem. Phys.*, 27, 631-639.

519 Karl, M., Brauers, T., Dorn, H.-P., Holland, F., Komenda, M., Poppe, D., Rohrer, F., Rupp, L., Schaub,
520 A., and Wahner, A., 2004. Kinetic Study of the OH-isoprene and O₃-isoprene reaction in the atmosphere
521 simulation chamber, SAPHIR, *Geophys. Res. Lett.*, 31.

522 Li, J., Li, H., Wang, X., Wang, W., Ge, M., Zhang, H., Zhang, X., Li, K., Chen, Y., Wu, Z., Chai, F.,
523 Meng, F., Mu, Y., Mellouki, A., Bi, F., Zhang, Y., Wu, L., and Liu, Y., 2021a. A large-scale outdoor
524 atmospheric simulation smog chamber for studying atmospheric photochemical processes:
525 Characterization and preliminary application, *J. Environ. Sci. (China)*, 102, 185-197.

526 Li, K., Lin, C., Geng, C., White, S., Chen, L., Bao, Z., Zhang, X., Zhao, Y., Han, L., Yang, W., and Azzi,
527 M., 2020. Characterization of a new smog chamber for evaluating SAPRC gas-phase chemical
528 mechanism, *J. Environ. Sci. (China)*, 95, 14-22.

529 Li, Y., Zhao, J., Wang, Y., Seinfeld, J.H., and Zhang, R., 2021b. Multigeneration Production of Secondary
530 Organic Aerosol from Toluene Photooxidation, *Environ. Sci. Technol.*, 55, 8592-8603.

531 Li, Z., Xue, L., Yang, X., Zha, Q., Tham, Y.J., Yan, C., Louie, P.K.K., Luk, C.W.Y., Wang, T., and Wang,
532 W., 2018. Oxidizing capacity of the rural atmosphere in Hong Kong, Southern China, *Sci. Total. Environ.*,
533 612, 1114-1122.

534 Ma, W., Liu, Y., Zhang, Y., Feng, Z., Zhan, J., Hua, C., Ma, L., Guo, Y., Zhang, Y., Zhou, W., Yan, C.,
535 Chu, B., Chen, T., Ma, Q., Liu, C., Kulmala, M., Mu, Y., and He, H., 2022. A New Type of Quartz Smog

536 Chamber: Design and Characterization, Environ. Sci. Technol., 56, 2181-2190.
537 McMurry, P.H., and Grosjean, D., 1985. Gas and aerosol wall losses in Teflon film smog chambers,
538 Environ. Sci. Technol., 19, 1176-1182.
539 Nakao, S., Clark, C., Tang, P., Sato, K., and Cocker Iii, D., 2011. Secondary organic aerosol formation
540 from phenolic compounds in the absence of NO_x, Atmos. Chem. Phys., 11, 10649-10660.
541 Ng, N.L., Kroll, J.H., Chan, A.W.H., Chhabra, P.S., Flagan, R.C., and Seinfeld, J.H., 2007. Secondary
542 organic aerosol formation from *m*-xylene, toluene, and benzene, Atmos. Chem. Phys., 7, 3909-3922.
543 Paulsen, D., Dommen, J., Kalberer, M., Prévôt, A.S.H., Richter, R., Sax, M., Steinbacher, M.,
544 Weingartner, E., and Baltensperger, U., 2005. Secondary Organic Aerosol Formation by Irradiation of
545 1,3,5-Trimethylbenzene–NO_x–H₂O in a New Reaction Chamber for Atmospheric Chemistry and Physics,
546 Environ. Sci. Technol., 39, 2668-2678.
547 Pye, H.O.T., D'Ambro, E.L., Lee, B.H., Schobesberger, S., Takeuchi, M., Zhao, Y., Lopez-Hilfiker, F.,
548 Liu, J., Shilling, J.E., Xing, J., Mathur, R., Middlebrook, A.M., Liao, J., Welti, A., Graus, M., Warneke,
549 C., de Gouw, J.A., Holloway, J.S., Ryerson, T.B., Pollack, I.B., and Thornton, J.A., 2019. Anthropogenic
550 enhancements to production of highly oxygenated molecules from autoxidation, Proceedings of the
551 National Academy of Sciences of the United States of America, 116, 6641-6646.
552 Qi, X., Zhu, S., Zhu, C., Hu, J., Lou, S., Xu, L., Dong, J., and Cheng, P., 2020. Smog chamber study of
553 the effects of NO_x and NH₃ on the formation of secondary organic aerosols and optical properties from
554 photo-oxidation of toluene, Sci. Total. Environ., 727, 138632.
555 Rohrer, F., Bohn, B., Brauers, T., Brüning, D., Johnen, F.J., Wahner, A., and Kleffmann, J., 2005.
556 Characterisation of the photolytic HONO-source in the atmosphere simulation chamber SAPHIR, Atmos.
557 Chem. Phys., 5, 2189-2201.
558 Saathoff, H., Moehler, O., Schurath, U., Kamm, S., Dippel, B., and Mihelcic, D., 2003. The AIDA soot
559 aerosol characterisation campaign 1999, J. Aerosol Sci., 34, 1277-1296.
560 Sato, K., Hatakeyama, S., and Imamura, T., 2007. Secondary Organic Aerosol Formation during the
561 Photooxidation of Toluene: NO_x Dependence of Chemical Composition, J. Phys. Chem. A, 111, 9796-
562 9808.
563 Shao, Y., Wang, Y., Du, M., Voliotis, A., Alfarra, M.R., O'Meara, S.P., Turner, S.F., and McFiggans, G.,
564 2022. Characterisation of the Manchester Aerosol Chamber facility, Atmos. Meas. Tech., 15, 539-559.
565 Smith, D.M., Fiddler, M.N., Sexton, K.G., and Bililign, S., 2019. Construction and Characterization of
566 an Indoor Smog Chamber for Measuring the Optical and Physicochemical Properties of Aging Biomass
567 Burning Aerosols, Aerosol Air Qual. Res., 19, 467-483.
568 Song, M., Liu, Ying, Li, Xin, Lu, Sihua, 2021. Advances on Atmospheric Oxidation Mechanism of
569 Typical Aromatic Hydrocarbons, Acta Chim. Sin., 79, 1214-1231.
570 Srivastava, D., Vu, T.V., Tong, S., Shi, Z., and Harrison, R.M., 2022. Formation of secondary organic
571 aerosols from anthropogenic precursors in laboratory studies, npj Climate and Atmospheric Science, 5,
572 22.
573 Tan, Y., Liu, C., Ho, K., Ma, Q., and Lee, S.-C., 2021. Characterization of an indoor environmental
574 chamber and identification of C1–C4 OVOCs during isoprene ozonolysis, Indoor. Built Environ., 30,
575 554-564.
576 Wagner, R., Möhler, O., Saathoff, H., Schnaiter, M., and Leisner, T., 2011. New cloud chamber
577 experiments on the heterogeneous ice nucleation ability of oxalic acid in the immersion mode, Atmos.

578 Chem. Phys., 11, 2083-2110.
579 Wang, J., Doussin, J.F., Perrier, S., Perraudin, E., Katrib, Y., Pangui, E., and Picquet-Varrault, B., 2011.
580 Design of a new multi-phase experimental simulation chamber for atmospheric photosmog, aerosol and
581 cloud chemistry research, Atmos. Meas. Tech., 4, 2465-2494.
582 Wang, W.-G., Li, K., Zhou, L., Ge, M.-F., Hou, S.-Q., Tong, S.-R., Mu, Y.-J., and Jia, L., 2015. Evaluation
583 and Application of Dual-Reactor Chamber for Studying Atmospheric Oxidation Processes and
584 Mechanisms, Acta Phys. -Chim. Sin., 31, 1251-1259.
585 Wang, X., Liu, T., Bernard, F., Ding, X., Wen, S., Zhang, Y., Zhang, Z., He, Q., Lü, S., Chen, J., Saunders,
586 S., and Yu, J., 2014. Design and characterization of a smog chamber for studying gas-phase chemical
587 mechanisms and aerosol formation, Atmos. Meas. Tech., 7, 301-313.
588 White, S., Angove, D., Li, K., Campbell, I., Element, A., Halliburton, B., Lavrencic, S., Cameron, D.,
589 Jamie, I., and Azzi, M., 2018. Development of a new smog chamber for studying the impact of different
590 UV lamps on SAPRC chemical mechanism predictions and aerosol formation, Environ. Chem., 15, 171-
591 182.
592 Wu, S., Lü, Z., Hao, J., Zhao, Z., Li, J., Takekawa, H., Minoura, H., and Yasuda, A., 2007. Construction
593 and characterization of an atmospheric simulation smog chamber, Adv. Atmos. Sci., 24, 250-258.
594

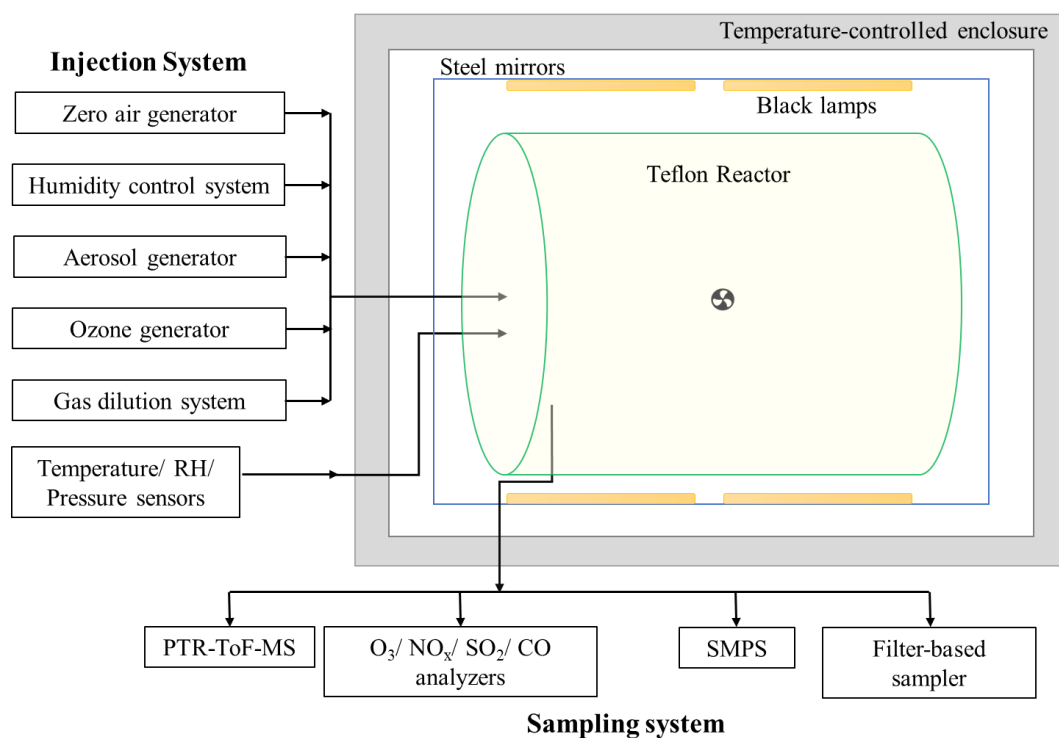


Figure 1. The schematics of the PolyU smog chamber facility.

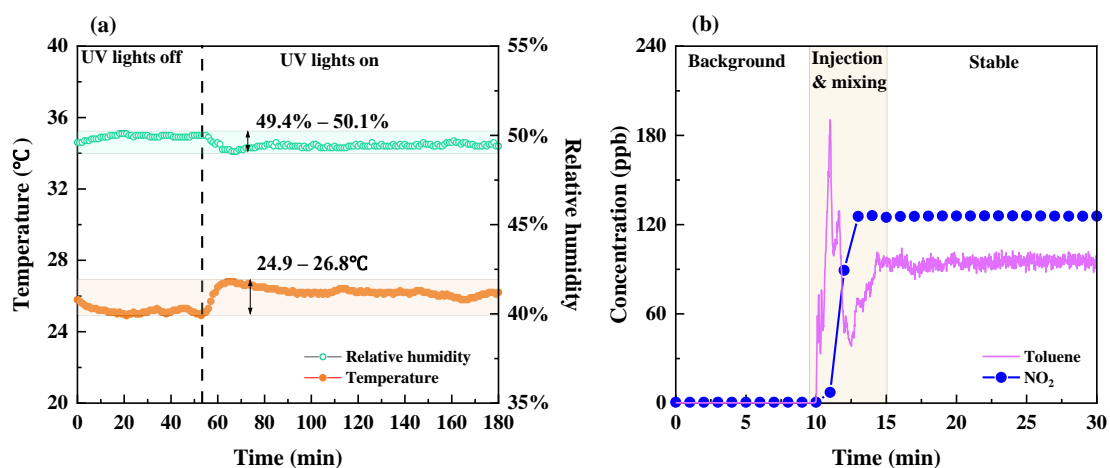


Figure 2. Evolution of (a) temperature and relative humidity inside the chamber with lights off and 42 lights on, (b) toluene and NO₂ before, during and after the injection.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60
61
62
63
64
65

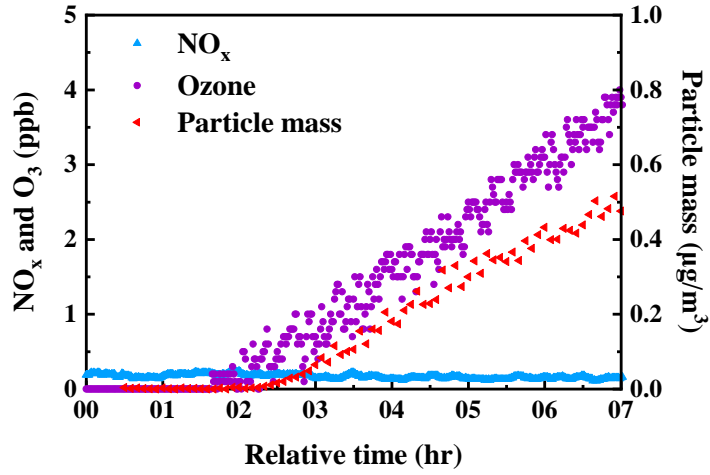


Figure 3. Concentrations of NO_x , Ozone and Particle number as a function of time during the photolysis of zero air.

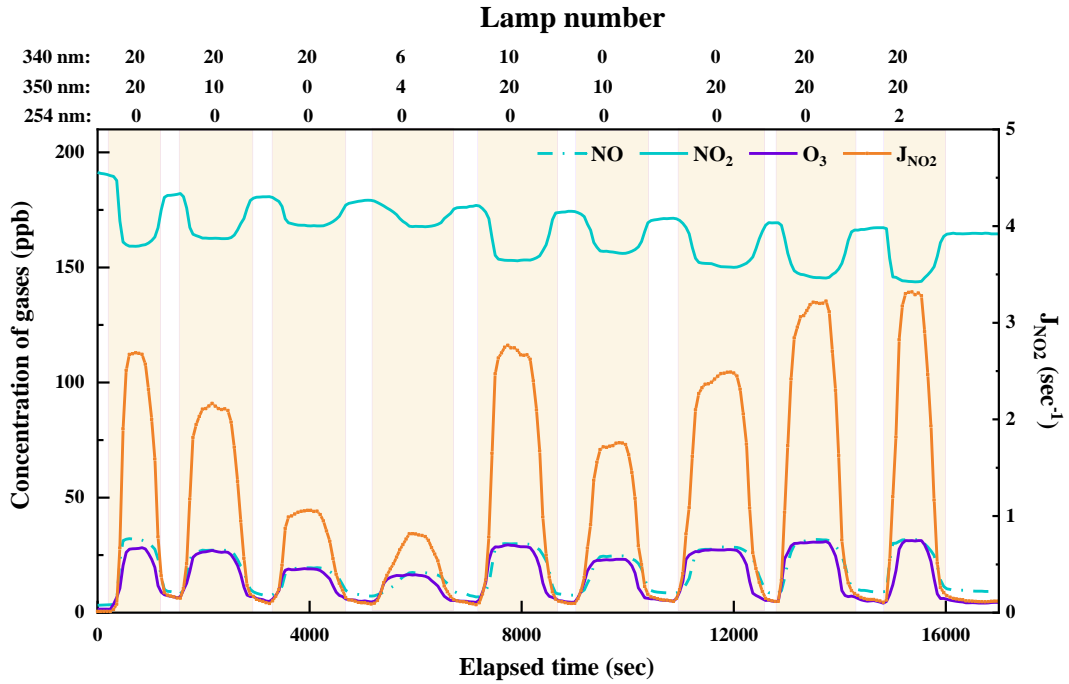


Figure 4. Variations of $j(\text{NO}_2)$ against concentrations of NO , NO_2 and O_3 . The filled purple area means when a certain number of lights are turned on.

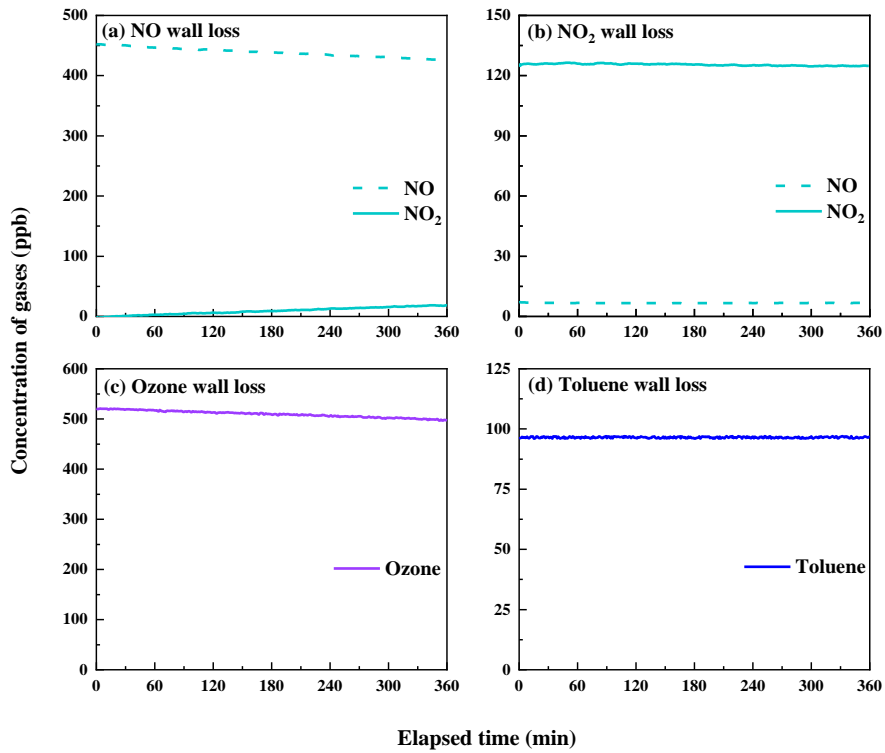


Figure 5. Experimental results for (a) NO, (b) NO₂, (c) O₃, (d) toluene wall loss.

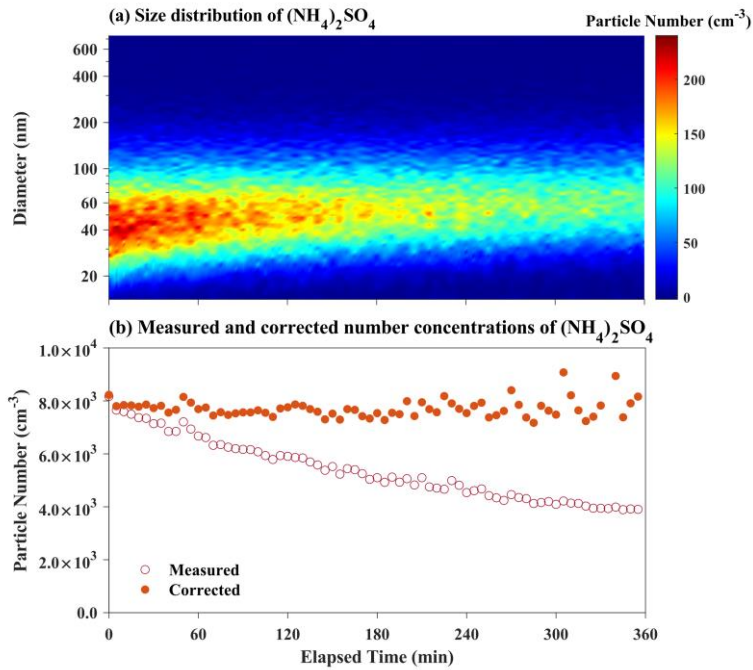


Figure 6. (a) Size distribution of $(\text{NH}_4)_2\text{SO}_4$ during a six-hours particle wall-loss experiment, (b) corrected particle number concentrations compared to measured particle number concentrations.

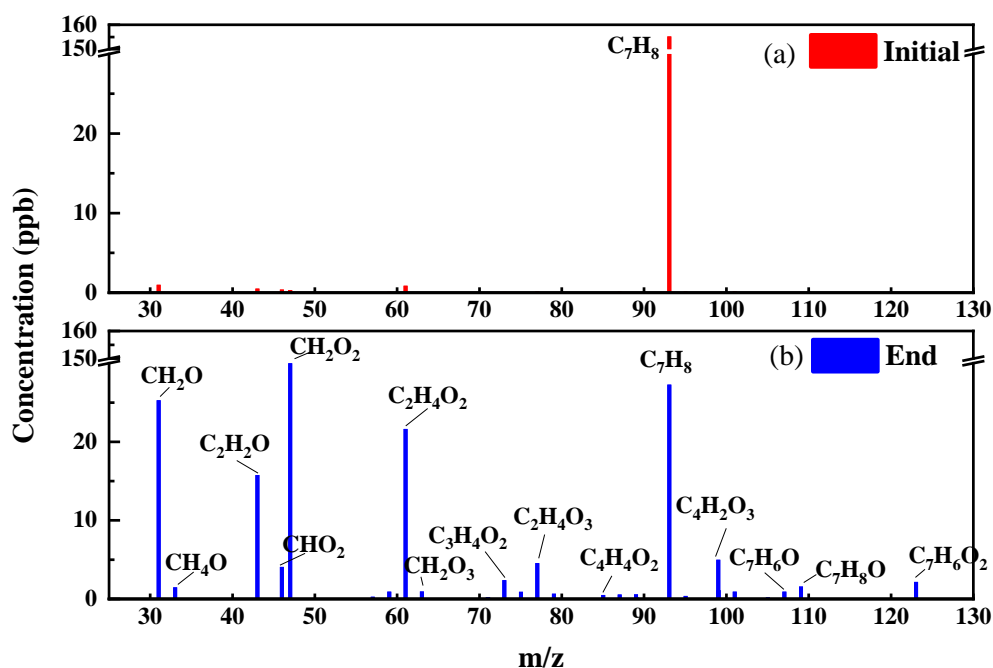


Figure 7. Mass spectra of VOCs detected by PTR-TOF-MS at the start (a) and the end (b) of toluene photooxidation experiment.

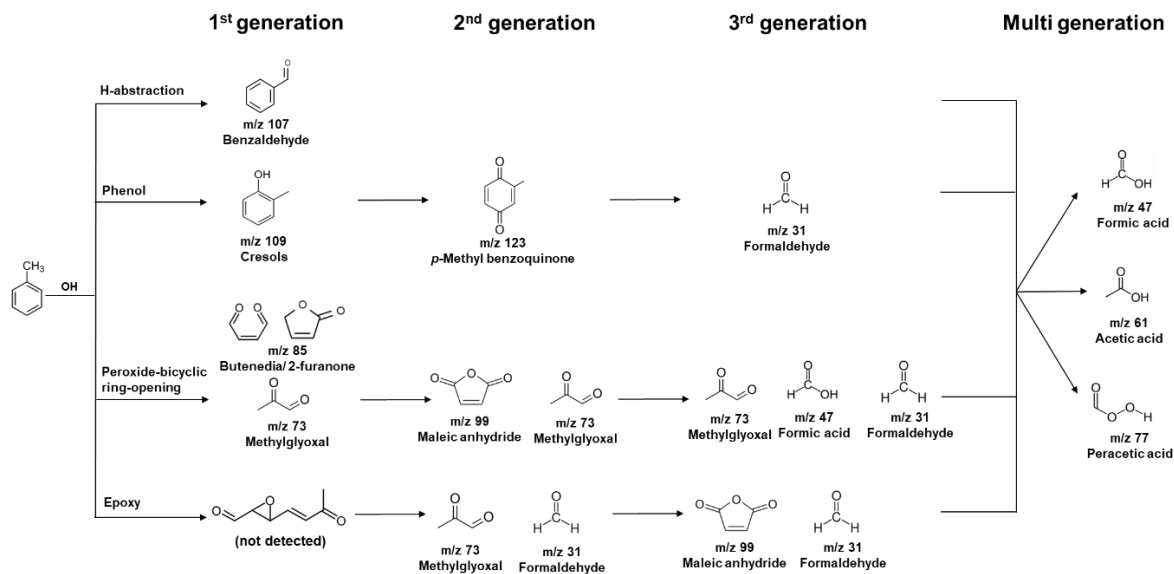


Figure 8. Major gas-phase products detected by PTR-TOF-MS in the toluene photooxidation system as implemented in the MCM v3.3.1.

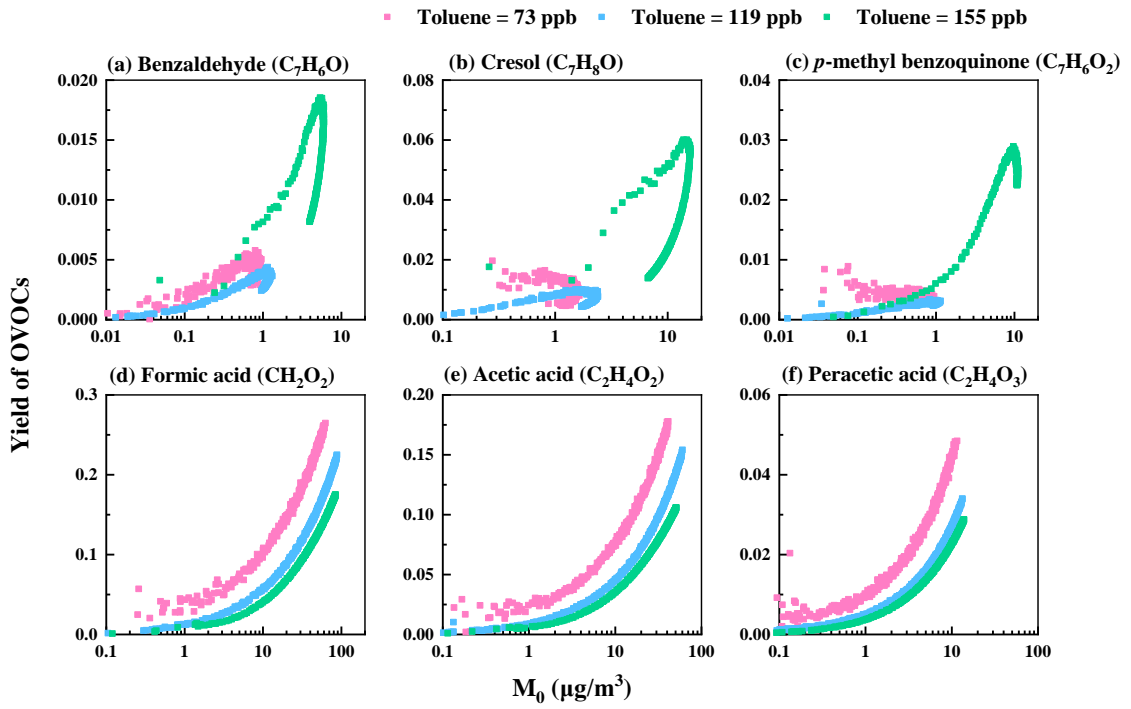


Figure 9. Yield of OVOCs in toluene-OH test.

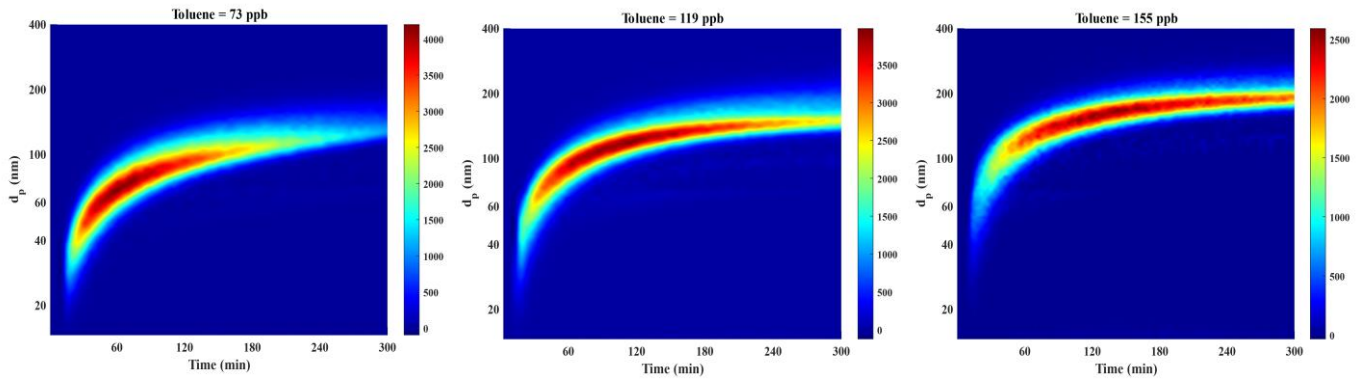


Figure 10. Variation of SOA size and number concentration for toluene photooxidation under different initial precursor concentration conditions.

Table 1. Overview of instruments.

Instrument	Target species and parameters	Detection limit/ Range	Accuracy	Flow rate (L/min)
Teledyne Model T200U	NO/NO ₂ /NO _x	< 50 ppt	0.50%	1
Teledyne Model T400	O ₃	0.4 ppb	0.5 ppb or 0.5%	0.8
Ionicon PTR-QiToF-MS	VOCs	< 10 ppt	± 5%	0.05–0.8
Aerodyne HR-ToF-CIMS	Oxygenated VOCs	< 0.01 ppt		1.2–10
Aerodyne Filter Inlet for Gases and AEROSols, FIGAERO	Particle composition	>10 ng		10
SMPS (TSI 3082 DMA and 3776 CPC)	Particle number	1–10 ⁷ cm ⁻³	± 3%	0.3
	Particle size	10–1000 nm	± 3%	
Custom built filter collector	Particle collection for offline analysis			20
Thermo Scientific LC- Orbitrap MS	Oxidation products in offline particulate samples		<1 ppm	NA
Beijing Convenient Environmental Tech Co.Ltd sensors	Temperature	10–40°C	± 1°C	NA
	Relative humidity	5%–85%	± 3%	NA
	Pressure	0-1000 Pa	± 3 Pa	NA

Table 2. Summary of wall loss rate of gas species and particle in PolyU chamber and comparison with other chamber facilities.

Chamber	Volume (m ³)	Surface to Volume Ratio (m ⁻¹)	NO (× 10 ⁻⁴ min ⁻¹)	NO ₂ (× 10 ⁻⁴ min ⁻¹)	O ₃ (× 10 ⁻⁴ min ⁻¹)	VOC (× 10 ⁻⁴ min ⁻¹)	k _{dep} at 100 nm (× 10 ⁻³ min ⁻¹)	J _{NO2} (× 10 ⁻³ sec ⁻¹)	Reference
PolyU	5.4	1.59	0.47	0.37	1.17	negligible (toluene)	0.2	3.32	This study
Tsinghua University	2	5	3.38	4.17	6.07	0.06	2.6	3.83	Wu et al. (2007)
KIST	2.5		7-20	4-20	12–24			2	Bae et al. (2003)
Zhejiang University	3	4.32		1.64	8.99		3.52	2	Li et al. (2020)
CESAM, CNRS	4.2	4.29			20	< 0.6 (propene & isoprene)	0.02–0.1	4.5	Wang et al. (2011)
ICCAS-DRC	5	3.59	3.0–3.1	3.8–4.5	2.5–3.1	0.31 (propene)	2.0–4.3	2.32–2.55	Wang et al. (2015)
Kyungpook National University	7	3.09	3.78	4.48	6.47	< 1.71	3.96	2.83	Babar et al. (2017)
NCAT	9	2.9	7.40 ± 0.01	3.47 ± 0.01	5.91 ± 0.08		2.46 ± 0.11	2.75 ± 0.08	Smith et al. (2019)
BUCT	10	2.6	4.50–4.58	2.16–3.54	6.76–8.90	0.11-0.46 (toluene)	1.33	6.67 ± 0.17	Ma et al. (2022)

Table 3. Summary of initial conditions and results for toluene photooxidation experiments.

	Toluene injection (ppb)	ΔH_C ($\mu\text{g}/\text{m}^3$)	M₀ of SOA ($\mu\text{g}/\text{m}^3$)	Yield of SOA	M₀ of OVOCs ($\mu\text{g}/\text{m}^3$)	Yield of OVOCs
Test 01	73	198.79	29.0	24.7	103.49	0.52
Test 02	119	336.11	71.0	26.4	197.92	0.59
Test 03	155	465.14	165.3	34.3	287.12	0.62

Table 4. Summary of SOA yields in previous studies on toluene photooxidation at low or no NO_x conditions.

[Toluene] ₀ (ppb)	Δ[Toluene] (ppb)	PM (μg /m ³)	SOA yield	Reference
73	32	29.0	24.7%	This study
119	72	71.0	26.4%	This study
155	128	165.3	34.3%	This study
1000	–	210 ± 20	15% ± 2%	
1000	–	87 ± 6	6.2% ± 0.5%	
1000	–	84	5.9%	
1000	–	28 ± 7	2.0% ± 0.5%	Hinks et al. (2018)
1000	–	26 ± 9	1.9% ± 0.6%	
300	–	27	5.5%	
300	–	9	2.2%	
3980	–	127	7.58%	
4004	–	97	5.61%	
3998	–	92	5.87%	Sato et al. (2007)
4074	–	109	4.55%	
4005	–	104	6.31%	
380	–	20	41%	
270	–	20	40%	
180	–	20	66%	Hildebrandt Ruiz et al. (2015)
200	–	20	31%	
570	–	20	26%	
570	–	20	26%	
690	306	93.9	8.15%	
603	270	62.2	6.12%	
425	245	42.6	4.61%	
275	172	34.2	5.26%	Chen et al. (2019)
66	39	12.6	8.44%	
59	44	13.8	8.23%	
37	25	7.2	7.76%	
960	318	0	0%	
990	511	87.1	4.3%	Qi et al. (2020)
1010	553.7	168.9	8.4%	
1010	596.2	291.5	12.1%	
–	32.1	37.4	30.8% ± 1.7%	
–	63.9	73.1 ± 5.6	30.2% ± 0.7%	Ng et al. (2007)
–	10	11.5 ± 1.6	30.4% ± 4.1%	
–	23.8	26.7 ± 2.5	29.8% ± 1.6%	



Click here to access/download
Supplementary Material
Supporting Information.docx

