Comparative Life Cycle Assessment of Composite Structures Incorporating Uncertainty and Global Sensitivity Analysis

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4 Abstract: In recent years, there have been increasing efforts seeking novel material and structural 5 alternatives to alleviate environmental and economic burdens caused by conventional engineering 6 structures. However, research on long-term environmental impact and cost of the design 7 alternatives are limited. This paper presents a comparative life-cycle assessment (LCA) and life-8 cycle cost analysis (LCCA) of three composite columns over a service life of 100 years. The 9 studied cases are typical composite structural forms consisting of: (1) concrete-filled steel tubular 10 column (CFST); (2) concrete-filled fiber-reinforced polymer (FRP) tubular column (CFFT); and (3) hybrid FRP-concrete-steel double-skin tubular column (DSTC). The CFFT is expected to have 11 12 extended service life due to corrosion resistance of FRP. The DSTC is designed to reduce concrete 13 consumption by leaving a void at the center of the column. Both deterministic and probabilistic 14 results are discussed in this research. Specifically, Sobol's index is selected to aid the probabilistic LCA and LCCA analyses. The deterministic LCA results indicate that CFFT has the least CO₂ 15 16 emission: 50 % less than DSTC and 60 % less than CFST. While LCCA results show that for the investigated scenario the DSTC costs the most across the studied service life, about 15 % more 17 18 than CFST and CFFT. The probabilistic results indicated that the production and maintenance 19 stage are two significant influential factors of the LCA and LCCA results. In general, CFFT and 20 DSTC are more economic and environmental-friendly alternatives compared to CFST.

Keywords: Fiber-reinforced polymer (FRP), concrete, steel, CO₂ emissions, life-cycle cost,
 uncertainty analysis.

23 **1. Introduction**

24 As concerns about environmental issues are raising constantly, construction environmental impact, 25 one of the major sources of greenhouse gas emissions, has drawn more public attention. 26 Infrastructure construction has been assessed to be responsible for one third of the global 27 anthropogenic greenhouse gas (GHG) emissions and approximately the same share of the global 28 energy production [1]. As two widely used construction materials, concrete and steel contribute 29 massive portion of global CO₂ emission. In 2012 alone, the production of cement was 30 approximately 3.8 giga tonnes (Gt), approximately equal to 3.2 Gt of CO_2 emissions worldwide, 31 around 8 % of the annual anthropogenic CO₂ emissions [2]. The carbon intensity of steel production varies among manufacturing techniques, data has shown that the average CO₂ emission 32 33 is 1.9 t CO_2 per tonne of steel production [3].

To mitigate these environmental impacts triggered by concrete and steel production, industries have attempted to substitute conventional materials with alternatives, such as recycled concrete [4-7], ultra-high performance concrete (UHPC) [8, 9], stainless steel [10], etc., which have been proved to show better performance from a life-cycle perspective [11-13]. Apart from

38 innovation at the material level, another concept of integrating multiple novel materials into 39 structural elements (e.g., concrete-filled steel tubular and concrete-fill FRP tubular columns) has 40 gained momentum. Concrete-filled steel tubular columns (CFSTs) offer a variety of structural benefits including high strength, superior fire resistance, outstanding ductility, and great energy 41 42 absorption capacity [14]. In addition, with the outer steel tube acting as shuttering, the construction 43 cost and duration could be reduced accordingly. Nevertheless, corrosion issue must be considered 44 under marine circumstances, where FRP-concrete composite structures may become a better 45 option for its corrosion-resistance characteristic. Likewise, the FRP confinements can also be used 46 as permanent shuttering of concrete. The applied FRP jackets can be fabricated by pultrusion, wet 47 lay-up or filament winding techniques, and are light-weighted enough to be easily shipped from 48 plants to construction sites [15]. Wet lay-up (e.g., FRP wraps) is more commonly used in retrofitting existing structures while pultrusion (e.g., FRP bars) and filament winding (e.g., 49 50 prefabricated FRP tubes) are more suitable for constructing new constructions. Filament-wound FRP tubes are widely used in fabricating columns with plain or steel-reinforced concrete infilled, 51 which are known as concrete-filled FRP tubular columns (CFFTs). CFFTs have been widely 52 53 applied in new constructions such as bridge columns and piles due to the outstanding corrosion 54 resistance [16-19]. Nevertheless, the major drawbacks to CFFTs are high initial costs, linear-55 elastic-brittle stress-strain behavior, low elastic modulus-to-strength ratio and inferior fire resistance [20]. To offset these disadvantages, an innovative structural form was proposed in [20], 56 57 namely hybrid FRP-concrete-steel double-skin tubular columns (DSTCs). A DSTC consists of an 58 inner steel tube and an outer FRP tube with concrete filled in between, which is expected to be 59 more ductile and thus can be used in flexural and seismic conditions. It particularly benefits the 60 need for sustainable construction as the presence of the inner void largely reduces concrete without significant strength loss. Furthermore, DSTC holds great potential for associating with high-61 strength concrete [21, 22] and even UHPC [23-25]. 62

63 While CFSTs, CFFTs, and DSTCs have been extensively studied regarding their mechanical 64 properties and structural performance [14, 26-31], limited efforts have been devoted to 65 investigating life-cycle performance of such hybrid structures and more studies should be 66 conducted on relevant research area. Steel, steel-concrete composite column with steel reinforcement and wooden column were investigated from a life-cycle perspective while the scope 67 only covered the cradle-to-gate span without considering the end-of-life scenario [32]. Han [33, 68 34] did some studies on CFST regarding its structural deterioration throughout the lifespan, the 69 70 results, however, were neither economy nor environment-oriented. A nonmonetary evaluation 71 model was developed to identify the life-cycle benefit-cost of CFST and FRP-confined concrete 72 structures [35], which was further refined to an advanced framework that utilizes material 73 properties to assess the performance-based life-cycle cost of composite materials in construction 74 [36]. Other studies were carried out to compute life-cycle cost or environmental impacts of FRP 75 reinforced concrete, FRP components or structures [37-43].

Life-cycle assessment (LCA) and life-cycle cost analysis (LCCA) are important techniques in the abovementioned studies, which have been widely used in buildings [44-46] and infrastructures [47-49]. LCA is a framework developed for evaluating life-cycle environmental 79 performance of a product system, including raw material acquisition, production, construction, use, 80 disposal, and transportation required during the process [50]. While LCCA is a supplementary 81 method to LCA, which employs life-cycle principles to account for all costs incurred over a system's lifetime rather than just initial cost [51]. The analysis has to be completed by aggregating 82 83 all life-cycle inventories (LCI) associated with every unit process. Dozens of LCI models have been developed, which can be classified into evaluation at midpoint (e.g. CML 2002, EDIP 97-84 85 2003, MEEuP and TRACI), endpoint (e.g. Eco-indicator 99 and EPS 2000), or combined (e.g. Impact 2002, Swiss Ecoscarcity 06, ReCiPe 2008) [52]. Midpoint impact categories include global 86 warming potential (GWP), primary energy (PE), acidification potential, eutrophication potential, 87 88 human toxicity potential, ozone layer depletion potential, and photochemical smog formation 89 potential. While endpoint focuses more on receptors including resources (R), ecosystem quality 90 (EQ), climate change (CC) and human health (HH) [53]. GWP, one of the most concerned 91 midpoint categories [53], is selected in this study. In the context of infrastructure, life-cycle cost 92 comprises agency cost, user cost, and environmental cost, except that environmental cost is rarely considered [54]. As this research focuses on LCCA of structural component rather than system, 93 94 only agency cost is evaluated herein.

95 Most of the previous research regarding composite structures are presented in a deterministic manner without considering uncertainties embedded in the assessment process [55]. The outcome 96 97 of the assessment model is affected by uncertainty (e.g. physical properties of materials, amount of pollutants emitted, etc.). Uncertainty plays a significant role within the life-cycle assessment 98 99 process and the effect of various uncertainty on the life-cycle performance could be very different. 100 It is of vital importance to take the uncertainties with the assessment process in a life-cycle context [56-59]. Without considering the uncertainty may introduce some arbitrariness in the comparison 101 102 of different design options. A common practice to study the effect of uncertainties on output is performing sensitivity analyses. Most of the sensitivity analyses focused on the linearity by 103 104 varying one parameter at a time, as such, the effect of the varied parameters can be reflected 105 individually [41]. Another approach requires input as probabilistic parameters rather than centered 106 values, by which the extent of each input parameter contributing to the output variance can be 107 determined [42]. These two methods belong to the realm of local sensitivity analysis and global 108 sensitivity analysis, respectively [55]. The global sensitivity analysis varies all the variables 109 simultaneously and the random variable changes through its entire range. In this paper, Sobol's 110 method is selected to perform global sensitivity analysis for its capability of determining the 111 critical input parameter, i.e., parameter that contributes the most to the output variance and 112 therefore should be most accurately known. Parameters with low Sobol indices can be removed, 113 as their contribution to the overall variance is insignificant and thus have marginal effect on the 114 output. Furthermore, Sobol's method is capable of identifying interaction effect between multiple 115 input parameters [60]. It has been proved by many researchers that Sobol's method performed well 116 on uncertainty treatment in LCA [55, 60-63]. To the best knowledge of the authors, there has been 117 no studies focusing on the probabilistic performance assessment and global sensitivity analysis of 118 CFST, CFFT, and DSTC in a life cycle context.

To address the current research gap, this paper aims to investigate deterministic and 119 probabilistic life-cycle performance of CFST, CFFT, and DSTC from environmental and 120 121 economic perspectives. Given three comparable composite structural cases, life-cycle assessment 122 and life-cycle cost analysis (LCA-LCCA), an approach that leads to long-term and preventive 123 assessment [54], is used in this study. The remainder of the paper is structured as follows: Section 124 2 elaborates the investigated structural forms including material and structural design, followed by 125 a detailed introduction of the life-cycle model in Section 3. Section 4 presents the computational 126 process of the deterministic life-cycle environmental impact and cost results while the accordingly 127 probabilistic results are summarized in Section 5.

128 **2. Investigated structural elements**

129 To demonstrate benefits of the novel structural systems (i.e. CFFT and DSTC) over traditional 130 CFST, fair comparison should be made among three. Herein, the case of traditional column CFST 131 is directly taken from [32], in which the strength, stability and stiffness were all taken into account 132 on the basis of Eurocode 3 [64] and Eurocode 4 [65]. [32] calculated the axial bearing capacity as 133 follows: 1. Consider a slab in a three-story building with panel dimension of 10×20 m²; 2. Take appropriate dead and live loads for the slab in accordance with the standards; and 3. Transform 134 135 dead and live loads into uniaxial load on a column of 5,622 kN. Therefore, CFFT and DSTC are 136 both designed to carry nominal axial load of 5,622 kN. Fig. 1 illustrates the loading condition and 137 cross-section profiles of the investigated columns. CFST consists of an outer steel tube and steel 138 reinforced concrete. CFFT comprises with a tubular FRP tube and a concrete core while DSTC 139 includes an FRP tube, an inner steel tube and an annular concrete infill.



140 2.1 Material design

- 141 Materials used for the three design alternatives are tabulated in Table 1. The benchmark example
- 142 is an excerpt from [32], where 1 m^3 of concrete production considers 400 kg of cement, 700 kg of
- sand, and 1200 kg of gravel. While for the other two scenarios, 1 m^3 of C40 concrete requires 350
- 144 kg of cement, 175 kg of water, 1194 kg of sand, and 614 kg of gravel [66]. A typical density of 145 7850 kg/m³ is assigned to steel. Filament-wound glass FRP is assumed to act as confinement for
- 146 CFFT and DSTC. Material properties of GFRP and steel tube (for CFFT and DSTC) are taken
- from the experimental results in [22].

148 **2.2 Structural design**

- 149 To make a fair comparison, CFFT and DSTC are supposed to be designed to have similar loading
- 150 capacity with CFST. To this end, the theoretical model proposed in [67] is used to determine
- 151 geometric parameters of CFFT and DSTC. The model in [67] has been verified against extensive
- 152 database [68-70]. This design method is described as follows:
- 153 Firstly, the confinement stiffness ratio ρ_{κ} and strain ratio ρ_{ε} are calculated as follows

$$\rho_{\kappa} = \frac{2E_{frp}t_{frp}}{(f_{co}/\varepsilon_{co})D}$$

$$\rho_{\varepsilon} = \frac{\varepsilon_{h,rup}}{\varepsilon_{co}}$$
(1)
(2)

where E_{frp} = elastic modulus of the FRP in the hoop direction, t_{frp} = thickness of the FRP tube, *D* = diameter of the confined concrete core, f'_{co} = unconfined concrete axial strength, ε_{co} = unconfined

- 156 concrete axial strain, and $\varepsilon_{h,rup}$ = hoop rupture strain of FRP.
- 157 With the above two ratios determined, the ultimate axial strain ε_{cu} and the compressive 158 strength f'_{cc} can be computed as

$$\frac{\varepsilon_{cu}}{\varepsilon_{co}} = 1.75 + 6.5\rho_{\kappa}^{0.8}\rho_{\varepsilon}^{1.45}$$
(3)

Given
$$\rho_{\kappa} \ge 0.01, \frac{f_{cc}}{f_{co}} = 1 + 3.5 (\rho_{\kappa} - 0.01) \rho_{\varepsilon}$$
 (4)

159 It should be noted that for DSTC [71], the ultimate axial strain ε_{cu} becomes

$$\frac{\varepsilon_{cu}}{\varepsilon_{co}} = 1.75 + 6.5 \rho_{\kappa}^{0.8} \rho_{\varepsilon}^{1.45} (1-\phi)^{-0.22}$$
(5)

- 160 where ϕ is void ratio, i.e., the ratio of outer diameter of the steel tube to diameter of the concrete 161 core.
- 162 The abovementioned models [67, 71] have been proved to be able to accurately predict axial 163 compressive strength of CFFT and DSTC via numerous experimental [72-75] and parametric 164 studies [70, 76-78].

The loading capacity of CFFT equals the confined concrete strength multiplied by concrete area
while that of DSTC approximately equals the load resisted by concrete and steel tube [73]. Given
the loading capacity around 5,622 kN, iteration processes were conducted for CFFT [67] and
DSTC [71] to determine the targeted cross-sectional parameters. The parameters so determined,
together with those of CFST [32], are summarized in Table 1. The calculated loading capacity of
CFFT and DSTC are 5,676 kN and 5,670 kN, respectively. Table 1 Detailed information of the
investigated elements

Component	Material property	CFST [32]	CFFT	DSTC
	Yield strength f_y (MPa)	355		360
Stool tubo	Ultimate strength f_u (MPa)	-		491
Steel tube	Outer diameter D_s (mm)	406	-	280
	Thickness <i>t</i> _s (mm)	6		8
Concrete	Compressive strength f'_{co} (MPa)	35	40	40
Colletete	Reinforcement area A_s (mm ²)	4517	-	-
	Elastic modulus E_{frp} (GPa)	-	44	44
GFRP tube	Inner diameter D (mm)	-	300	350
	Thickness t_{frp} (mm)	-	3.0	2.5

172 **3. Life-cycle model**

173 This section elucidates the studied life-cycle model. Material design of the illustrative examples is

174 primarily described, which will be integrated with unit environmental and economic impact for

175 final evaluation. The deterioration models for each alternative are then discussed, and in turn to

define maintenance schemes [79]. To cover major maintenance actions for each alternative,

FHWA [80] has suggested the investigated time period to be as long as enough. In balance, the life-cycle assessment herein is conducted over a period of 100 years. The results are also presented

178 life-cycle assessment herein is conducted over a period of 100 years. The results are also presented 179 in a cumulative way for each year such that the results for any shorter period can be referenced.

180 Fig. 2 depicts the framework of life-cycle assessment consisting of goal and scope definition,

inventory analysis, impact assessment, and interpretation of the results [50]. The following section

182 will define the goal and scope of the LCA-LCCA, while the inventory analysis and impact

assessment will be discussed in Section 4 along with the analysis results.



Fig. 2. Life-cycle assessment framework

184 **3.1 Goal and scope definition**

185 This section specifies the functional unit and system boundaries. In this study, the function unit is

186 defined as different column forms having similar loading capacity. As discussed in Section 2,

187 CFST, CFFT and DSTC are all designed to carry a nominal axial load of 5,622 kN.

188 This study aims to conduct a cradle-to-grave life-cycle assessment, covering all aspects of the 189 life-cycle process from raw materials to the end of construction waste. A typical life-cycle process 190 for constructions starts from producing structural components, including raw material extraction, 191 concrete mix, steel rolling, etc. Transportation of finished products and on-site construction is 192 considered as well. Maintenance over the whole lifespan is associated with durability of the 193 investigated design alternatives. As for the end-of-life stage, construction demolishing and waste 194 disposal are essential components. All considered phases together with their subsections are 195 schematically shown in Fig. 3.



Fig. 3. Flow chart of the studied life-cycle stage

196 **3.2 Maintenance schedule**

FRP composites are expected to outperform steel counterparts with respect to enduring behaviorwhich, in turn, differentiates the corresponding maintenance schedules. Herein, previously

established deterioration models for CFST, CFFT, and DSTC are incorporated to determine the relevant maintenance schedules. It is assumed that repair actions are taken when the design alternatives are subjected to a certain level of performance loss. According to the time-dependent corrosion model proposed in [81], a CFST column subjected to constant compressive loading and harsh ambient has 10 % performance loss at around year 10.

A series of accelerated experiments were conducted to study the deterioration behavior of GFRP-confined concrete solid columns and GFRP-concrete-steel double-skin tubular columns [82, 83]. The testing regime was designed to represent severe weather exposure for a period of 20 years. Both CFFT and DSTC will end up retaining slightly more than 90 % performance under constant compression loading, i.e., 10 % performance loss at year 20.

To sum up, the assumed baseline maintenance schedule for CFST is every 10 years while the alternatives of CFFT and DSTC will be repaired every 20 years. 10 % original materials will be replaced for every repair action. This is admittedly a simplified maintenance model, which can be easily updated once given more available information associated with deterioration.

213 **3.3 Life-cycle unit processes**

As depicted in Fig. 3, the first life-cycle stage is production, covering the whole manufacturing process from raw material 'in the earth' to finished products ready for use on site. For concrete, commonly used ingredients include ordinary Portland cement, river gravel, and sand.

Upon completion of the production process, the constituent materials will be transported to the construction site for assembling. All materials are assumed to be transported from local suppliers or those from neighboring cities. Locations were marked on Google map so that mean values of driving distances could be determined accordingly (Table 2).

221

Phase	Concrete (x_{14})	Steel (x_{15})	FRP (x_{16})
Average distance (km)	34	48	66
Maximum distance (km)	45	110	175
Minimum distance (km)	18	4	4
Number of spots	5	8	4

The following construction phase is associated with activities like mixing, vibrating or pumping concrete and putting shuttering or construction elements in place. Environmental and economic impact at this stage tends to relate to energy consumption and labor intensity. The investigated life-cycle stage, i.e. production, transportation, and construction, that incur at year 0 is also known as the 'initial stage'.

As discussed in Section 3.2, it is assumed that CFST will be repaired every 10 years while the alternatives will be maintained every 20 years. Common practice in previous studies is to take environmental and economic impacts produced from maintenance actions as a ratio of those from the initial stage [11, 13]. Given a predefined 10 % performance loss at the point where maintenance actions are implemented, the baseline maintenance-to-initial ratio is taken as 0.1, meaning that 10 %
 material will be replaced or repaired during each individual maintenance action.

233 During the end-of-life stage, the structural components are expected to be demolished and the 234 resultant construction wastes need to be disposed of in proper ways. Disposal techniques in 235 industry include recycling, landfill, and incineration. Steel is considered as recyclable metallic 236 material that can be refabricated into new products [84]. Herein, the recycled steel weight is taken 237 as 90 % of the original, with the remaining ended up in landfill [13]. Both landfill and recycling 238 apply to concrete wastes. In fact, there has been a growing trend of reusing crushed concrete wastes 239 as aggregates in new concrete [85, 86]. Nevertheless, this seemingly sustainable scenario still has 240 downsides such as quality inconsistency and cost variance [87]. Therefore, concrete wastes are 241 assumed to be shipped to the nearest landfill plant after demolition. In contrast with concrete and 242 steel, neither landfilling nor reuse is the proper solution to FRP treatment. Owing to the potential 243 deleterious substances in composite materials, extra treatment may be required prior to landfill. 244 Incineration, which allows for energy recovery, is thus a more reasonable way for disposal of FRPs 245 [88].

246 **4. Computational process of life-cycle environmental impact and cost**

The previous section has elaborated the framework to assess life-cycle environmental impact and cost of the three cases. Given the assessment framework, the following sections present the computational process using established models.

4.1 Environmental impact results

To compute respective environmental impacts of each design alternative, CO₂ emissions 251 252 associated with individual unit process were collected from the Ecoinvent database [89] or open 253 literatures. The unit CO₂ emissions from producing cement, aggregate, steel tube, steel reinforcement, and GFRP composites are 0.951 kg CO₂/kg, 1.06×10⁻³ kg CO₂/kg [90], 1.802 kg 254 255 CO₂/kg [91], 1.106 kg CO₂/kg [92], and 2.63 kg CO₂/kg [39], respectively. CO₂ emissions from 256 the construction stage are believed to come from energy consumption of equipment, e.g. generator, 257 truck crane, vibrator, etc., and are associated with use hours. A value of 0.016 kg CO_2 emission 258 for constructing 1 kg concrete is referenced herein [93]. Consider half of the time for steel 259 construction, i.e. 0.008 kg CO₂ emits from constructing per kg steel [94]. As the installation of 260 FRP tube barely requires energy-consumed equipment, the unit CO₂ emission of FRP construction 261 is estimated as 0.004 kg CO₂/kg. Cho and Chae [95] has suggested 0.70×10^{-2} kg CO₂/kg and 262 0.379×10^{-2} kg CO₂/kg for concrete/steel landfill and steel recycling respectively, while incinerating 1 kg GFRP is associated with 0.61 kg CO₂ emission [96]. The abovementioned GWP 263 factors are listed in Table 3 along with the material consumption of the studied cases. 264

Given the predefined life-cycle model and inventory data collection, the total CO_2 emissions *Total*_{CO2} of investigated design alternative can be calculated as

$$Total_{\rm CO_2} = P_{\rm CO_2} + T_{\rm CO_2} + C_{\rm CO_2} + M_{\rm CO_2} + E_{\rm CO_2}$$
(6)

where P_{CO_2} , T_{CO_2} , C_{CO_2} , M_{CO_2} , and $E_{CO_2} = CO_2$ emission from production, transportation, construction, maintenance, and end-of-life stages, respectively [11, 97].

During the initial stage, CO₂ released from production phase is $P_{CO_2} = \sum_{i=1}^{n} m_i \cdot C_{mi}$, where m_i e the amount of the associated material *i* (kg or *t*), *n* = numbers of material types, and C_{mi} = the unit CO₂ emission associated with unit production process (kg CO₂/kg).

272 CO₂ emission produced from transportation process is $T_{CO_2} = \sum_{i=1}^{n} m_i \cdot d_i \cdot C_{ti}$, where $d_i =$ 273 transportation distances (km) and C_{ti} = the unit CO₂ emission for transporting 1 kg material (kg 274 CO₂/t·km).

275 CO₂ released from construction phase is computed as $C_{CO_2} = \sum_{i=1}^{n} m_i \cdot C_{ci}$, where C_{ci} = the unit CO₂ emission factor associated with unit construction process (kg CO₂/kg). Based on previous 276 277 durability experimental data [81-83], 10 % performance loss is assumed every 10 years for CFST 278 and every 20 years for CFFT and DSTC. Therefore, 10 % of original materials are assumed to be 279 replaced or repaired associated with each maintenance action. CO₂ emitted from each maintenance action during the service stage is thus taken as 10 % of those emitted from the initial stage. Thus, 280 $M_{CO_2} = [(P_{CO_2} + T_{CO_2} + C_{CO_2}) \times 0.1] \cdot t_m$, where t_m = times of maintenance actions within the service 281 life. CO₂ emission from the end-of-life phase is $E_{CO_2} = \sum_{i=1}^{n} (m_i \cdot C_d \cdot m_{salvage} \cdot C_{salvage})$, where 282 $m_{salvage}$, C_d and $C_{salvage}$ = the amount of treated waste, the unit CO₂ emission for 283 284 landfilling/incinerating and recycling 1 kg waste, respectively.

285 The total life-cycle CO₂ emissions of all three investigated cases are compared as illustrated 286 in Table 3 and Fig. 4. CFFT obviously presents the lowest environmental impact, only half the emission of DSTC. DSTC emits about 50 % less CO2 than CFST does, which is second only to 287 CFFT regarding environmental benefits. Despite 50 % reduction of concrete consumption 288 289 comparing to CFFT, the presence of inner steel tube significantly increases CO₂ emission from 290 producing DSTC. However, the disadvantage is later offset by less emissions from maintenance 291 stage. For the benchmark scenario CFST, CO₂ emissions from the production and maintenance 292 phases account for similar fractions, around 50 %, of the total. For the other two cases, emissions 293 associated with production hold the largest shares while maintenance emission only account for 294 30 % of the total. Transportation and end-of-life processes only contribute limited share of 295 emission.

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Table 3 Life-cycle	environmental	impact results
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Column	Life-cycle stage	Consumed material		Amount	Unit	GWP coefficient (kg CO ₂ /kg)	GWP result (kg CO ₂ eq)
CEST	Droduction	Concrete	Cement	244	kg	0.951	232.04
Cr31	FIGUICION	Concrete	River gravel	732	kg	0.00106	0.78

]		Sand	427	kg	0.00106	0.45	
		St	eel tube	300	kg	1.802	541.29	
		Reinford	cing steel bars	177	kg	1.106	196.09	
				Sum			970.64	
		C	oncrete	1.52×34	t [.] km	0.6	31.01	
	Transportation	St	eel tube	0.3×48	t [.] km	0.6	8.64	
	Transportation	Reinford	Reinforcing steel bars		t [.] km	0.6	5.10	
				Sum			44.75	
		C	oncrete	1521	kg	0.016	24.33	
	Construction	St	eel tube	300	kg	0.008	2.40	
	Construction	Reinford	cing steel bars	177	kg	0.008	1.42	
				Sum			28.15	
	Maintenance			Sum			939.18	
		Concr	ete (landfill)	1521	kg	0.007	10.64	
		Steel tu	ube (landfill)	30	kg	0.007	0.21	
		Steel tu	be (recycling)	270	kg	0.00379	1.02	
	End-of-life	Reinforcement (landfill)		18	kg	0.007	0.12	
		Reinforcement (recycling)		160	kg	0.00379	0.60	
				Sum			12.61	
			Sum				1995.33	
			Cement	122	kg	0.951		
		Concrete	River gravel	215	kg	0.00106	117.17	
	Production	Concrete	Sand	418	kg	0.00106		
	Troduction		water	61	kg	0.00091	0.06	
		F	RP tube	31	kg	9.35	293.74	
				Sum			410.96	
		C	oncrete	0.85×34	t [.] km	0.6	17.34	
	Transportation		FRP	0.0314×66	t [.] km	0.6	1.24	
CFFT				Sum			18.58	
		C	oncrete	855	kg	0.016	13.68	
	Construction	ruction FRP tube			kg	0.004	0.13	
				Sum			13.81	
	Maintenance			Sum			177.34	
		Concr	ete (landfill)	855	kg	0.007	5.99	
	End-of-life	FRP tube (incineration)31kg0.61					19.16	
		Sum				25.15		
			Sum	1			645.85	
			Cement	63	kg	0.951		
DSTC	Production	Concrete	River gravel	111	kg	0.00106	60.26	
			Sand	215	kg	0.00106		

	Water	32	kg	0.00091	0.03
	Steel tube	268	kg	1.802	483.50
	FRP tube	30	kg	9.35	284.69
		Sum			828.48
	Concrete	0.419×34	t•km	0.6	8.55
T c c'	Steel tube	0.268×48	kg	0.6	7.72
Transportation	FRP tube	0.03×66	t•km	0.6	1.20
		Sum			17.47
	Concrete	419	kg	0.016	6.71
	Steel tube	268	kg	0.008	2.15
Construction	FRP tube	30	kg	0.004	0.12
	Sum				8.97
Maintenance		Sum			341.97
	Concrete (landfill)	419	kg	0.007	2.93
	Steel tube (recycling) 241	1	0.00379	0.92
End_of_life	Steel tube (landfill)	27	кg	0.007	0.19
Litu-01-inte	FRP tube (incineration)30kg0.61		0.61	18.57	
	Sum			22.61	
	S	um			1219.50

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298 Sensitivity analysis is conducted by varying the baseline maintenance-to-initial ratio from 0 299 to 0.4. Compared to CFFT and DSTC, the effect seems more prominent on CFST. The life-cycle

- 300 environmental impact increases significantly as the ratio increases as shown in Fig. 5. However,
- 301 CFFT remains the most environmentally-friendly alternative while CFST is the least in any case.



Fig. 5. Life-cycle environmental impact under different maintenance-to-initial ratios

302 **4.2 Life-cycle cost results**

316

303 In this section, initial cost refers to the sum of material/production cost, transportation cost, and 304 construction cost. Individual costs of related materials are displayed in Table 4. Unlike 305 environmental impact that is associated with energy consumption of equipment use, construction 306 cost has more to do with labor use over machinery use with respect to concreting, erection, 307 installation, etc. A labor rate, referred to the unit labor-to-material cost ratio, is introduced to 308 roughly estimate the construction cost. It is well known that labor costs vary from one country to 309 another. Average labor rate of Australia, UK, and US for concrete, steel reinforcement, and 310 structural steel derived from [98] are 2.32, 0.2, and 0.27, respectively. As FRP tubes are reportedly 311 much less labor-intensive and easier for installation [35, 99], the labor rate for GFRP construction is accordingly reduced to 0.1. The same maintenance schedule is applied to life-cycle cost analysis, 312 313 with each maintenance action costing 10 % of the initial cost. The unit costs related to end-of-life 314 stage are also listed in Table 4, where the costs were converted with present inflation rate via the 315 calculator provided on http://inflationdata.com/Inflation/Inflation_Rate/HistoricalInflation.aspx.

Item	Unit price (USD)	Source		
Ready mix concrete C35	177 \$/m ³	https://twisitymagdymin.gom/miggligt/		
Ready mix concrete C40	186 \$/m ³	https://tricityreadymix.com/price-nst/		
Carbon steel bars	0.63 \$/kg	https://worldsteelprizes.com/		
Carbon steel sections	0.78 \$/kg	https://wondsteetpinces.com/		
GFRP tube	10.71 \$/m ²	Local supplier		
Demolition-concrete	124.2 \$/m ³	[13]		
Landfill rate	0.089 \$/kg	[13]		
GFRP incineration	0.23 \$/kg	[100]		
Carbon steel-scrap value	0.11 \$/kg	[13]		

Table 4 Unit cost data of the studied raw materials

317 Given the inventory data in Table 4 and the consumed material amount listed in Table 3, the 318 total life cycle cost can be computed as follows [13]:

$$LCC = \sum_{t=0}^{T} \frac{C_{\rm t}}{(1+\gamma)^t} \tag{7}$$

319 where t = the year of incurred cost, T = the investigated period, $C_t =$ the cost incurred at the 320 corresponding year, and $\gamma =$ the monetary discount rate.

321 Incorporating discount rate is to reflect the potential monetary inflation or deflation over the 322 service life. Herein, the discount rate is taken as 3.3 % [101]. The life-cycle cost $C_t = C_P + C_C + C_C$ $C_M + C_E$, where C_P , C_C , C_M , and C_E refer to production, construction, maintenance, and end-of-life 323 cost, respectively. The discount rate is not applied to the initial stage, i.e., production and 324 construction, as these activities all incur at year 0. Therefore, $C_P + C_C =$ 325 $\sum_{i=1}^{n} m_{m,i} \cdot c_{m,i} + w_i \cdot \sum_{i=1}^{n} m_{m,i} \cdot c_{m,i}$, where $c_{m,i}$ = unit price of manufacturing material *i* (USD/kg or 326 USD/m³) and $w_{m,i}$ = the predefined labor ratio. Considering the monetary fluctuation over the 327 studied period, $C_M = \sum_{t=1}^T (C_P + C_C) \cdot w$ and $C_E = m_{E,i} \cdot c_{E,i} - m_{s,i} \cdot c_{E,i}$, where w = the assumed 328 329 percentages of maintenance cost to the corresponding production and construction costs, $m_{E,i}$ = 330 amount of material to be disposed (kg or m^3), $c_{E,i} = \cos t$ of demolishing, incinerating, landfilling or recycling a unit amount of material i (USD/kg or USD/m³), and $m_{s,i}$ = amount of material to be 331 332 recycled (kg).

333 Given the presumably defined life-cycle model and unit cost inventory, life-cycle costs for 334 three design alternatives were accordingly calculated using Eq. (7). The aggregated production and 335 construction cost is denoted as the initial cost incurred at year 0. Fig. 6 presents the life-cycle cost breakdown for the initial, maintenance, and end-of-life stages. CFST apparently has the cheapest 336 337 initial cost with CFFT second to it. The initial cost of CFFT and DSTC are nearly 14 % and 30 % 338 more than that of CFST, respectively, mainly attributed to higher material price of FRP composites 339 and structural steel. That said, the drawback of FRP initial cost is offset at later stage, i.e., service 340 duration, by lower maintenance cost. CFFT and DSTC have around 50 % maintenance cost 341 reduction compared to their CFST counterpart. Due to the higher salvaged value of steel, DSTC 342 has the lowest end-of-life cost among three. In balance, CFFT almost costs the same as does CFST 343 despite the higher initial cost.



Fig. 6. Comparison of life-cycle costs at different stages

The life-cycle cost results are also presented in a way such that the costs are added up to the total value (Fig. 7). Given the baseline discount rate of 3.3 %, CFST starts with initial cost advantage but ends up with a similar life-cycle cost with CFFT due to more intensive maintenance actions throughout the studied time period.



Fig. 7. Comparison of life-cycle costs over time given $\gamma = 3.3\%$

348 The results of local sensitivity analysis are presented in Figs. 8 and 9. Fig. 8 reveals that

349 discount rate has significant effect on the resultant life-cycle costs. CFST is the costliest alternative

given discount rates less than 1 %. Given discount rates between 1 % and 3.3 %, life-cycle cost of

351 CFST remains intermediate between CFFT and DSTC. As discount rate increases, life-cycle costs

352 of CFST and CFFT come closer with DSTC costing the most.



Fig. 8. Life-cycle cost given varied discount rates

As illustrated in Fig. 9, life-cycle cost increases as maintenance-to-initial ratio increases. DSTC is the costliest alternative given a maintenance-to-initial ratio of less than 25 %. Given a ratio between 10 % and 25 %, the life-cycle cost of CFST remains intermediate between CFFT and DSTC. DSTC remains comparatively expensive than CFFT irrespective of the maintenanceto-initial ratio. As the ratio increases, the life-cycle cost of CFST becomes the most expensive alternative of all.



Fig. 9. Life-cycle cost under different maintenance-to-initial ratios

359 Taken together the deterministic life-cycle results, FRP-confined scenarios display prominent

360 environmental merits in the long run. DSTC, though costs more than CFST over the studied period,

361 remains a potentially advantageous alternative when more practical conditions, e.g. flexural

362 resistance, seismic performance, fire resistance, etc., are considered [20].

363 5. Probabilistic assessment and sensitivity analysis considering uncertainties

364 **5.1 Sensitivity analysis techniques**

The above deterministic results have shed some lights on life-cycle environmental impact and cost assessment of the investigated three design alternatives. It is worth noting that the deterministic results were derived from historical datasets and reasonable presumptions without considering the uncertainties during life-cycle modeling process.

To embody the effect of uncertainty propagation, influential factors are identified and given respective distributions, thereby obtaining probabilistic life-cycle results. This process is implemented via Monte Carlo Simulation (MCS) which entails *N* times random sampling.

A subsequent global sensitivity analysis is conducted to determine how input parameters would simultaneously affect the outcome. Sobol's method is introduced to compute both the first order effects and total effects, the results are presented to reveal the sensitivity levels of each individual input parameter.

The fundamental of Sobol's method is briefly introduced as below. The Sobol's total effect index (STE) measures how much input parameter *i* affects the output, with all potential interactions with other parameters l, m, k... taken into account:

$$S_i^{\text{STE}} = S_i + S_{il} + S_{im} + \dots + S_{ilm} + \dots + S_{ilm\dots k}$$

$$\tag{8}$$

Given an output function y = f(x) entailing *n* input parameters, an input parameter space $\Omega^n = 380$ (*x_i*| *i* = 1, 2, ..., *n*) is created. With a constant $f_0 = \int_{\Omega^n} f(x) dx \approx \frac{1}{N} \sum_{n=1}^N f(x_n)$, the total output variance *S* is determined by

$$S = \int_{\Omega^{n}} f^{2}(x) - f_{0}^{2} \approx \frac{1}{N} \sum_{n=1}^{N} f^{2}(x_{n}) - f_{0}^{2} S$$
(9)

382 The first-order Sobol's index S_i is computed as

$$S_{i} = S - \frac{1}{2} \left[f(x) - f(x_{i}, x_{-i}) \right]^{2} dx dx_{-i} \approx S - \frac{1}{2N} \sum_{n=1}^{N} \left[f(x_{n}) - f(x_{in} - x_{-in}) \right]^{2}$$
(10)

383 The total Sobol's index S_i^{STE} is computed as [55, 102]

$$S_{i}^{STE} = \frac{1}{2} \int \left[f(x) - f(x_{i}', x_{i}) \right]^{2} dx dx_{-i}' \approx \frac{1}{2N} \sum_{n=1}^{N} \left[f(x_{n} - f(x_{in}', x_{-in})) \right]^{2}$$
(11)

where *N* is MCS sampling size and x_{i} is the vector complementary to x_{i} . A flowchart that describes the uncertainty quantification and global sensitivity analysis procedure is depicted in Fig. 10.



Fig. 10. Flowchart of uncertainty and global sensitivity analysis [102]

386 **5.2 Life-cycle environmental impact considering uncertainties**

387 Referring to [42] and [103], the sensitive parameters include key inputs, e.g., GWP coefficients 388 and maintenance schedule, which are characterized as uniform and normal distributions, 389 respectively. Tables 2, 5, and 7list the random distributions of the selected sensitive parameters, 390 where each of the parameter is designated a number. Given the probability density function of each 391 input parameter, Monte Carlo simulation, the most common technique for uncertainty 392 quantification, is adopted to assess the life-cycle environmental impact. Accordingly, the relevant 393 results are expressed in terms of a probability density function, from which the expected value and 394 the variances can be computed.

395

Table 5 GWP coefficient of individual item at each life-cycle stage

Phase		Parameter	Assumed value	Distribution (±)	Source
		Cement (x_1)	0.951 kg CO ₂ /kg	40 %	
	R	iver gravel (x ₂)	1.06×10 ⁻³ kg CO ₂ /kg	20 %	[90]
Droduction		Sand (x_2)	1.06×10 ⁻³ kg CO ₂ /kg	20 %	
FIGURCHOIL		Steel tube (x_3)	1.802 kg CO ₂ /kg	20 %	[91]
	Rei	nforcing bars (x ₄)	1.106 kg CO ₂ /kg	20 %	[92]
	GFRP composites (x_5)		9.35 kg CO ₂ /kg	40 %	[39]
Transportation	CO ₂ emissions from 1 t·km of transportation (x_6)		0.6 kg CO ₂ / t·km	20 %	[103]
	Concrete (x_7)		0.016 kg CO ₂ /kg	40 %	-
Construction		Steel (x_8)	0.008 kg CO ₂ /kg	20 %	-
	FRP (x_9)		0.004 kg CO ₂ /kg	40 %	-
Cone		crete landfill (<i>x</i> ₁₀)	0.70×10 ⁻² kg CO ₂ /kg	40 %	
End of life	Stool	Recycling (x_{11})	0.379×10 ⁻² kg CO ₂ /kg	20 %	[95]
End-of-file	-or-me Steel	Landfill (x_{12})	0.70×10 ⁻² kg CO ₂ /kg	20 %	
	GFR	P incineration (x_{13})	0.61 kg CO ₂ /kg	40 %	[96]

Based on the MCS results, the maximum, 75% quartile, mean, 25% quartile, and minimum life-cycle CO₂ equivalent emissions are extracted and depicted in Fig. 11. The total CO₂ emission of CFST have the least variation which is around ± 25 % while the variations of CFFT and DSTC are both around ± 60 %. For the investigated three cases, the uncertainties embedded in the maintenance schedule are shown to contribute the most to the output variance.



Fig. 11. Life-cycle CO₂ emission results considering uncertainty

401 Fig. 12a shows the probability density function (PDF) of life-cycle CO_2 emissions associated 402 with the investigated design alternatives. CFFT obviously releases the least CO_2 in any case while 403 an overlap can be seen between CFST and DSTC. Fig. 12b quantifies the possibility that DSTC is 404 more environmental-friendly than CFST, where CI_{GWP} defines the relationship between two 405 impact results of CFST and CFFT, i.e., $CI_{GWP} = Z_{CFST}/Z_{CFFT}$. As shown in Fig. 12b, the possibility

406 that CFST releases more CO_2 than does DSTC is 0.8.





Fig. 12b. PDF of life-cycle CO₂ emission comparison

407 The Sobol's sensitivity results are listed in Table 6. It is found that for CFST and CFFT, 408 maintenance timing (x_{19}) is the most sensitive parameter. As for DSTC, GWP coefficient of steel 409 tube (x_3) and maintenance timing (x_{19}) are the two most sensitive parameters.

Sobol's indices	CFS	ST	CFF	Т	DST	ГC
Parameters	First-order	Total	First-order	Total	First-order	Total
<i>x</i> ₁	0.1062	0.1045	0.3219	0.3209	0.0344	0.0327
x_2	0.0056	0	0.0027	0	0.0021	0
<i>x</i> ₃	0.1159	0.1128	-	-	0.4229	0.4214
x_4	0.0477	0.0432	-	-	-	-
x_5	-	-	0.0447	0.0434	0.1459	0.1436
x_6	0.0056	0	0.0026	0	0.0021	0
<i>x</i> ₇	0.0056	0	0.0026	0	0.0021	0
x_8	0.0056	0	-	-	0.0021	0
<i>X</i> 9	-	-	0.0026	0	0.0021	0
x_{10}	0.0056	0	0.0026	0	0.0021	0
x_{11}	0.0056	0	-	-	0.0021	0
<i>X</i> 12	0.0056	0	-	-	0.0021	0
<i>x</i> ₁₃	-	-	0.0026	0	0.0021	0
x_{14}	0.0066	0.0010	0.0067	0.0038	0.0064	0.0043
<i>X</i> 15	0.0073	0.0015	-	-	0.0042	0.0023
<i>x</i> ₁₆	-	-	0.0029	0.0002	0.0022	0.00008
x_{19}	0.7380	0.7377	0.6318	0.6316	0.3947	0.3956

410 Table 6 Sensitivity analysis results of life-cycle environmental impact using Sobol's method

411 **5.3 Life-cycle cost considering uncertainties**

With respect to life-cycle cost, the sensitive parameters are selected as initial cost, discount rate, and maintenance timing (Table 7). Initial cost refers to the sum of production and construction cost, which covers almost all uncertainties of input parameters. The FHWA has suggested an acceptable range for discount rate within 3-5 % while a typical one hovers at 4 % [104]. Herein, a triangular distribution is assumed and the minimum, most likely, and maximum values are 0 %, 3.3 %, and 5 % respectively [101].

418

Table 7 Random variables for life-cycle cost analysis

Random variable		Distribution	Description	
Initial cost (x_{17})		Uniform	± 20 %	
Discount rate (x_{18})		Triangular	Min = 0, most likely = 0.33 %, max = 5 %	
Maintananaa timina	CFST	Normal	Mean = 10 years, COV = 0.15	
Maintenance timing $(r_{\rm er})$	CFFT		Mean = 20 years, $COV = 0.07$	
(X19)	DSTC	[42]	Mean = 20 years, COV = 0.11	

419 The life-cycle cost variation of CFST is around ± 40 %, which is slightly larger than those of 420 CFFT and DSTC, i.e., around 30 % (Fig. 13). A clear trend is that the end-of-life cost variations 421 $(\pm 90\%)$ are larger than maintenance cost variations (around $\pm 60\%$) which are larger than initial 422 cost variations $(\pm 20 - \pm 30\%)$. Therefore, the output variance of life-cycle cost is supposedly 423 attributed to the varying discount rate as its impact is cumulated with time (Equation 7).



Fig. 13. Life-cycle cost results considering uncertainty

424 As presented in Fig. 14a, PDFs of life-cycle cost show considerable overlaps between life-

425 cycle cost results of the three cases. PDFs of comparison between the three cases indicate that the

426 probabilities that CFST costs more than does CFFT, CFST costs more than does DSTC, and CFFT

427 costs more than does DSTC are 0.59, 0.25, and 0.14, respectively (Fig. 14b).





Fig. 14b. PDF of life-cycle cost comparison

The Sobol's sensitivity analysis results are presented in Table 8. It is found that maintenance timing is the most sensitive factor while discount rate is the least. Table 8 also shows that discount rate has no interaction with other parameters.

431

Sobol's indices	CFST		CFFT		DSTC	
Parameters	First-order	Total	First-order	Total	First-order	Total
<i>X</i> 17	0.1789	0.1765	0.4843	0.4764	0.4747	0.4747
x_{18}	0.0046	0	0.0086	0	0.0007	0
<i>x</i> ₁₉	0.8234	0.8211	0.5236	0.5157	0.5253	0.5253

Table 8 Sensitivity analysis results of life-cycle cost using Sobol's method

434 **6. Discussions: further applications**

The developed framework of the present study may be extended to other case scenarios such as beams or beam-columns subjected to static and dynamic conditions. This This section focuses on the potential application of the proposed method within other configurations and boundary conditions.

439 **6.1 Beams**

Fig. 15 illustrates typical forms of CFFT and DST beams (DSTB) studied in previous research.
Test results indicate that DSTB with an inner steel tube eccentrically placed within the outer FRP
tube outperforms the alternatives regarding flexural resistance and ductility [105, 106]. However,
slip between the infill concrete and steel tubes were spotted in [105, 107, 108]. Attempts were
made to address such bonding issues by adding steel rings [107] or shear connectors [106] which,

though effective, largely increase steel consumption and labour intensity.



⁴⁴⁶ Therefore, future studies may consider full-scale beams with similar shear resistance, flexural

432

433

⁴⁴⁷ resistance, crack limit, ductility, etc.[109], as comparable case studies for life-cycle environmental

⁴⁴⁸ impact and cost analysis.

449 **6.2 Seismic performance**

Test data regarding CFFT and DST columns, beams, and beam-columns subject to cyclic loading can be found in [22, 28, 110-112]. The seismic performance of CFST and CFFT beamcolumns were compared in [112], where CFST showed better ductility and energy dissipation than CFFT. However, the brittleness of CFFT was shown to be compensated by the presence of an inner steel tube [22, 28, 110, 111]. More uncertainties may be considered when analyzing seismic performance.

To associate life-cycle seismic performance with life-cycle environmental impact and cost analysis, earthquake scenarios that a structural element may be experienced during its lifetime need to be assumed [113]. The probabilistic seismic loss should be quantified considering all potential earthquake scenarios which, in turn, can be linked to economic and environmental impact [114].

461 **7. Conclusions**

462 A life-cycle assessment and life-cycle cost analysis (LCA-LCCA) was conducted in this study. The LCA-LCCA aims to evaluate life-cycle performance of comparative concrete-filled steel 463 tubular column (CFST), concrete-filled FRP tubular column (CFFT), and hybrid FRP-concrete-464 465 steel double-skin tubular column (DSTC) from environmental and economic perspectives. Apart 466 from deterministic results, uncertainty quantification and global sensitivity analysis results are also 467 presented. First order effects and total effects were computed and parameters were ranked 468 according to their contribution to the output variance of the life-cycle environmental impact and 469 life-cycle cost. The following conclusions can be drawn:

- As can be seen from the deterministic life-cycle environmental results, CFFT is the most environmental-friendly alternative of all with DSTC only second to it. In contrast, life-cycle CO₂ emission of CFST is as much as three times that of CFFT.
- Given a reasonable discount rate of 3.3 %, DSTC has the largest life-cycle cost while CFST and CFFT cost nearly 15 % less than DSTC does. A follow-up local sensitivity analysis demonstrates that the life-cycle cost varies as discount rate varies between 0 and 5 %. CFST is very likely to become the most cost-effective option with increasing discount rate.
- Despite that the deterministic results indicate that CFFT is the most environmentally and economically benefited alternative of all, the probabilistic results tell something different.
 CFST is potentially more environmental-friendly than DSTC if input uncertainties are considered. The probabilities of each alternative being more cost-effective are presented in the form of probability density functions. Finally, the key influential parameters are identified by means of Sobol's method.
- The above findings are limited to the simplifications and assumptions in some aspects. For instance, very limited information is available in terms of maintenance schedule of the investigated composite structural elements.

- The results are expected to provide more evidence for stakeholders to make decisions by considering the tradeoffs among multiple perspectives associated with novel structural systems. Life-cycle performance of composite structural elements in other case scenarios, e.g., columns, beams or beam-columns subject to static or seismic conditions, will be explored in future work.
- 491

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