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Shift in controlling factors of carbon stocks across biomes on the
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E-mail: huzm@igsnr.ac.cn**Keywords:** Qinghai-Tibetan Plateau, SOC mapping, machine learning algorithms, driving factors, uncertaintiesSupplementary material for this article is available [online](#)

Abstract

The Qinghai-Tibetan Plateau (TP) accumulated a large amount of organic carbon, while its size and response to environmental factors for the whole area remain uncertain. Here, we synthesized a dataset to date with the largest data volume and broadest geographic coverage over the TP, composing of 7196 observations from multiple field campaigns since the 1980s, and provided a comprehensive assessment of the size and spatial distribution of carbon pools for both plant and soils on the TP using machine learning algorithms. The estimated soil organic carbon (SOC) storage to 1 m depth was $32.01^{47.9}_{19.69}$ Pg ($11.72^{17.53}_{7.2}$ kg m⁻² on average), accounting for approximately 37.2^{55.6}_{22.9}% of China's SOC stock on its <30% land area. There was 15.52^{23.52}_{9.91} Pg C stored in grassland soils (1 m), which played as the largest C pool on the TP, followed by shrubland (7.52^{11.6}_{4.8} Pg) and forest (3.72^{5.36}_{2.5} Pg). The estimated plant C pool was 2.4^{5.16}_{0.95} Pg (1.03^{2.7}_{0.2} Pg in aboveground biomass (AGB) and 1.37^{2.45}_{0.75} Pg in belowground biomass). Soil and biomass C density presented a similar spatial pattern, which generally decreased from the east and southeast parts to the central and western parts. We found both vegetation and soil C (1 m depth) were primarily regulated by climatic variables and C input across the entire TP. However, main driving factors of the C stocks varied among vegetation types and depth intervals. Though AGB played as an important role in SOC variation for both topsoil (0–30 cm) and subsoil (30–100 cm), the strength of the correlation weakened with depth and was gradually attenuated from grassland to shrubland, and forest. The outcomes of this study provided an updated geospatial estimate of SOC stocks for the entire TP and their relationships with environmental factors, which are essential to carbon model benchmarking and better understanding the feedbacks of C stocks to global change.

1. Introduction

The Qinghai-Tibetan Plateau (TP) is the largest and highest plateau on the earth with unique vegetation types and climate conditions. The soil on the TP is a large carbon (C) reservoir because of the low soil carbon turnover rate due to cold weather [1, 2]. While the rapid warming rate in recent decades has long made it been recognized as an amplifier of global climate [3, 4], and would potentially accelerate the carbon loss to the atmosphere [5, 6]. Vegetation on the TP also plays an essential role in the regional C cycling given its high sensitivity of productivity to climate change [7, 8]. It is thus crucial for quantifying the carbon pools on the TP, clarifying their spatial patterns, as well as the dominant factors controlling the spatial variations of the C stocks [9, 10].

Great attention has been paid to the soil organic carbon (SOC) stock on the TP and its spatial pattern owing to its high carbon density. Early studies have been conducted on SOC stocks in 0–1 m depth on the TP [11–14], the size ranged from 18.37 to 49 Pg (1 Pg = 10^{15} g) with considerable uncertainty. These estimations were mostly based on data from China's 2nd national soil survey (mid-1980s) with sparse sampling sites. The upscaling methods (soil group area-weighted) are usually inadequate to account for spatial heterogeneity, which resulted in large uncertainties. Recent field survey campaigns have supplemented a large number of soil samplings that refer to broader geographic coverage and deeper soil cores [15–17]. Machine learning algorithms with geospatial and remote sensing information have been applied for SOC mapping to better capture the spatial heterogeneity of frozen-affected areas on the TP [2, 15, 17]. However, these studies mainly focus on grasslands [1, 18] or permafrost areas [2, 14–16]. The C stocks in forests and shrublands on the TP remain unclear. In addition, compared with soil, less attention has been paid to the vegetation C stocks on the TP regarding its small size compared with the soil C pool and the scanty of field observations [8]. Therefore, to clarify the size and spatial pattern of soil and vegetation C pools and reduce the uncertainty, it is necessary to provide an updated assessment with advanced compiled datasets and improved upscaling/mapping methodology for the entire TP.

Spatial variations in soil C are affected by environmental factors, including climate [19], land types [20, 21], soil properties [22, 23], and geomorphology [10]. While sparse sampling points or incongruous study areas usually lead to inconsistent results. For example, Yang *et al* [1] found a positive relation between SOC and mean annual temperature (MAT) in alpine grassland, while Mishra *et al* [10] claimed that the relationship was not significant in the permafrost area (mainly grassland) on the TP. In addition, previous studies primarily focused on a single vegetation type (i.e. grasslands) or specific geological

region (i.e. permafrost), while lacking the exploration of the relationships between SOC and environmental factors of the entire Qinghai-TP.

To provide a comprehensive estimate of carbon stocks especially SOC across the entire TP, we collected 7196 observations from multiple field sampling campaigns and peer-reviewed articles, including 4809 measures for SOC with four depth intervals down to 2 m and 2387 sites for plant above- and belowground biomass (AGB and BGB). By integrating 48 environmental variables referred to meteorological dataset, edaphic and topographic factors and their interactions, we interpolated these site-level C densities to the regional scale using machine learning algorithms. We aimed to (a) comprehensively quantify the size and spatial distribution of C stocks on the TP and (b) clarify the main driving factors of the spatial variation in vegetation and soil carbon pools with depth.

2. Materials and methods

2.1. Study area

The Qinghai-TP lies between 26°0′–39°47′N, 73°19′–104°47′E, with an area of $\sim 2.5 \times 10^6$ km² [24]. Its climate is characterized by intense solar radiation, long durations of sunshine, low temperature, and extensive diurnal temperature range. Grassland as the dominant ecosystem on the plateau occupying over 60% of the total area [25]. Forests and shrublands summed together to $\sim 5.35\%$ of the land surface and scattered in the central and western TP. Barren lands were mainly distributed in the northern TP with dry and cold conditions.

2.2. Data collection

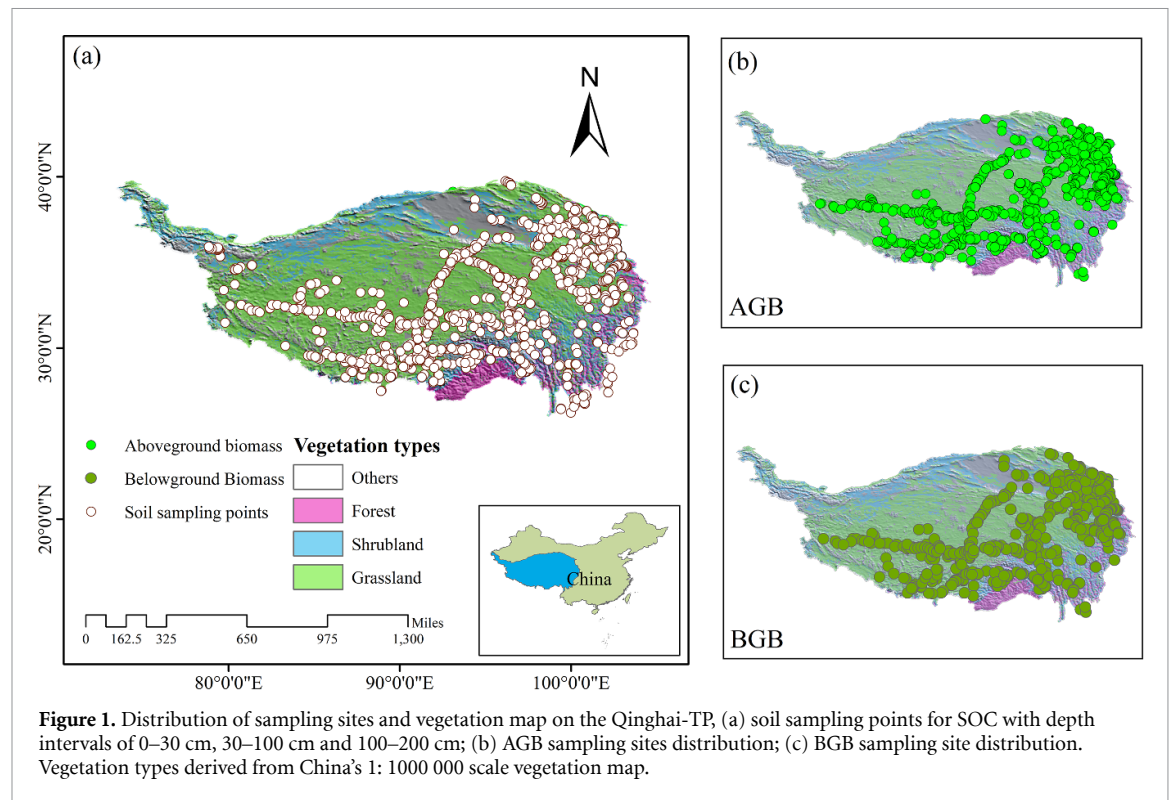
2.2.1. SOC and biomass data

We collected 4809 SOC observations and 2387 biomass data from multiple regional field sampling campaigns across the plateau since the 1980s [1, 2, 15–17], and established the most representative soil dataset to date, which covers the main vegetation types and spatial scope (see supplementary text in details). The sites distribution map is shown in figure 1.

SOC stock (SOC_s, kg C m⁻²) was calculated by the following equation (1):

$$\text{SOC}_s = \frac{1}{100} \times \sum_{j=1}^n \text{SOC}_j \times (1 - C_j) \times \text{BD}_j \times D_j \quad (1)$$

where j is the soil horizon (i.e. 1, 2, 3, ..., n), SOC is SOC concentration on a mass basis (g kg⁻¹), C_j is the volume percentage of gravel fraction, BD is the soil bulk density corrected (g cm⁻³), and D_j is the thickness of the j th horizon (cm). There were $\sim 7.5\%$ (360/4809) SOC measurements missed their corresponding bulk density data at specific sites and layers, we hereby built an empirical (nonlinear) relationship between SOC and BD [18] to estimate the missing BD (supplementary figure S1).



Generally, each site in our dataset has multiple layers with different depth intervals. Sites with the maximum depth larger than 30 cm were used to calculate SOC stock of 0–30 cm. While sites with the maximum depth larger than 70 cm were used to estimate SOC stock of 30–100 cm, 0–100 cm, sites with depth larger than 100 cm were used to estimate SOC stock of 100–200 cm. In sum, we got 2019 sites for 0–30 cm, 1181 sites for 30–100 cm, 1302 sites for 0–100 cm, and 307 sites for 100–200 cm. These measurements are sufficient for us to assess the C stocks of the TP and further explore its controlling factors.

2.2.2. Environmental variables

To detect the relationship between environmental variables and SOC stock of TP and subsequently use a machine learning approach to predict its size and spatial pattern, we compiled an environmental data collection with 61 covariates (supplementary table S1) that refers to meteorological variables (i.e. evapotranspiration, temperature, etc), topographic indexes (i.e. slope, curvature, etc), soil physico-chemical properties (i.e. pH, clay, silt, etc) and their interactions for corresponding observation sites. For detailed information, see supplementary text. The minimum distance of paired observations was ~ 3.6 km on average, considering the representative and sample density compared to former studies [2, 15], all the predictors were uniformly resampled into $5 \text{ km} \times 5 \text{ km}$ resolution, and the corresponding covariates of the sampling sites were extracted for further analysis.

2.3. Principal component analysis (PCA) and reclassify the covariates

Rotated PCA (details see supplementary text) was used to reduce dimension of the independent variables and further improve the interpretation of larger groups of controls [26]. According to the variable loadings of rotated PCA and correlations among the meteorological variables (supplementary figure S2, tables S2–S5), we found temperature related 11 bioclimatic factors and precipitation related eight bioclimatic factors were always classified into different components. In addition, MAT has high correlation with variables, such as solar irradiation (RD), vapor pressure deficit (VPD) and ratio of MAT to solar irradiation (MAT_RD), and mean annual precipitation (MAP) has high correlation with variables like aridity (AI), evapotranspiration (ET), soil moisture (SM) and ratio of ET to solar irradiation (ET_RD) (supplementary tables S2–S5), we thereby grouped these principal components (PCs) related to these climatic variables into two larger categories: temperature related factors, including 11 temperature derived bioclimatic factors, RD, VPD and MAT_RD, and water related factors, including eight precipitation derived bioclimatic variables and AI, ET, SM and ET_RD. The rest PCs were further grouped into three different environmental categories including AGB, which always related with diurnal temperature range, soil physico-chemical properties and topographic factors (supplementary tables S2–S5). Machine learning algorithms was then used to build the relationships between these PCs and different C stocks and further identify the importance of each PC.

2.4. Carbon mapping

Combined with the spatial environmental covariates, we trained and tuned (supplementary text 1.3.1) the models with several machine learning algorithms, including support vector machine (SVM), random forest (RF), extreme gradient boosting and feedforward neural network. To test the model performance, we used the five-fold cross-validation approach and calculated the mean square errors as the criteria [27]. RF finally outperformed other algorithms (supplementary figure S3, table S7) and was adopted to mapping C stocks (at a resolution of 5 km) on the TP.

2.5. Uncertainty quantification of the mapping

We used two methods to quantify the uncertainty of our estimate, absolute uncertainty and relative uncertainty (details see supplementary text). Briefly, for the former one, we used a quantile regression forest approach to calculate the interquartile of each pixel (difference between the 75th and 25th percentiles), further the summed and averaged value of the 25th and 75th percentiles were used to represent the uncertainty of OC stocks on the TP [2]. To get the relative uncertainty, the sensitivity and model structure uncertainty were then summed up to represent the total uncertainty, which then was divided by the median values of each pixel to obtain the relative uncertainty [2, 28, 29].

2.6. Statistical analysis

If the *P*-value of Kolmogorov Smirnov [30] test was smaller than 0.05, we used log transformation for the data for further model establishing and analysis. Biomass C, including AGB and BGB, in this study were all transformed into base-10 log scale for further analysis. They were transformed back after model prediction. All statistical analysis were conducted in R platform [31]. The packages we mainly used were listed in supplementary text.

3. Results

3.1. Relationships between different carbon stocks and climate

The relationship between C stocks (bin-averaged) and climatic factors varied with vegetation types (figure 2). Specifically, both grassland AGB and BGB showed a significant positive tendency with MAP ($p < 0.01$, figure 2(a)), but their correlations with MAT were sharply weakened (figure 2(d)). While for shrubland, a decreasing relationship with MAT was found for both its AGB and BGB ($p < 0.01$, figure 2(e)). In addition, only AGB in forest showed a positive relationship with MAP ($p < 0.01$, figures 2(c) and (f)).

Moreover, the correlations between SOC and MAP gradually shifted from positive in grassland ($p < 0.01$) to negative in forest ($p < 0.01$) for both topsoil and subsoil (figures 2(g) and (i)). SOC stock

in shrubland showed an opposite trend to MAP at the two depth intervals, increasing in topsoil and decreasing in subsoil. However, the response of SOC stock to MAT was almost synchronized at both soil layers (figures 2(i)–(l)), which all showed a negative relationship with MAT for the three vegetation types.

3.2. The relative importance of environmental variables to vegetation and soil C stocks

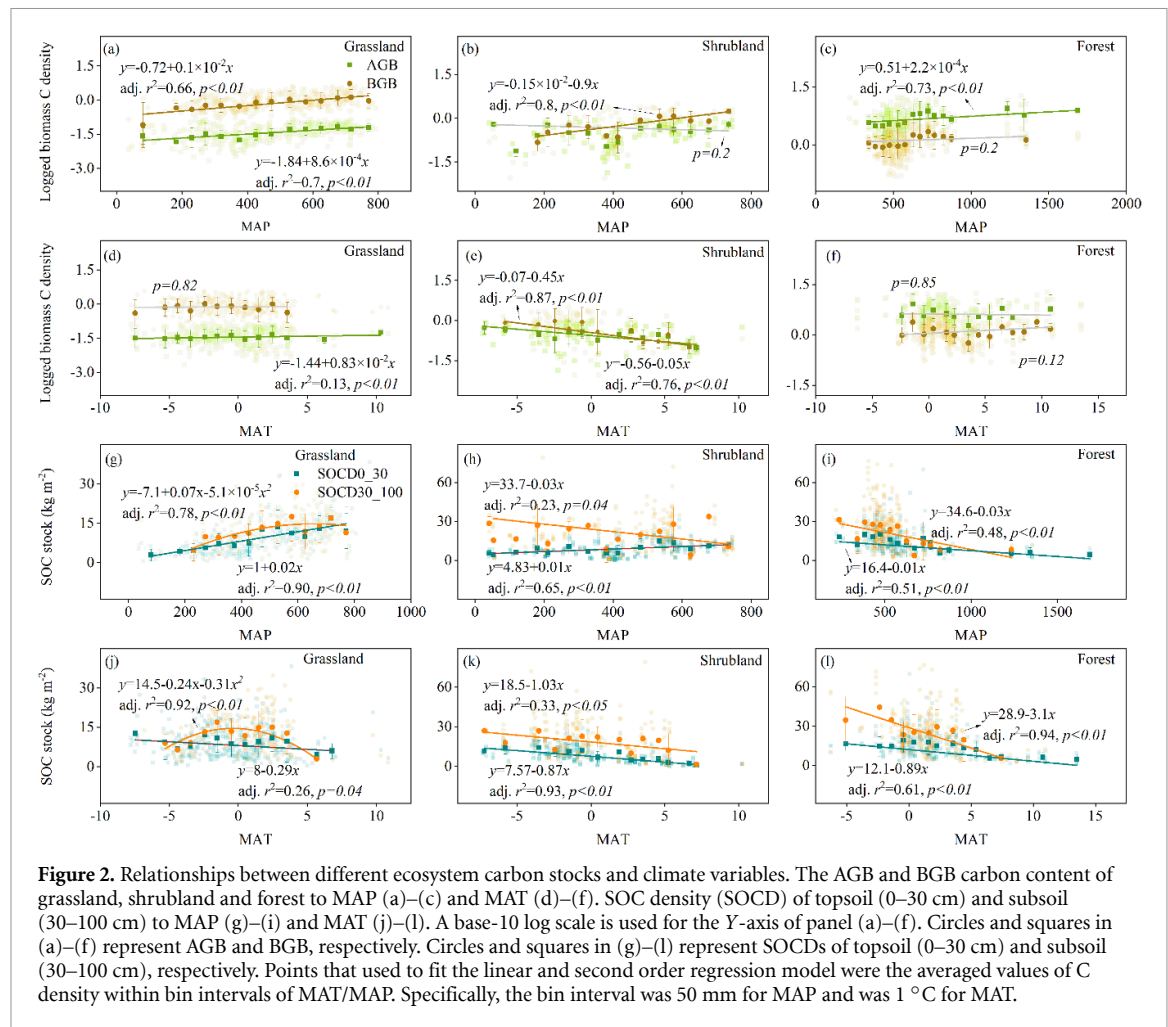
The primary governing environmental category differs for vegetation and soil C stocks on the TP. For AGB, temperature and water related factors contributed the most to its spatial distribution, taking half of its variation across the TP (figure 3(a)), indicating the dominant role of climatic variables in shaping amount of AGB. While for the individual land covers, the main driving environmental category changed from precipitations related factors in grassland to soil properties and temperature related factors in shrubland and forest (figure 3(a)), respectively.

For SOC stock, AGB played a primary role for both topsoil (0–30 cm) and subsoil (30–100 cm) (figures 3(b) and (c)), demonstrating the importance of plant-derived C input to SOC stocks on the TP. For topsoil, both AGB and water related factors (MAP, etc) showed positive effect on the SOC stocks across the TP, while temperature related factors like temperature fluctuation ratio of monthly to yearly (isothermality, supplementary figure S5(e)) almost showed a linear negative effect on SOC stocks. However, the dominant driving environmental category varied among different vegetation types (figure 3(b); supplementary figures S4(e)–(h)), which changed from water related factors in grassland to topographic factors, temperature related factors in shrubland and forest (figure 3(b)), suggesting divergent controlling factors among land covers on the TP.

Unlike topsoil, besides AGB, there were no exceptional environmental categories contributing to subsoil SOC stocks. Categories such as temperature related and soil physico-chemical properties had almost the same contribution to SOC in sublayers throughout the TP (figure 3(c)). Moreover, the dominant regulatory categories were relatively consistent for the individual land cover (figure 3(c)), which were all referring to water and temperature related factors, revealing a relative complex mechanism in regulating the SOC in deeper soils. Specifically, the influence of individual PCs on different SOC stocks with depth exhibited the same trend, which all increased with precipitation and decreased with temperature related factors (supplementary figures S4(i)–(l)).

3.3. Size and spatial patterns of vegetation and soil C stocks on the TP

The spatial distribution of both vegetation and soil C densities presented a similar pattern. Namely, the C density decreased from the east and southeast parts to the central and western parts and then increased



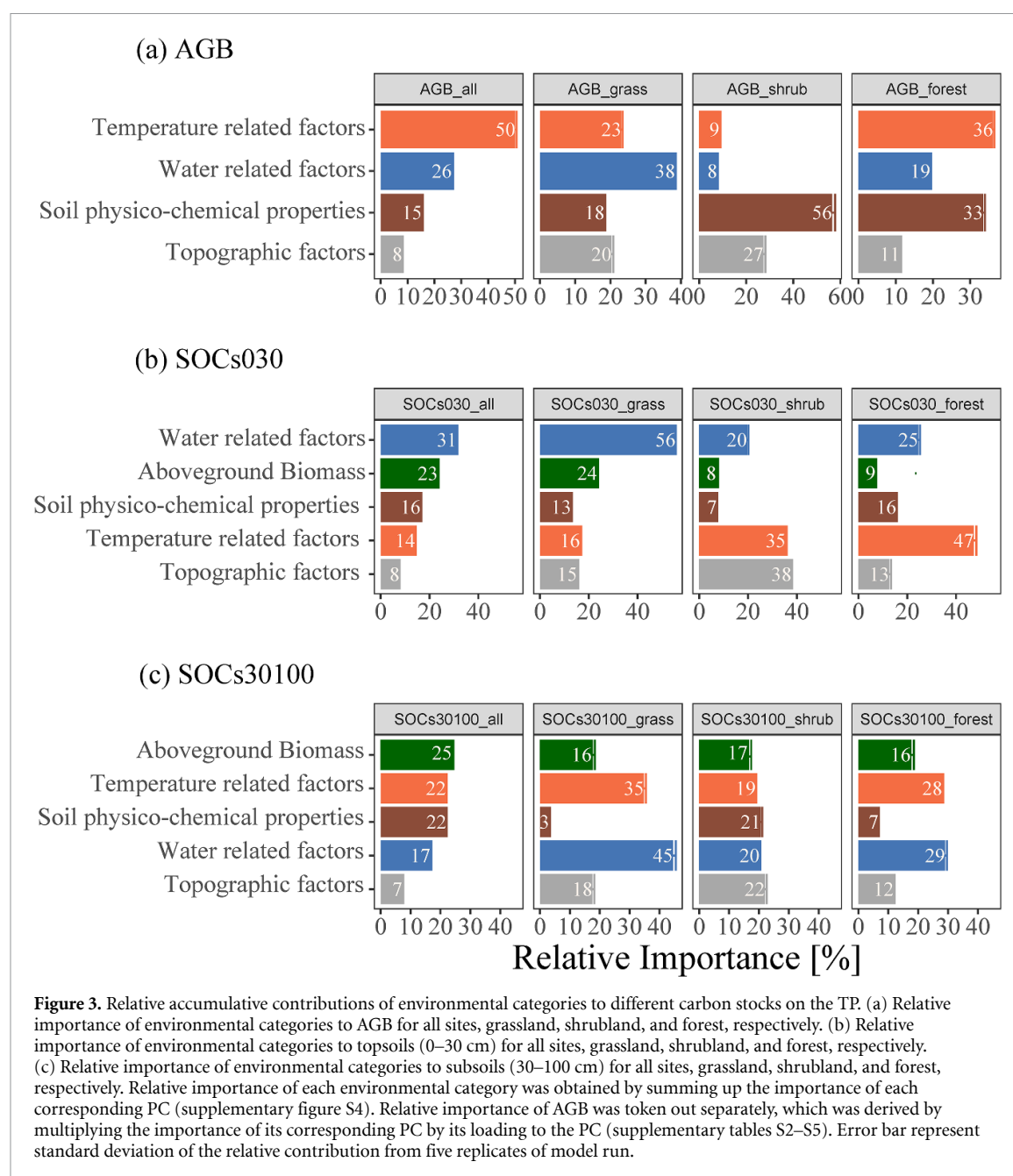
again at the northwest edge within Xinjiang Uygur autonomous region (figures 4(a)–(c) and (g)–(i)). While, their uncertainties showed an opposite tendency between plant and soil carbon pools (insets of figures 4(a)–(c) and (g)–(i)). The uncertainties of vegetation C density exhibited a similar pattern with its density distribution, suggesting proportional uncertainties to their C density. While the highest uncertainty for soil C densities within 1 m depth was concentrated in the northwest part and then decreased to the eastern part (insets of figures 4(g) and (h)).

There was a total of $2.4^{5.16}_{0.95}$ Pg C (confidence interval of upper and lower estimate indicated 75th and 25th percentiles to quantify the prediction uncertainty) stored in vegetation on the TP. Of which $1.03^{2.7}_{0.2}$ Pg was allocated to AGB and $1.37^{2.45}_{0.75}$ Pg was assigned to BGB (figures 4(a)–(c)). The average carbon density of AGB was increased from 0.10 kg m^{-2} in grassland to 0.52 and 2.86 kg m^{-2} in shrubland and forest, respectively. Correspondingly, their carbon stock size increased from $0.16^{0.5}_{0.04}$ Pg in grassland to $0.27^{0.8}_{0.06}$ and $0.56^{1.22}_{0.09}$ Pg in shrubland and forest. In contrast, the carbon stored in BGB was decreased from $0.71^{3.1}_{0.36}$ Pg in grassland to $0.29^{0.49}_{0.18}$ and $0.26^{0.41}_{0.16}$ Pg in forest (figures 4(d)–(f)).

The estimated SOC stock on the entire TP was $32.01^{47.9}_{19.69}$ Pg ($11.72^{17.53}_{7.2}$ kg m^{-2} on average) for 0–100 cm soil and $40.11^{65.06}_{24.55}$ Pg for 0–200 cm depth ($14.68^{23.81}_{8.98}$ kg m^{-2} on average). Though the average SOC stock (0–100 cm) increased from grassland (9.95 kg m^{-2}) to forest (19.12 kg m^{-2}), the opposite trend of their total storage was found, the largest SOC pool was grassland soils ($15.52^{23.5}_{9.91}$ Pg, 48.5%), of which $10.2^{15.3}_{6.62}$ Pg stored in alpine meadow and $3.87^{5.88}_{2.36}$ Pg stored in alpine steppe, followed by shrubland ($7.52^{11.6}_{4.8}$ Pg, 23.5%) and forest ($3.72^{5.36}_{2.5}$ Pg, 11.6%) (figures 4(f)–(h)). There was $14.27^{20.15}_{9.77}$ Pg SOC stored in 0–30 cm depth, which accounted for 44.57% of 0–100 cm SOC storage, and of which $7.81^{11.0}_{5.4}$ Pg stored in grassland ($5.47^{7.4}_{3.89}$ Pg stored in alpine meadow and $1.81^{2.82}_{1.18}$ Pg stored in alpine steppe), $3.13^{4.34}_{2.2}$ Pg in shrubland and $1.6^{2.23}_{1.11}$ Pg in forest, respectively. All indicated that grassland, especially its alpine meadow system, played as the largest C reservoir to the TP.

3.4. Relationships between vegetation and soil C stocks

SOC and AGB were closely related on the TP, especially for grassland and shrubland (figure 5). SOC increased significantly with the increase of AGB in



grassland for the topsoil ($p < 0.01$) (figure 5(a)). Though the strength of the correlation (bin averaged) decreased in deeper soil (30–100 cm), the coefficient still at a high level (0.78) (figure 5(b)). There was also a strong positive correlation between AGB and SOC in shrubland in topsoil (0–30 cm), of which the coefficient was 0.70 ($p < 0.01$). For the forest, though there was a negative trend between AGB and SOC, they are not statistically significant ($p > 0.05$) (figure 5).

4. Discussion

4.1. Carbon stocks on the TP

Based on the unique compiled dataset, widely collected environmental variables, and the advances of machine learning (RF) approach, this study provided a comprehensive assessment of the TP for both its

biomass and soil carbon pools. The total estimated biomass carbon stock was $2.4^{5.16}_{0.95}$ Pg, which accounted for 23.4% of China's total vegetation carbon stock (figure 6(a)). The estimated biomass C stock size of grassland was 0.86 Pg, which accounted for 63.7% of China's total grassland biomass C size (1.35 Pg from Tang *et al* [34]), this proportion was similar with the findings by Fan *et al* (56.4%) [35], but the amount was much lower than theirs (1.87 Pg). Our results of biomass C densities were higher in shrubland and lower in forest compared with China's average value by Tang *et al* [34] (1.07 vs 0.64 kg m⁻² in shrubland, and 4.2 vs 5.32 kg m⁻² in forest).

The estimated SOC stock of the entire TP was $32.01^{47.9}_{19.69}$ Pg C (1 m depth), with an average density of $11.72^{17.53}_{7.2}$ kg C m⁻², correspondingly accounting for $37.2^{55.6}_{22.9}$ % (25%–75% quantile) of China's SOC stocks

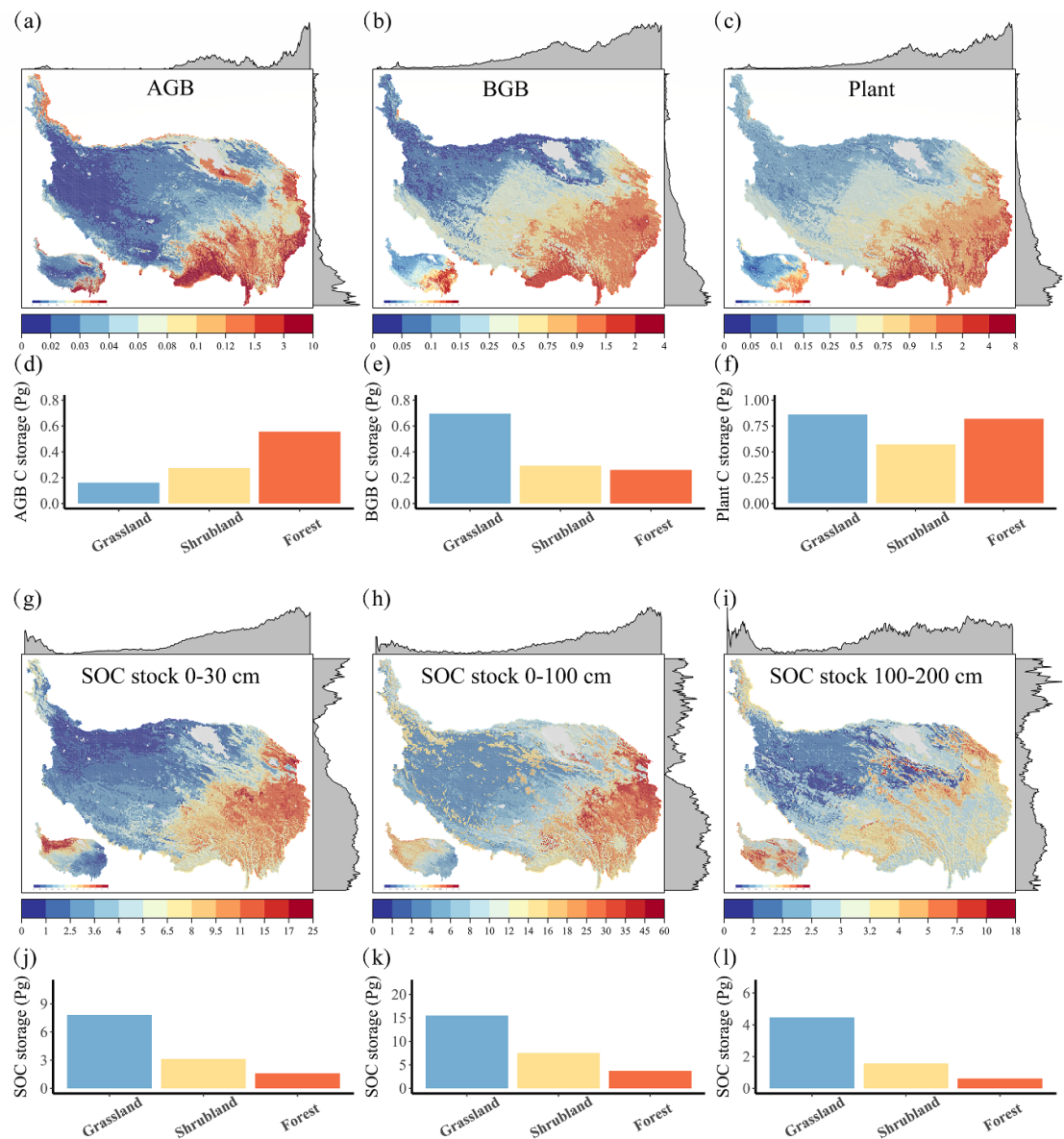


Figure 4. Spatial pattern of different C stocks (kg m^{-2}) on the TP and their C storage size within the three vegetation types. Spatial pattern of (a) AGB C stock, (b) BGB C stock, (c) plant C stock, (g) 0–30 cm SOC, (h) 0–100 cm SOC, (i) 100–200 cm SOC. Total C stocks (Pg) of grassland, shrubland and forest in (d) AGB, (e) BGB, (f) plant, (j) 0–30 cm soil, (k) 0–100 cm soil, (l) 100–200 cm soil.

(median = 86.09 Pg) (figure 6(a)), and $\sim 2.2\%$ of global SOC stock (synthesized ~ 1448.9 Pg) [36, 37]. This size was much similar to Jiang's *et al* [14] report (32.79 Pg with 12.68 kg m^{-2} on average) but lower than the estimates of Fang [13] (38.4 Pg for 0–0.72 m depth) and Tian *et al* [12] (49 Pg for 0–0.75 m depth) (table 1). To compare with the results from earth system models (ESMs), we chose ten ESMs derived from Coupled Model Intercomparison Project v.5 over the TP, and found our estimates (0–100 cm) was similar with the averaged prediction (~ 29.32 Pg C) of the ten ESMs (supplementary figure S6). However, almost all of these ten ESMs failed to capture the spatial pattern of SOC distribution [38], our outputs could be

provided as a benchmark for future carbon model prediction on the TP.

Though the larger proportion of SOC on the TP compare to its area (figure 6(a)), the sensitivity of which to warming such as Q_{10} was found to be $\sim 20.7\%$ larger than the average level of China (2.32–1.93) (figure 6(b)), suggesting its vulnerability in response to climate warming on the TP. Process-based model (ORCHIDEE) has indicated that a 2°C warming could invoking $\sim 10\%$ of SOC stock lost (1.2 Pg) in grassland, which would make the TP a weak C source [39]. While Chen *et al* [40] synthesized 65 publications and found the total amount of SOC on the TP would not change under warming, the

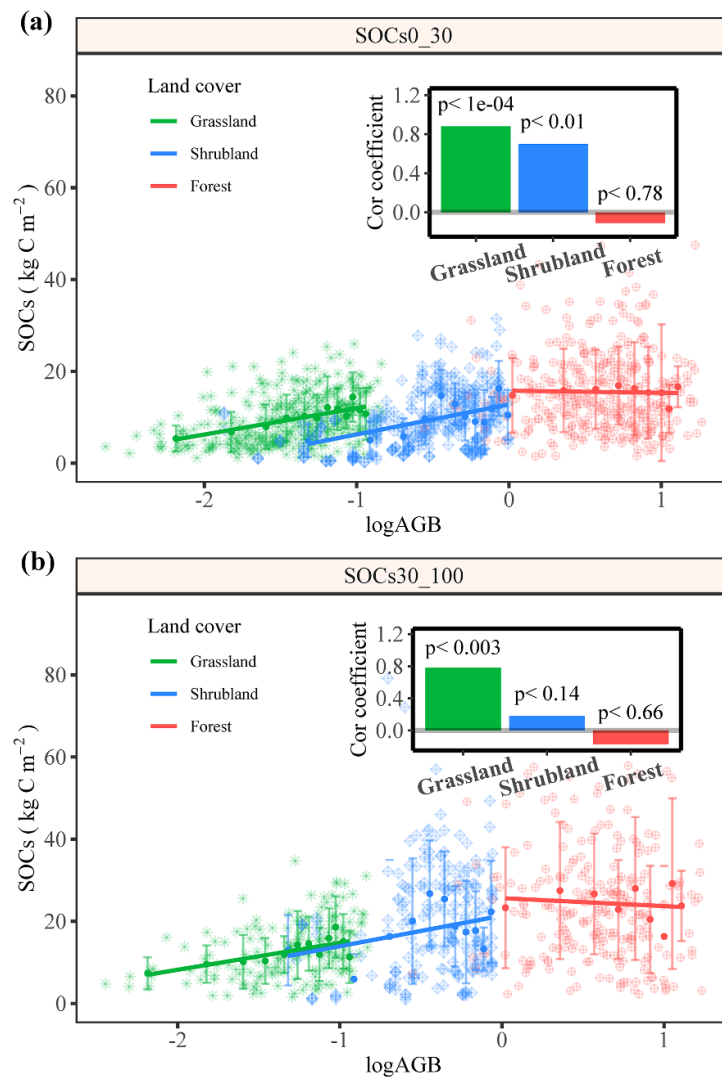


Figure 5. Relationship between AGB carbon stock and SOC stocks in 0–30 cm (a) and 30–100 cm (b). The inserts represent the correlation coefficient of AGB to SOC with depth for the three vegetation types. A base-10 log scale is used for the X-axis of (a) and (b).

lost C induced by enhanced respiration could be compensated by increased plant biomass and dissolved OC. Concerning the uneven distribution of current warming experimental sites on the TP, more attention should be further focused on its central and west part and besides grassland to obtain a more comprehensive assessment of its SOC feedbacks to global warming.

SOC stored in the grassland was $15.52^{23.52}_{9.91}$ Pg, which accounted for 64.5% of China's grassland carbon size (24.03 Pg), indicating the large contribution (~64%) of TP's grassland to China's grassland C stock on ~55% of its area [34]. The large size of SOC reservoir in grassland primarily due to the slow OC decomposition rate by low temperature in most alpine grassland regions [2, 10, 19]. To make a further comparison, SOC storage of alpine grassland was estimated (14.07 Pg, 9.92 kg m^{-2} on average), and the amount was almost two times the size of Yang *et al* [1] results for 0–1 m (7.36 Pg , 6.52 kg m^{-2} on average).

Additionally, we overlay a digital permafrost distribution map on our predicted maps, and found that there were $4.71^{6.87}_{3.07}$, $10.74^{16.06}_{5.83}$, and $3.41^{8.73}_{1.94}$ Pg C stored in 0–0.3 m, 0–1 m and 1–2 m in permafrost regions, respectively, which were all fell into the range of previous estimations, but mainly at a high level for each depth interval (table 1).

Differences between our estimates to previous studies could be mainly induced by three possible reasons. Firstly, the dataset was unprecedentedly improved for both its data volume and geographical coverage, making the data more representative and minimizing the uncertainties [10, 15]. In particular, we have appended a large number of measurements from regional field works conducted in shrub and forest ecosystems, which were rarely concerned by former estimates. Moreover, number of observations from permafrost regions were complemented recent years, which has largely increased the average value of SOC density (SOCD) in grassland than before

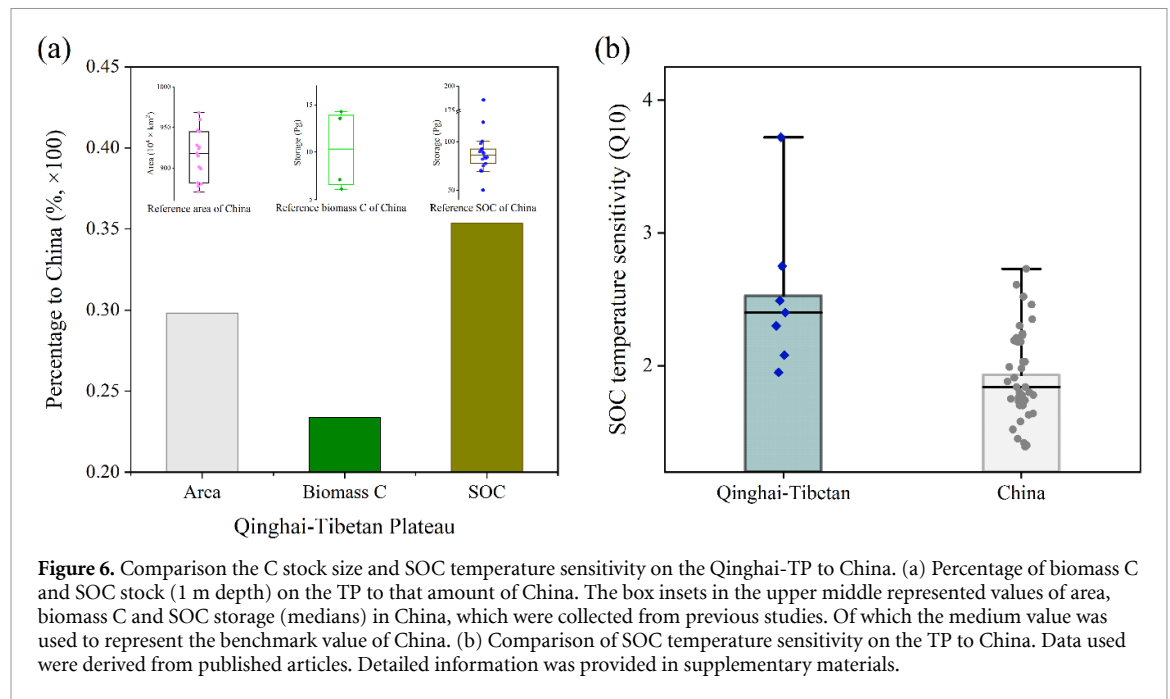


Figure 6. Comparison the C stock size and SOC temperature sensitivity on the Qinghai-TP to China. (a) Percentage of biomass C and SOC stock (1 m depth) on the TP to that amount of China. The box insets in the upper middle represented values of area, biomass C and SOC storage (medians) in China, which were collected from previous studies. Of which the medium value was used to represent the benchmark value of China. (b) Comparison of SOC temperature sensitivity on the TP to China. Data used were derived from published articles. Detailed information was provided in supplementary materials.

(increased to 6.96 and 9.95 kg C m⁻² for 0–0.3 m and 0–1 m, respectively). Secondly, algorithms have a non-negligible effect on regional SOC prediction. The machine learning approach used in this study was expected to have better performance than traditional up-scaling methods [2]. A recent work by Liu *et al* [41] used ensemble machine learning algorithms mapping China's soil information, and achieved more accurate predictions compared with previous soil maps like the harmonized world soil database (HWSD, v1.2), which produced by linking raster to harmonized soil property data. Thirdly, differences in scale may also result in estimation deviations among vegetation types. For instance, area of alpine grassland used in this study was $\sim 1.42 \times 10^6$ km² which was larger than 1.13×10^6 km² that used in Yang *et al* [1].

Though machine learning has proven to have great advantages in OC mapping [2, 17], this study still has uncertainties. For biomass C stocks, the large uncertainty from forest AGB and grassland BGB suggested that additional observations in above ground forest and belowground grassland are needed in the future (figures 4(a)–(c)). For SOC stocks, the uncertainty decreased for 0–100 cm compared with 0–30 cm, which was consistent with the tendency from Ding *et al* [2] and Jiang *et al* [42]. While the uncertainty increased sharply when it concerned 100–200 cm depth, which was in accordance with the findings by Mishra *et al* [10], who claimed the uncertainty was much higher at 100–200 cm depth interval on the TP's permafrost region. These above uncertainties could mainly be resulted from the lack of sampling points, which again suggested the necessity of supplemental samplings for the non-sampling areas and deeper soils.

4.2. Driving factor for different carbon stocks in the TP

Ecological correlation does not necessarily imply causation. Temperature related factors, specifically the mean diurnal temperature range (Tdr), was shown as the primarily factor negatively controlling AGB across the TP (figure 3(a), supplementary figures S4 and S5). Usually, vegetation through physiological processes like evapotranspiration and biomass energy absorption modifying ambient air temperature during daytime, and releasing the absorbed energy to atmosphere during nighttime affecting ambient air temperature, and thus influence the Tdr. Therefore, Tdr, as an indicator, is regulated by interaction among plant and surrounding factors like solar irradiation, VPD etc [43]. On the TP, the spatial pattern of AGB was asynchronous with the Tdr, which increased from forest in southwest to barren desert steppe in west central part (figure 4(a)). From this point of view, the dominant driving factor for AGB should be water related factors on the TP (supplementary figure S4(a)). However, our result implied that Tdr can be used as an effective covariate for AGB prediction on the TP.

Our results suggested that AGB and water related variables presented as the key categories for the topsoil (0–30 cm) SOC stock, which was consistent with former studies [1, 44, 45]. SOC of topsoil on the TP was susceptible to precipitation because it could increase organic C input into the soil by enhancing plant productivity. For subsoil (30–100 cm), we found AGB and climatic variables still played as a dominant role in driving its SOC, though the accumulative contribution of soil physico-chemical properties increased (figure 3(c)). This result appears to be inconsistent with the early findings on a global

Table 1. Comparison the estimates of SOC stocks on the TP.

Depth (m)	Study area	N	Area (10 ⁶ km ²)	Methods	SOD (kg m ²)	SOCS (Pg)	Reference
0–0.3	Grassland	135	112.82	Linear regression with satellite observations	4.42	4.99	[1]
	Grassland	135	112.82	Linear relationship with satellite observations	3.9	4.4	[32]
	Permafrost (mainly grassland)	103	114.4	Mean value with area	4.63	—	[9]
	Permafrost (mainly grassland)	200	148	Mean value with area	—	6.05 (5.74–6.39)	[16]
0–1	Permafrost (mainly grassland)	173	110	Regression-kriging interpolation	6 (3.5–9)	7 (4–10)	[10]
	Grassland	2019	~157	Machine learning—RF	5.0	7.8	This study
	Grassland	135	112.82	Linear regression with satellite observations	6.52	7.36	[1]
	Permafrost (mainly grassland)	173	114.4	Machine learning—SVM	7.44 (6.38–8.68)	8.51 (7.3–9.93)	[2]
	Permafrost (mainly grassland)	200	148	Mean value with area	—	12.73	[16]
	Permafrost (mainly grassland)	190	135	Mean value with area	—	17.3 ± 5.3	[33]
	Permafrost (mainly grassland)	114	110	Regression-kriging interpolation	8 (6–9.5)	9.2 (7–11)	[10]
	Frozen ground	1302	158.8	Upscaling with soil subgroups	11.21	17.79	[14]
	Grassland	1302	~157	Machine learning—RF	9.95	15.52	This study
	Permafrost (mainly grassland)	114	~117	Machine learning—RF	9.25	10.74	This study
1–2	Permafrost (mainly grassland)	200	148	Regression-kriging interpolation	2 (0.3–5)	2.5 (0.4–6)	[10]
	Permafrost (mainly grassland)	200	148	Upscaling with land covers	—	3.57 (1.3–6.66)	[16]
	Permafrost (mainly grassland)	190	135	Mean value with area	—	10.6 ± 2.7	[33]
	Permafrost	307	117	Machine learning—RF	2.95	3.41	This study

(Continued.)

Table 1. (Continued.)

Depth (m)	Study area	N	Area (10^6 km^2)	Methods	SOC D (kg m^{-2})	SOCS (Pg)	Reference
0–2	Permafrost (mainly grassland)	173	114.4	Machine learning—SVM	10.68 (9.1–12.51)	12.22 (10.41–14.31)	[2]
	Permafrost (mainly grassland)	200	148	Upscaling with land covers	—	17.07 (11.34–25.33)	[16]
	Permafrost (affected region)	—	—	Review	—	19 \pm 6.6	[6]
	Permafrost (affected region)	190	135	Mean value with area	—	27.9	[33]
0–3	Permafrost	307	~117	Machine learning—RF	6.09	13.88	This study
	Permafrost	1114	—	Machine learning—SVM	15.4 (14.4–16.4)	36.6 (34.3–38.9)	[15]
Actual depth	Permafrost	314	—	Machine learnings—RF	6.14	15.33	[17]
	Frozen ground	—	—	Upscaling with soil subgroups	25.76	40.89	[14]
	Entire TP	the 2nd National Soil Survey	158.8 255.1	Upscaling with soil subgroups	7.2	18.37	[12]
	Entire TP	the 2nd National Soil Survey	197	Upscaling with soil subgroups	—	38.4	[13]
0–0.75	Entire TP	124	162.7	Mean value with area	—	49.0	[12]
0–1	Entire TP	144 subgroups	258.7	Upscaling with soil subgroups	12.68	32.79	[14]
0–3	Entire TP	144 subgroups	258.7	Upscaling with soil subgroups	28.45	73.6 (63.9–83.3)	[14]
0–1	Entire TP	7196	273	Machine learning—RF	11.72	32.01	This study
1–2	Entire TP	307	273	Machine learning—RF	2.96	8.1	This study

scale (with ~2700 soil profile) [19] which claimed a switched importance of SOC driving factors with depth, and emphasized the importance of edaphic factors for deep soils. However, Luo *et al* [46] recently used an updated global soil dataset (141 581 soil profile) detecting the controlling factors and found the importance of climatic variables did not show a decreasing trend with depth. Additionally, Mishra *et al* [10] used the regression kriging method (AGB was not included) also suggested the importance of climate to deep SOC which found precipitation could explain the largest portion of the SOC variation for both 0–0.3 m (47%) and 0–1 m (24%) SOC on the TP. It is important to note that the relative high SOC_D the TP was mainly attributed to its cold weather. As the C output pathway was inhibited by low temperature, the mass C balance should be dominated by C input from plant which was primarily controlled by climatic variables (figure 3(a), supplementary figure S4(a)). In addition, from the data aspect, the topsoil (0–30 cm) and subsoil (30–100 cm) showed a high correlation (figure S7) and also a similar spatial pattern (figures 4(g)–(h)), therefore our findings, these two layers were synchronized on the main driving factor, should be reasonable. However, C stability in subsoils might still be controlled by edaphic related factors rather than climatic variables. A recent study used radiocarbon ($\Delta^{14}\text{C}$) measurements detecting controlling factors on C persistence on the TP's grassland, and found the dominant factor for SOC persistence changed to mineral properties in subsoil due to a decreased root C input [45]. While evidence from temperature control experiments have also demonstrated that deep SOC had higher sensitivity (Q_{10}) to warming than topsoils [47, 48], and was vulnerable to climate warming [49]. Given the large uncertainty and vulnerability of SOC to climate, more field observations and manipulative experiments of whole-soil profile should be paid to clarify the feedbacks of deep soils to the ongoing climate change.

For vegetation types, the dominant driving factors of topsoil SOC in grassland were coming from water related categories (supplementary table S5), which was in accordance with former studies [1, 10]. While for shrubland, the key driving factors transformed into temperature related variables. The significant negative relationship between temperature related variables (supplementary table S5) and SOC stock in shrubland in this study was consistent with the findings from shrubland on a global scale [19, 50], which indicated an increasing risk for C loss in shrubland due to warming stimulated biological degradation. The reason might be that alpine shrubs are mainly distributed on shady slopes and foothills with lower solar radiation and thus low temperature and water consumption. Therefore, water becomes no longer a limit factor and the temperature becomes the main constraint of the shrub ecosystems.

4.3. Relationship between different C stocks

The tendency between plant productivity and SOC varies with vegetation types. The correlation coefficients decreased from grassland to shrubland and disappeared in forest (figures 5(a) and (b)). Plant productivity (i.e. AGB) in grassland was mainly controlled by water related variables (figure 3(a)), and the low temperature further constrains the microbe activities for SOM decomposition. Therefore, plant productivity was closely related to its SOC stocks in grassland, which has also been demonstrated by remote sensing observations [1, 18]. Whereas, for shrubland, temperature related variables was the main driver on its topsoil SOC which mainly affect the C loss path [50]. Therefore, the relationship between AGB, which determine C input, and SOC was weakened. For forest, only a small portion of its AGB, mainly leaves, fall on the ground and has been involved in the C cycle every year, resulting in no significant relationship between them.

Besides, the strength of the correlations between AGB and SOC weakened with depth. For subsoil (30–100 cm), only grassland exhibited a significant ($p < 0.01$) relationship between AGB and SOC, with a decreased R^2 (figure 5(b)). Which confirmed by previous findings that the importance of C input from plant decrease with increased soil depth [19, 45]. However, AGB still played a dominant role in shaping subsoil SOC distribution, indicating that the importance of plant C input for deep SOC (to 1 m) was non-negligible on the TP. The similar contribution of other environmental categories (i.e. temperature and soil properties) suggested a relative complex regulatory environment for the deep soils [51]. In sum, our findings implicated that the responses of SOC to future climate change may differs depending on interactions between vegetation types and regional climate conditions.

Data availability statement

The data that support the findings of this study are available upon reasonable request from the authors.

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Conflict of interest

None declared.

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References

- [1] Yang Y *et al* 2008 Storage, patterns and controls of soil organic carbon in the Tibetan grasslands *Glob. Change Biol.* **14** 1592–9
- [2] Ding J *et al* 2016 The permafrost carbon inventory on the Tibetan Plateau: a new evaluation using deep sediment cores *Glob. Change Biol.* **22** 2688–701
- [3] Pan B and Li J 1996 Qinghai-Tibetan Plateau: a driver and amplifier of the global climatic change III. The effects of the uplift of Qinghai-Tibetan Plateau on climatic changes *J. Lanzhou Univ.* **1** 108–15
- [4] Chen H *et al* 2013 The impacts of climate change and human activities on biogeochemical cycles on the Qinghai-Tibetan Plateau *Glob. Change Biol.* **19** 2940–55
- [5] Dörfer C, Kühn P, Baumann F, He J-S and Scholten T 2013 Soil organic carbon pools and stocks in permafrost-affected soils on the Tibetan Plateau *PLoS One* **8** e57024
- [6] Mu C *et al* 2020 The status and stability of permafrost carbon on the Tibetan Plateau *Earth Sci. Rev.* **211** 103433
- [7] Mao J *et al* 2016 Human-induced greening of the northern extratropical land surface *Nat. Clim. Change* **6** 959–63
- [8] Piao S *et al* 2019 Responses and feedback of the Tibetan Plateau's alpine ecosystem to climate change *Chin. Sci. Bull.* **64** 2842–55
- [9] Ding J *et al* 2017 Decadal soil carbon accumulation across Tibetan permafrost regions *Nat. Geosci.* **10** 420–4
- [10] Mishra U *et al* 2021 Spatial heterogeneity and environmental predictors of permafrost region soil organic carbon stocks *Sci. Adv.* **7** eaaz5236
- [11] Wang G, Qian J, Cheng G and Lai Y 2002 Soil organic carbon pool of grassland soils on the Qinghai-Tibetan Plateau and its global implication *Sci. Total Environ.* **291** 207–17
- [12] Tian Y, Ouyang H, Xu X, Song M and Zhou C 2008 Distribution characteristics of soil organic carbon storage and density on the Qinghai-Tibetan Plateau *Acta Pedol. Sin.* **45** 933–42
- [13] Fang J 1996 Soil carbon pool in China and its global significance *J. Environ. Sci.* **8** 249–54
- [14] Jiang L, Chen H, Zhu Q, Yang Y, Li M, Peng C, Zhu D and He Y 2019 Assessment of frozen ground organic carbon pool on the Qinghai-Tibet Plateau *J. Soils Sediments* **19** 128–39
- [15] Ding J *et al* 2019 The paleoclimatic footprint in the soil carbon stock of the Tibetan permafrost region *Nat. Commun.* **10** 4195
- [16] Zhao L *et al* 2018 Soil organic carbon and total nitrogen pools in permafrost zones of the Qinghai-Tibetan Plateau *Sci. Rep.* **8** 3656
- [17] Wang T *et al* 2020 Permafrost thawing puts the frozen carbon at risk over the Tibetan Plateau *Sci. Adv.* **6** 1–8
- [18] Yang Y *et al* 2010 Soil carbon stock and its changes in northern China's grasslands from 1980s to 2000s: SOC dynamics in northern China's grasslands *Glob. Change Biol.* **16** 3036–47
- [19] Jobbágy E G and Jackson R B 2000 The vertical distribution of soil organic carbon and its relation to climate and vegetation *Ecol. Appl.* **10** 423–36
- [20] Don A, Schumacher J and Freibauer A 2011 Impact of tropical land-use change on soil organic carbon stocks—a meta-analysis: soil organic carbon and land-use change *Glob. Change Biol.* **17** 1658–70
- [21] Wiesmeier M *et al* 2014 Carbon sequestration potential of soils in southeast Germany derived from stable soil organic carbon saturation *Glob. Change Biol.* **20** 653–65
- [22] Yang Y, Ji C, Chen L, Ding J, Cheng X and Robinson D 2015 Edaphic rather than climatic controls over ¹³C enrichment between soil and vegetation in alpine grasslands on the Tibetan Plateau *Funct. Ecol.* **29** 839–48
- [23] Rasmussen C *et al* 2018 Beyond clay: towards an improved set of variables for predicting soil organic matter content *Biogeochemistry* **137** 297–306
- [24] Zhang Y, Li B and Zheng D 2002 A discussion on the boundary and area of the Tibetan Plateau in China *Geogr. Res.* **21** 1–8
- [25] Li W and Zhou X 1998 *Ecosystems of Qinghai-Xizang (Tibetan) Plateau and Approach for Their Sustainable Management* (Guangzhou: Guangdong Science & Technology Press)
- [26] Haaf D, Six J and Doetterl S 2021 Global patterns of geo-ecological controls on the response of soil respiration to warming *Nat. Clim. Change* **11** 623–7
- [27] Rose S 2013 Mortality risk score prediction in an elderly population using machine learning *Am. J. Epidemiol.* **177** 443–52
- [28] Mishra U, Drewniak B, Jastrow J D, Matamala R M and Vitharana U W A 2017 Spatial representation of organic carbon and active-layer thickness of high latitude soils in CMIP5 earth system models *Geoderma* **300** 55–63
- [29] Yigini Y, Olmedo G F, Reiter S, Baritz R, Viatkin K and Vargas R 2018 Soil organic carbon mapping: cookbook (Food and Agriculture Organization)
- [30] Chakravarti I M, Laha R G and Roy J 1967 *Handbook of Methods of Applied Statistics* (New York: Wiley)
- [31] R Core Team 2020 *R: A Language and Environment for Statistical Computing* (Vienna: R Foundation for Statistical Computing)
- [32] Yang Y *et al* 2009 Changes in topsoil carbon stock in the Tibetan grasslands between the 1980s and 2004: SOC changes in Tibetan grasslands *Glob. Change Biol.* **15** 2723–9
- [33] Mu C *et al* 2015 Editorial: organic carbon pools in permafrost regions on the Qinghai-Xizang (Tibetan) Plateau *Cryosphere* **9** 479–86
- [34] Tang X *et al* 2018 Carbon pools in China's terrestrial ecosystems: new estimates based on an intensive field survey *Proc. Natl Acad. Sci.* **115** 4021–6
- [35] Fan J *et al* 2008 Carbon storage in the grasslands of China based on field measurements of above- and below-ground biomass *Clim. Change* **86** 375–96
- [36] Koechy M, Hiederer R and Freibauer A 2015 Global distribution of soil organic carbon—part 1: masses and frequency distributions of SOC stocks for the tropics, permafrost regions, wetlands, and the world *Soil* **1** 351–65
- [37] Scharlemann J P, Tanner E V, Hiederer R and Kapos V 2014 Global soil carbon: understanding and managing the largest terrestrial carbon pool *Carbon Manage.* **5** 81–91
- [38] Li S, Liu Y, Lyu S, Wang S, Pan Y and Qin Y 2021 Change in soil organic carbon and its climate drivers over the Tibetan Plateau in CMIP5 earth system models *Theor. Appl. Climatol.* **145** 187–96
- [39] Tan K *et al* 2010 Application of the ORCHIDEE global vegetation model to evaluate biomass and soil carbon stocks of Qinghai-Tibetan grasslands *Glob. Biogeochem. Cycles* **24** 1–12
- [40] Chen Y, Feng J, Yuan X and Zhu B 2020 Effects of warming on carbon and nitrogen cycling in alpine grassland ecosystems on the Tibetan Plateau: a meta-analysis *Geoderma* **370** 114363
- [41] Liu F *et al* 2021 Mapping high resolution National Soil Information Grids of China *Sci. Bull.* **67** 328–40
- [42] Jiang H *et al* 2020 Progress and challenges in studying regional permafrost in the Tibetan Plateau using satellite remote sensing and models *Front. Earth Sci.* **8** 1–17

- [43] Meier R, Davin E L, Swenson S C, Lawrence D M and Schwaab J 2019 Biomass heat storage dampens diurnal temperature variations in forests *Environ. Res. Lett.* **14** 084026
- [44] Del Grosso J S, Halvorson A D and Parton W J 2008 Testing DAYCENT model simulations of corn yields and nitrous oxide emissions in irrigated tillage systems in Colorado *J. Environ. Qual.* **37** 1383–9
- [45] Chen L *et al* 2021 Soil carbon persistence governed by plant input and mineral protection at regional and global scales *Ecol. Lett.* **24** 1018–28
- [46] Luo Z, Viscarra-Rossel R A and Qian T 2021 Similar importance of edaphic and climatic factors for controlling soil organic carbon stocks of the world *Biogeosciences* **18** 2063–73
- [47] Xu M, Li X, Kuyper T W, Xu M, Li X and Zhang J 2021 High microbial diversity stabilizes the responses of soil organic carbon decomposition to warming in the subsoil on the Tibetan Plateau *Glob. Change Biol.* **27** 2061–75
- [48] Li J *et al* 2020 Rising temperature may trigger deep soil carbon loss across forest ecosystems *Adv. Sci.* **7** 1–8
- [49] Hicks Pries C E, Castanha C, Porras R C and Torn M S 2017 The whole-soil carbon flux in response to warming *Science* **355** 1420–3
- [50] Schimel D S *et al* 1994 Climatic, edaphic, and biotic controls over storage and turnover of carbon in soils *Glob. Biogeochem. Cycles* **8** 279–93
- [51] Gray J M, Bishop T F A and Wilson B R 2015 factors controlling soil organic carbon stocks with depth in Eastern Australia *Soil Sci. Soc. Am. J.* **79** 1741–51