

1 **Physicochemical properties of clayey deposits slurry treated by**
2 **lime-activated ISSA and GGBS considering seawater salinity**
3 **effect**

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25 **Abstract**

26 This study sheds light on the engineering and environmental performance of lime-activated incinerated
27 sewage sludge ash (ISSA) and ground granulated blast furnace slag (GGBS) treated Hong Kong marine
28 deposits (HKMD) slurry by stabilisation/solidification (S/S) technology, which is proposed using as fill
29 material in reclamation projects. The S/S performance of the treated HKMD with distilled water and
30 seawater under different salinities was investigated. The results show that seawater could help S/S
31 treated HKMD gain strength by using activated industrial wastes (ISSA and GGBS). The hydration and
32 pozzolanic reactions between ISSA, GGBS, CaO and clayey compositions in HKMD make
33 contributions to the strength development, porosity decrease and heavy metals stabilization, which is
34 supported by the characterization analysis including thermo-gravimetric (TG) analysis, mercury
35 intrusion porosimetry (MIP) tests, nitrogen adsorption/desorption isotherms (NAI), scanning electron
36 microscopy coupled with energy-dispersive spectrometry (SEM-EDS) and the leaching test of toxicity
37 characteristic leaching procedure (TCLP). Seawater of 1.8% salinity (18 g/kg) is better than the distilled
38 water and seawater of 3.6% salinity as a substrate solution in the S/S treated HKMD, because of the
39 highest unconfined compressive strength and lowest porosity in the treated samples. The highest pH
40 may account for its highest strength under the 1.8% salinity conditions. The S/S process could
41 effectively stabilize the contaminants regardless of the curing time and the salinity of the mixing solution,
42 and the leachates from the stabilized HKMD are environmentally safe and meet the requirement of
43 standard in Hong Kong on the recycling treated soil. Therefore, recycling wastes—ISSA and GGBS
44 with lime can be used as an appealing binder to stabilise/solidify marine deposits as environmental-
45 friendly reusable materials in reclamation projects.

46 **Keywords:** Stabilisation/solidification, Hong Kong marine deposits, Industrial wastes reuse, Salinity
47 of seawater, Metal leachability

48

49 **1 Introduction**

50 Soft soils have been used as fill material in land reclamation projects in mainland China and Singapore.
51 However, most of the newly blow-filled soft soil are clayey soils possessing high water content, low
52 permeability and low strength. In Hong Kong, dredged marine sediments are produced 29 Mm³ per year
53 and approximately 1200 tonnes per day of dewatered sewage sludge are produced converting into
54 incinerated sewage sludge ash (ISSA) after incineration (Zhou et al., 2021a). It would take up many
55 landfills to dispose of both dredged marine sediments and ISSA resulting in an urgent issue to identify
56 suitable disposal sites. The current study sheds light on possible beneficial use of industrial wastes--
57 ISSA and ground granulated blast furnace slag (GGBS) to treat Hong Kong marine deposits (HKMD)
58 slurry to produce fill material in reclamation project. In order to achieve this goal, there are two basic
59 requirements: 1) the treated HKMD using ISSA and GGBS gains competent bearing capacity; 2) the
60 leachability of contaminants from the treated HKMD meets the reuse standards of treated soil, which is
61 more stringent than that in standard of landfill soil (HK EPD 2011). Since industrial wastes (e.g., ISSA)
62 contain a variety of heavy metals, and many types of dredged sediments are classified as unacceptably
63 contaminated in Hong Kong (Wang et al., 2015). Thus, the potential environmental risks exist. While
64 stabilisation/solidification (S/S) technology can meet both of the requirements in improving the physical
65 and mechanical properties of treated soft soil (Lofrano et al., 2017; Li et al., 2019) and reducing the
66 heavy metals dissolution, migration and leaching toxicity in soil environment (Malviya and Chaudhary,
67 2006).

68
69 In recent years, the development of new sustainable S/S materials is sought and of global interest for the
70 immobilization of heavy metals in contaminated soils (Emmanuel et al., 2020). In this work, the lime-
71 activated ISSA and GGBS were adopted as the binders in the S/S treatment of HKMD with seawater as
72 matrix solution. It makes more sense to use activated industrial wastes as an alternative binder material
73 to cement from both engineering and environmental point of view. Numerous previous works have
74 studied that lime can increase soil pH, promote precipitation of metal oxides, hydroxides or carbonates
75 and decrease heavy metal solubility (Castaldi et al., 2005; Chen et al., 2021; Gong et al., 2021; Xu et
76 al., 2021). Additionally, a large amount of heavy metals would have a negative impact on the mechanical
77 properties of geomaterials (Zhang et al., 2022a). It is found that high pH in the system is conducive to

78 immobilizing toxic metals by adsorption, precipitation and ion exchange reactions (Malviya and
79 Chaudhary, 2006). ISSA as a stabilisation agent was found effectively in S/S soft soil treatment (Chen
80 and Lin, 2009; Li et al., 2019; Zhou et al., 2020; Jagaba et al., 2019). Zhou et al. (2020) found that a
81 moderate pozzolanic activity in the ISSA - soft soil system. Another industrial waste-GGBS can help
82 the soft soil to improve the strength (Zentar et al., 2012; Celik and Nalbantoglu, 2013; Cai et al. 2022;
83 Yi et al., 2015; Song et al., 2023). Furthermore, the stabilizing effect of activated GGBS as a binder on
84 heavy-metal contaminated soils has been confirmed (Jin et al., 2016; Wu et al., 2018). Sun et al. (2023a)
85 investigated the treated HKMD slurry by the lime, activated ISSA and GGBS with dosage of 30% by
86 dry weight of HKMD. It was found that the hydration products such as ettringite (AFt), Friedel's salt
87 (F's salt) and calcium/magnesium silicate hydrate (C/M-S-H) and fill into the pores of the treated soil
88 and interlock the particles in the mixture, which increases the soil strength. Additionally, seawater was
89 reported playing a positive role as matrix solution in the S/S treatment of HKMD (Sun et al., 2023b),
90 which is in agreement with the findings from Ashraf et al. (2022) that due to the alkalinity (pH 8.2) of
91 seawater, it could act as an activator for the calcined clay-portlandite mixtures.

92
93 Previous research has confirmed that the factors, such as the binder material (Zhou et al., 2022), the
94 binder content (Deng et al., 2015; Li and Yi, 2023), the water content (Lang et al., 2021), the soil type
95 (Deng et al., 2015; Zhou et al., 2022), and chemical compositions of mixing solution (Zhao et al., 2021;
96 Sun et al., 2023b), etc., have impacts on the property of treated soil. Recently, studies have shown that
97 the change of pore water salinity can also change the inherent properties of reconstituted clay (Song et
98 al. 2017; Deng et al. 2018; Sun et al., 2020); but, the change of physiochemical properties of stabilized
99 marine clay due to seawater salinity has rarely been evaluated (Bayesteh and Bayat, 2021). Some
100 researchers investigated on the effect of salinity on the engineering properties of cement mixing soils or
101 dredged materials (e.g., Horpibulsuk et al. 2012; Khoshsirrat et al., 2019; Monsif et al., 2020).
102 Horpibulsuk et al. (2012) studied the unconfined compressive strength (UCS) of cement stabilization of
103 soft saline clays and found that the presence of salt (95.5% of NaCl) not only reduces the attraction and
104 bonding strength between clay particles, but also adversely affects the hydration of cement. Zhang et al.
105 (2014) indicated that a high salinity of NaCl has a detrimental effect on the UCS of cement stabilized
106 clay. The findings in Monsif et al. (2020) presented the impact of salinity (by mixing NaCl) on the

107 strength of cement-treated Champlain Sea clay and found that the UCS of treated soil shows no obvious
108 rule with the salinity increasing. Khoshsir et al. (2019) pointed out that seawater salinity could control
109 the UCS of cement-stabilized marine clay by affecting the microstructural configuration of the clay and
110 the pozzolanic reaction. In addition, the UCS of treated soil also presents a trend of increasing first and
111 then decreasing while the salinity increasing from low to high level. Zhou et al. (2021b) observed that
112 the strength of the lime-ISSA paste with seawater was double higher than that of paste with freshwater
113 at 28 curing days. However, these studies only examined the salinity of NaCl on the cement treated soil
114 properties or focused on the phenomenon of cement treated soil strength affected by the seawater salinity.
115 Sun et al. (2023b) attempted to analyse the effect of seawater components on the S/S mechanism of
116 treated HKMD by lime-activated ISSA-GGBS. Nevertheless, the effect of seawater salinity on the S/S
117 treated HKMD was not taken into account.

118 Thus, there is no conclusive finding about the effect of seawater salinity on the S/S performance of
119 treated marine clay treated by materials other than cement. This study aims to evaluate the effect of
120 seawater salinity on the stabilisation/solidification performance of HKMD by lime-activated industrial
121 wastes from the aspects of engineering strength and environmental risk assessment. The objectives of
122 this study are to reveal: a) the effect of seawater with different levels of salinity on the physical and
123 mechanical property of the S/S HKMD; b) the microstructural property of the S/S HKMD affected by
124 seawater salinity; c) the morphology of the neoformations and heavy metals in the S/S HKMD; d) the
125 leaching behavior of heavy metal(loid)s including arsenic (As), cadmium (Cd), chromium (Cr), copper
126 (Cu), zinc (Zn), nickel (Ni) and lead (Pb) from HKMD before and after S/S treatment. According to
127 Huang et al. (2019), heavy metals such as As, Cu, Zn, Ni, Cd, Cr, Pb, etc., are enrich in Chinese
128 contaminated soils. In particular, Cd is categorized as the most hazardous metal by the United States
129 Environmental Protection Agency (Gong et al., 2021). Macroscopic tests were carried out to investigate
130 the macro- performance of S/S treated HKMD. Thermo-gravimetric (TG) analysis, mercury intrusion
131 porosimetry technique (MIP) tests, nitrogen adsorption/desorption isotherms (NAI) and scanning
132 electron microscopy coupled with energy-dispersive spectrometry (SEM-EDS) were performed to
133 analyse the S/S mechanism of the treated HKMD. Besides, the toxicity characteristic leaching method
134 (TCLP) was used for the first time to evaluate the environmental risk of HKMD after S/S treatment.

135 2 Materials and Methodologies

136 2.1 Hong Kong marine deposits and binder materials

137 The dredging site for deposits collecting is near lamma island in Hong Kong. The basic parameters of
138 dredged HKMD are listed in Table 1. Among them, the BET SSA is the specific surface area (SSA)
139 calculated by Brunauer-Emmett-Teller (BET) method. The main compositions of the HKMD include
140 kaolinite, muscovite, quartz, halite, albite and phillipsite with the specific content of 14.4%, 11%, 55%,
141 8.3%, 4%, 3.7%, respectively.

142 Table 1. Basic properties of Hong Kong marine deposits.

G _s	Plastic limit (%)	Liquid limit (%)	PI	BET SSA (m ² /g)	pH
2.525	33	50	17	12.94	7.54

143

144 The oxide compositions of ISSA, GGBS and lime detected by X-ray fluorescence (XRF, Rigaku
145 Supermini 200) was described in Sun et al. (2023b). The particle size of HKMD, ISSA and GGBS was
146 presented in Sun et al. (2023a). Seven heavy metals from the materials including As, Cd, Cr, Cu, Zn, Ni
147 and Pb were investigated. As shown in Table 2, the leachability of these heavy metals from the raw
148 HKMD, ISSA and GGBS does not exceed the Landfill Disposal Criteria for Contaminated Soil standard
149 in Hong Kong (HK EPD 2011). While the theoretical value of the heavy metals in the treated soil as-
150 mixed does not meet the On-site Reuse of Cement Stabilisation/Solidification Treated Soil standard (HK
151 EPD 2011), especially for Cd, Cr, Pb and Zn. Thus, the environmental risk should be studied before the
152 direct reuse of the treated HKMD.

153

154 Table 2. Heavy metals and their concentrations from marine deposits and the industrial wastes.

Materials	Heavy metal (loid)s (mg/kg)						
	As	Cd	Cr	Ni	Cu	Pb	Zn
HKMD	0.175	0.222	1.98	3.01	0.426	1.272	3.95
ISSA	5.9	0.84	2.02	2.38	25.56	0.78	23.94
GGBS	0.134	0.17	1.87	0.287	0.418	1.04	0.56

Raw mixture [#]	1.06	0.30	1.80	2.50	4.35	1.07	6.56
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155 [#]The value of the heavy metal (loid)s in the mixture just mixed: $C^{\#} = C_{HKMD} \times 70\% + (C_{ISSA} \times 75\% + C_{GGBS} \times 25\%)$
156 $\times 70\% \times 30\%$.

159 2.2 Testing methods

160 2.2.1 Sample preparation

161 Firstly, the dried ISSA and GGBS with a mass ratio of 3:1 was mixed. Then, the lime and ISSA-GGBS
162 were mixed in a ratio of 3:7 to form the binder. Before sample preparation, the HKMD paste was
163 adjusted to 110% water content with distilled water (DW) and seawater (SW) with salinity of 18 and 36
164 g/kg (1.8% and 3.6%). The compositions of seawater are shown in Table 3. The mixed binders were
165 adopted with a dosage of 30% HKMD in dry weight. The cylindrical moulds of 100 mm in height and
166 50 mm in diameter were used to place the treated mud. The moulds were finally stored in a chamber
167 with humidity of around 95 % and temperature at 20 ± 2 °C. The specific sample preparation process is
168 shown in Fig. 1.

169



170

171 Fig. 1. Schematic diagram of sample preparation (Sun et al., 2023a).

172

173 Table 3. Compositions (in mg/L) of Red Sea salts used in this work

Compound	Cl	Na	Mg	S	Ca	K	C (inorg.)	Br
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Concentration	18,772	10,691	1320	890	480	390	45	15
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175 *2.2.2 Property tests*

176 The treated samples were dried in an oven at 105 °C for 24 h to determine their water contents. The
 177 specific gravity of materials was measured by a BELPycno L Gas Pycnometer (Microtrac MRB, Japan).
 178 The wax sealing method was used to measure the bulk density of S/S samples. All these results used for
 179 data analysis were the average of triplicates. The unconfined compression tests were conducted on the
 180 S/S HKMD samples by using a VJ-Tech Tri-Scan 50 compression machine.

181

182 *2.2.3 Characterization tests*

183 After curing, the specimens would be cut and frozen-dried for 48 h in a freeze-drier (Labconco Freezone
 184 2.5 Plus). TG analysis (Netzsch STA 449F3) was carried out on S/S HKMD powder with the weight
 185 around 10 mg after the frozen-dried process. The temperature was increased from 40 to 900 °C at a rate
 186 of 10 °C/min in the TG tests. To prevent carbonation of S/S HKMD samples, the nitrogen gas was used
 187 with a flow rate of 50 mL/min during the heating. MIP tests were carried out on Autopore V 9600 from
 188 Micromeritics to measure the porosity and pore size distribution (PSD) of S/S treated HKMD. NAI
 189 method was adopted by using a TriStar II 3020 porosity and surface area analyzer (Micromeritics
 190 Instruments Corporation) to detect the treated material porosity in powder state. The morphology and
 191 elemental distribution of the S/S treated HKMD samples were observed by SEM with EDS on TESCAN
 192 VEGA 3 in the secondary electron (SE) mode.

193

194 *2.2.4 Metal(loid)s leaching tests*

195 TCLP test was adopted to determine the leachability of heavy metal(loid)s in the materials before and
 196 after the S/S treatment according to the EPA Method 1311 (US EPA 1992). The leachates were subjected
 197 to acid digestion by concentrated HNO₃ (68%). Then, the heavy metal concentrations of the leachates
 198 were measured by Inductively Coupled Plasma-Atomic Emission Spectroscopy (ICP-AES, PerkinElmer
 199 Optima 3300DV). Two or three measurements were conducted for each sample, and the average value
 200 was used in the analysis.

201

202 **3 Results and analysis**

203 **3.1 Macro- property of treated material**

204 **3.1.1 Physical properties**

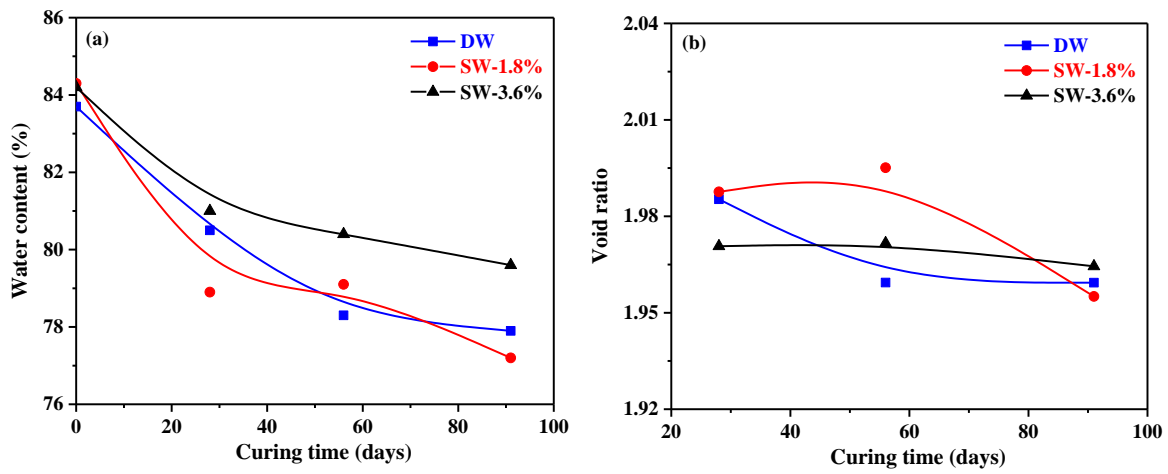
205 The water content and void ratio of the treated HKMD samples with curing times are presented in Fig.
206 2. The true density and bulk density of treated HKMD are listed in Table 4. From Fig. 2(a), it is seen
207 that the water content of the freshly mixed paste is around 84%. The reduction of water content with
208 time is found in the treated samples. From the slope of the curves, it is obvious that the decrease of water
209 content at the period of 0-28 days is most intense, which indicates the faster reaction speed in the system
210 within this period. That is the S/S HKMD mixed with SW of 1.8% salinity have the highest reaction rate
211 during the curing process. The external water content decrease should be due to the water consumption
212 in the hydration and pozzolanic reactions generating new hydration products in the system, which is in
213 agreement with the findings in Cai et al. (2022). Thus, the lowest water content in the mixture with
214 seawater of 1.8% salinity suggests a high content of hydration products and a high content of absorbed
215 water as well (Li et al., 2022).

216 The void ratio (e) of the S/S treated HKMD specimens at different curing time is drawn in Fig. 2 (b). It
217 was calculated from the equation $e = G_s(1 + w)\rho_w/\rho - 1$, in which G_s is the specific gravity with its value
218 taken as true density in this work, w is the water content, ρ_w is the water density, and ρ is the bulk density
219 of the S/S HKMD specimens. It is found that the calculated void ratio lies between 1.95 and 2 and
220 decreases with the curing time generally. Therefore, it can be inferred that the voids in the bulk sample
221 were gradually filled by hydration products (Li et al., 2022).

222

223 The true density and bulk density of treated HKMD after different curing periods are presented in Table
224 4. The true density of HKMD, GGBS and ISSA is 2.525, 2.69 and 2.68 g/cm³, respectively. It is seen
225 that the bulk density and true density of the samples mixed with three kinds of solution increase with
226 time overall, and most of the true density are less than 2.525. It should be due to the hydration products
227 (e.g., C-S-H), which has a lower density of about 2.25-2.35 g/cm³ (Bernard et al., 2017). In addition, the
228 true density increases with curing time, which may be partly because of the more absorbed water during

229 the reaction processes (Li et al., 2022), and partly because of the crystallization of amorphous products
 230 or the densification of the C-S-H or calcium aluminium silicate hydrate (C-A-S-H) over time (Avet et
 231 al., 2019).



232
 233 Fig. 2. The water content and void ratio of treated HKMD with curing time.

234 Table 4. The bulk density and true density of treated HKMD.

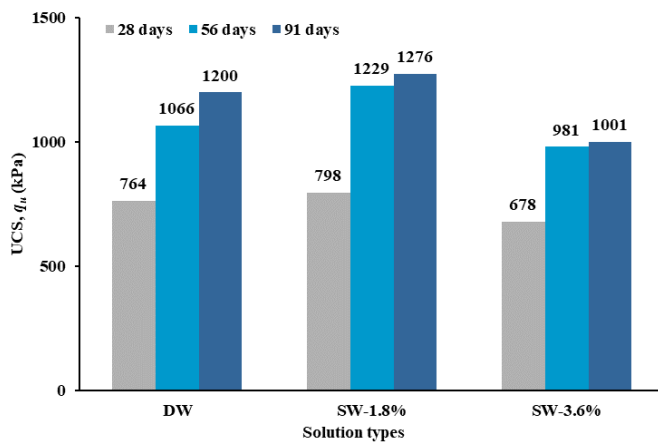
Density	Time (days)	DW	SW-1.8%	SW-3.6%
Bulk density	28	1.514	1.509	1.522
	56	1.512	1.508	1.520
	91	1.515	1.523	1.534
True density	28	2.504	2.52	2.498
	56	2.510	2.521	2.503
	91	2.520	2.539	2.532

235

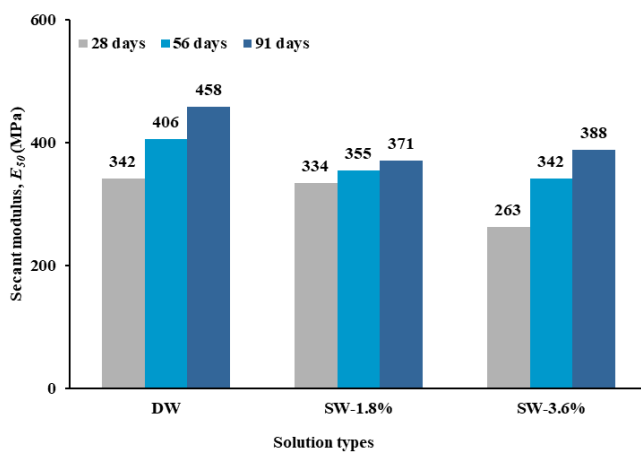
236 3.1.2 Mechanical properties of treated samples

237 The mechanical properties of S/S HKMD samples with seawater of different levels of salinity are shown
 238 in Fig. 3, in which the results in treated samples mixed with SW of 3.6% salinity is referred to Sun et al.
 239 (2023b). From Fig. 3(a). It is seen that all the specimens gain their UCS, and the UCS increases with
 240 curing time. The samples mixed with seawater of 0, 1.8% and 3.6% salinity yield a maximum UCS at
 241 the 91st curing day, which is 1200, 1276 and 1001 kPa, respectively. At 56 days, the UCS of treated
 242 samples reaches more than 85% of that at 91 days under each solution conditions. While at 28 days, the
 243 UCS is around 62-68% of that at 91 days. In general, the UCS of samples with seawater of 1.8% salinity
 244 is higher than that of samples under the other two conditions. While the HKMD slurry at initial water
 245 content of 110% exhibits no strength at all, it demonstrates that the binders of ISSA- GGBS in this work

246 are effective in S/S treatment of HKMD. Additionally, the highest strength investigated from the 91-
 247 day curing samples suggests moderate reactions in the system (Zhou et al., 2022). The deformation
 248 modulus E_{50} of the treated HKMD calculated as the secant modulus at 50% of UCS is an important
 249 parameter reflecting the deformation resistance of geomaterials. According to Fig. 3(b), the E_{50} of S/S
 250 treated HKMD specimens with three mixing proportions are also increasing with time. It is presented
 251 that the modulus of the specimens with DW is higher than that of samples with SW solutions. The
 252 findings are in consistent with the investigations about the impact of salinity on the strength of cement-
 253 treated Champlain Sea clay by Monsif et al. (2020), which found that a high salinity such as 30.55 g/L
 254 has a negative impact on the soil strength.



255
 256 (a) UCS of treated samples



257
 258 (b) Modulus of treated samples

259 Fig. 3. The mechanical properties of the treated HKMD samples.

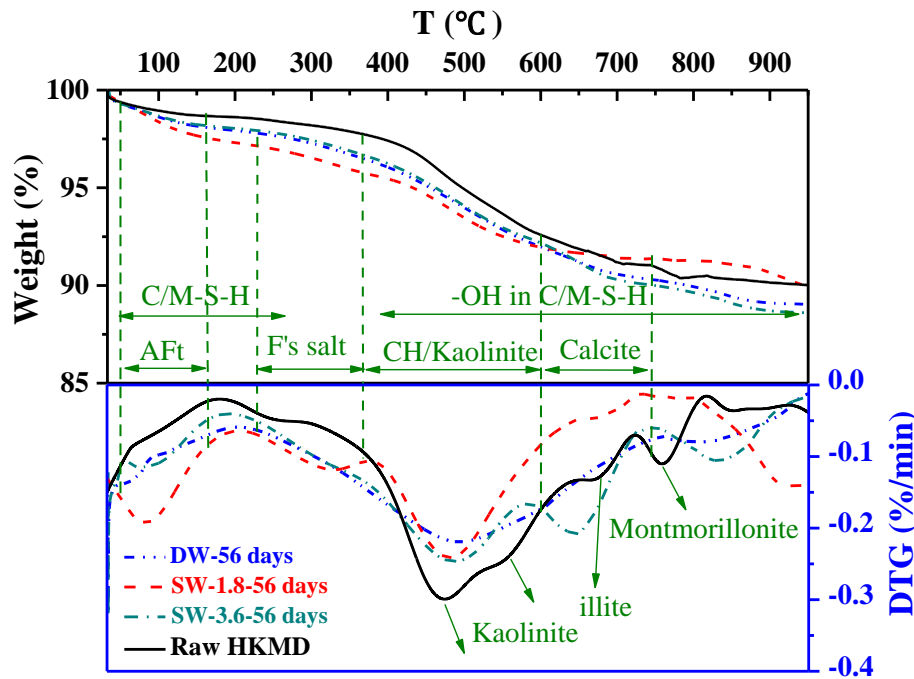
260

261 3.2 TG analysis

262 The TG and DTG curves of raw HKMD and S/S treated HKMD are shown in Fig. 4 with the temperature
263 ranging from 30-900 °C. The weight loss of each temperature period obtained from TG data are
264 presented in Table 5. Among them the data of treated samples under the SW-3.6% salinity conditions
265 are referred to Sun et al. (2023b). Firstly, from the TG and DTG curves of raw HKMD, it is obvious that
266 the most weight loss occurs at temperature from 400 to 800 °C. The peak on the DTG curve from 400
267 to 500 °C should be contributed to the dehydroxylation in kaolinite (Avet and Scrivener, 2018).
268 Additionally, the weight loss peaks at 500-700 °C represents the decomposition of illite (Húlan et al.,
269 2015), and the peak at 700-800 °C is associated with the dehydroxylation in montmorillonite (Li et al.,
270 2022).

271 After the S/S treatment of HKMD, the weight loss of the possible products in the complex system was
272 analysed. It was reported that the mass loss at 40-280 °C is due to the loss of the poorly bound water in
273 several hydration formations including AFt (Shen et al., 2017), C/M-S-H (Nied et al., 2016; Bernard et
274 al., 2017) and part of F's salt (230-370 °C) (Zhou et al., 2021b). From Fig. 4, the water loss of AFt
275 appears at 50-160 °C with the peak at around 100 °C in the treated soils (Avet and Scrivener 2018;
276 Wang et al., 2004). The Ca(OH)₂ (CH) loss is considered from 370 to 600 °C (Alarcon-Ruiz et al., 2005;
277 Wang et al., 2004; Li et al., 2019), and the decarbonation of CaCO₃ (CC) loss ranges from 600 to 720
278 °C (Prochon and Piotrowski, 2020; Lu and Poon, 2018; Scrivener et al., 2017; Sun et al., 2023b). The
279 dihydroxylation of Ca–OH in C-S-H occurs from 300-600 °C, and that of Mg–OH in M-S-H is at 400-
280 750 °C (Bernard et al., 2017, 2019). In addition, the weight loss occurs in silanol groups on M-S-H at
281 720-900 °C (Nied et al. 2016). The inner OH groups in the crystallized C/M-S-H such as the talc
282 structure dehydroxylates at 750-1000 °C (Dumas et al., 2013; Gallucci et al., 2013; Bernard et al., 2019;
283 Chakraborty et al., 2022). From Table 5, it shows that the weight loss between 400 and 600 °C is highest
284 in all samples. Since the mass content of raw HKMD in the mixture is 70%, the maximum weight loss
285 of HKMD in the treated samples is about 70% of that in raw HKMD, that is 3.37%, which is lower than
286 the values in the treated material. It confirms the new products in the mixed system.

287



288

289 Fig. 4. TG and DTG curves of HKMD before and after treatment at 56 curing days.

290 Table 5. Weight loss of HKMD before and after treatment at 56 curing days.

Condition	Weight loss (% of original)					Total
	30-200 °C	200-400 °C	400-600 °C	600-750 °C	750-900 °C	
HKMD*	0.969	0.844	3.373	1.155	0.554	6.896
DW	2.069	1.882	4.073	1.700	1.172	10.896
SW-1.8%	2.691	1.851	3.507	0.546	0.755	9.350
SW-3.6%	1.958	1.805	4.062	2.181	1.291	11.299

291 *Theory weight loss of HKMD in the treated samples is shown here, which is 70% that of the raw HKMD.

292

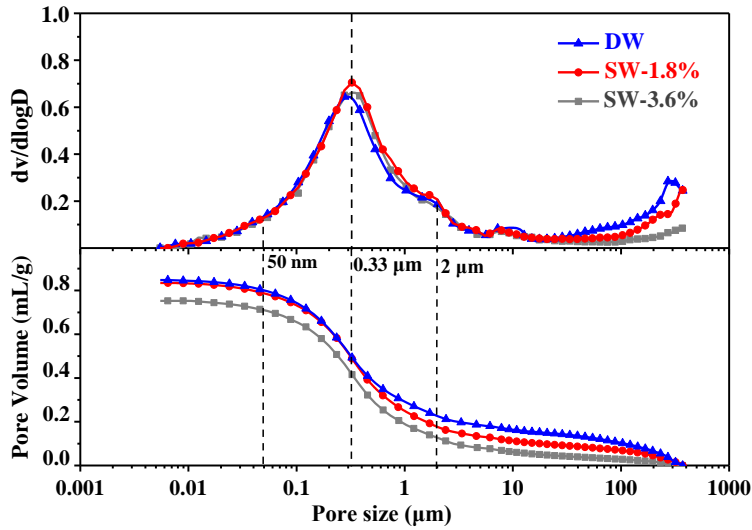
293 3.3 MIP analysis

294 The microstructure of the treated specimens at 56 curing days was detected by MIP tests with the PSD
 295 results shown in Fig. 5. The figure illustrates a unimodal PSD with a centre pore diameter of
 296 approximately 0.33 μm in the S/S samples, which suggests that the diameter of the largest pore group is
 297 approximately 0.33 μm (Sun et al., 2023b). The result is similar with the diameter of the largest pore
 298 group in the PSD of stabilized soils with metakaolin by 15% cement reported by Deng et al. (2015). The
 299 cumulative intrusion volume of the treated specimens at 56 days is also presented in Fig. 5(a), which is
 300 around 0.8 mL/g.

301 To better understand the PSD in HKMD samples after S/S treatment, the pores are classified to three
 302 groups: $> 2 \mu\text{m}$, 50 nm-2 μm and 5-50 nm. Because of the limitation of MIP method, the nano-pores (1-

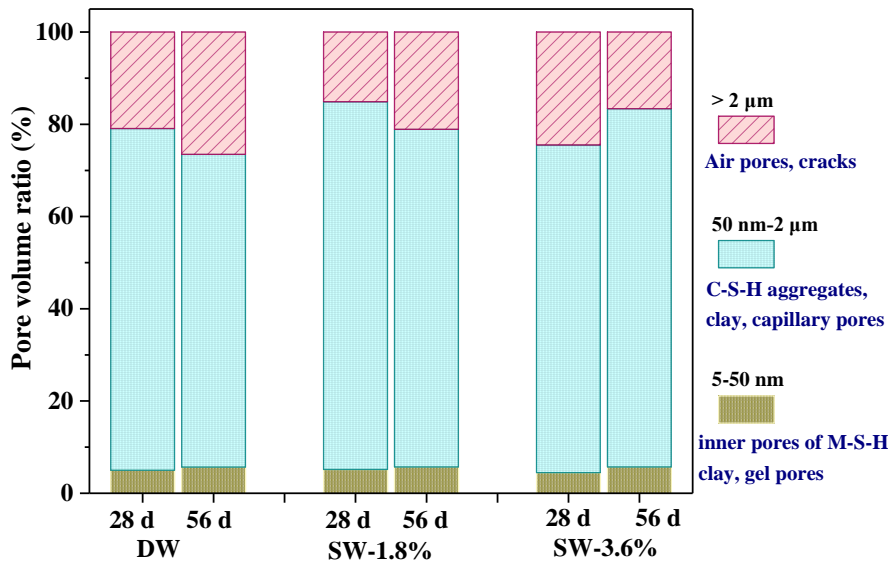
303 5 nm) in C/M-S-H gels could not be detected, and thus the nano-pore volume. It was investigated by
304 SEM observation with high magnification that the average size of inner pores of M-S-H aggregate is
305 around 30 nm (Chiang et al., 2014; Tonelli et al., 2016). While that of C-S-H aggregate is more than 50
306 nm (Chiang et al., 2014). Thus, 50 nm was selected to distinguish the inner pore size of M-S-H and C-
307 S-H aggregates. The diameter of 2 μm is a typical width of the cracks as estimated by SEM images of
308 S/S treated HKMD using lime-activated ISSA-GGBS (Sun et al., 2023b). Thus, the biggest pore size
309 group are assigned with size $> 2 \mu\text{m}$ representing the cracks and air pores in the treated soil (Dewitte et
310 al., 2022; Horpibulsuk et al., 2009).

311 Fig. 5(b) shows the pore volume percentage of the treated samples at curing age of 28 and 56 days. The
312 pore group with diameter from 5 to 50 nm is mainly composed of gel pores, pores in the kaolinite particle
313 from the deposits and pores in the M-S-H aggregates (Horpibulsuk et al., 2010; Tonelli et al., 2016; Yu
314 et al., 2016; Dewitte et al., 2022). As the pores of 5-50 nm partly represent the inner pores in M-S-H
315 aggregates, it is obvious that this part of pore volume fraction in S/S HKMD specimens increases with
316 curing days, which confirms the increasing content of M-S-H products. Salinity effect on this part of
317 pore volume is not obvious. The pore of diameter from 50 nm to 2 μm includes the inter-aggregate pores
318 in kaolinite and C-S-H, and capillary pores (Dewitte et al., 2022; Yu et al., 2016; Zeng et al. 2012). The
319 volume of pores with size from 50 nm to 2 μm in the S/S treated HKMD with three solutions is the
320 largest. The volume of largest pore group ($> 2 \mu\text{m}$) increases with time in the samples with DW and SW
321 of 1.8% salinity, which suggests the more cracks. It is in agreement with the opinion from Horpibulsuk
322 et al. (2009) that after mixing clayey soil with cementitious material, because of the physicochemical
323 interaction in the mixture the formation of clusters would reduce the small pore (e.g. $< 0.1 \mu\text{m}$) volume
324 and increase the large pore (e.g. $> 0.1 \mu\text{m}$) volume. Then, it would result in the increase of the dry unit
325 weight, which is in agreement with our density results listed in Table 4. Moreover, the clusters, such as
326 C-S-H particles, would harden and shrink during the curing process of transforming from a gel to a
327 crystalline state, which will also reduce the small pores and increase the large pores resulting in more
328 air pores or cracks (Sun et al., 2022). It is the reason that the proportions of air pores increased from 28
329 to 56 days for DW and SW-1.8% mixed samples. While, increasing salinity of seawater (from 0 to 3.6%)
330 is conducive to inhibit the development of air pores or cracks (especially $> 40 \mu\text{m}$) in the samples as
331 shown in Fig. 5(a).



332

333 (a) PSD curves of treated HKMD from MIP method



334

335 (b) Pore volume distribution of treated HKMD by MIP method.

336 Fig. 5. MIP results for the samples after 56 days curing.

337

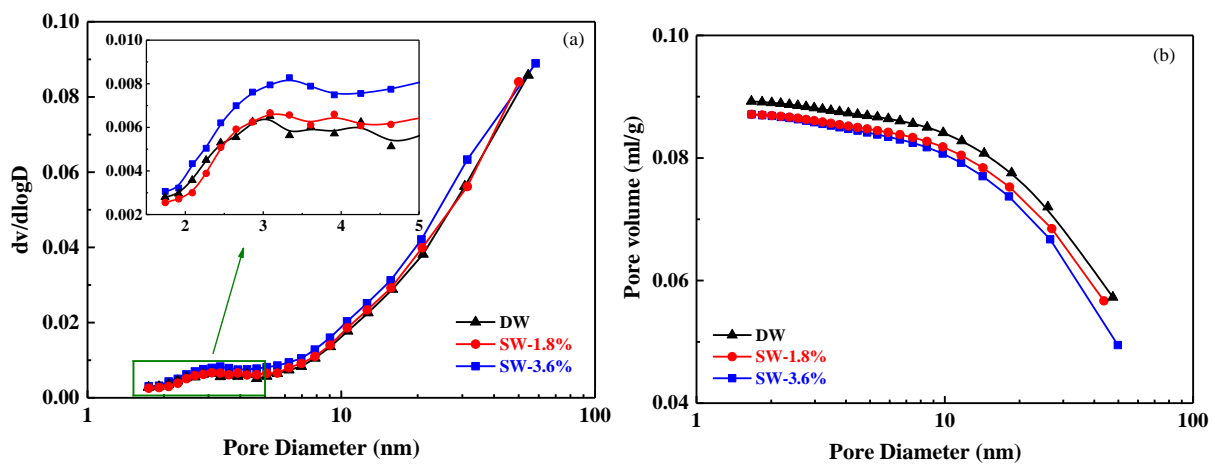
338 3.4 NAI analysis

339 The PSD of the treated HKMD in powder state after 56 curing days was determined by NAI tests with
 340 the results compared in Fig. 6. The average pore diameter and BET SSA of the treated HKMD prepared
 341 by seawater with different levels of salinity are also determined. It is identified that the SSA value of
 342 the S/S HKMD powder increases with the salinity increasing, which is 15.99, 16.46, and 17.02 m²/g
 343 under the conditions of DW, SW-1.8% and SW-3.6%, respectively. On the contrary, the average pore
 344 diameter of the treated HKMD decreases with the salinity increasing, which is 25.65, 24.59, and 23.25

345 nm under the conditions of DW, SW-1.8% and SW-3.6% (Sun et al., 2023b), respectively. It is reported
 346 that the development of surface area in cement pastes is dependent on highly porous phases, such as C-
 347 S-H (Garci Juenger and Jennings 2001; Thomas et al., 1999; Pan et al., 2015) and M-S-H (Bernard et
 348 al., 2019). Therefore, the improvement of SSA in the treated HKMD is mainly contributed to the
 349 development of C/M-S-H content. The results indicate that the treated HKMD sample with a 3.6%
 350 salinity of seawater exhibited the highest content of C/M-S-H phases, which agrees with the findings
 351 from TG analysis (see Table 5). The increasing quantity of C/M-S-H with salinity may be due to more
 352 Mg^{2+} in the SW (Garci Juenger and Jennings 2001).

353 The PSD of the S/S HKMD powders hydrated by three solutions is shown in Fig. 6. From Fig. 6(a), it
 354 is shown that the density of pores with diameter from 1.7-45 nm increases, and a peak at pore diameter
 355 of around 3 nm presents. The pore volume calculated from NAI data by using the BJH model increases
 356 in the treated HKMD in Fig. 6(b). It is also presented that the density of pores with diameter from 1.7-
 357 3 nm increases significantly, and then it almost keeps constant from 3-5 nm. The pore density of the 56-
 358 day sample mixed with SW of 3.6% salinity presents more nano pores. It can be inferred that the peak
 359 at 1.7 to 5 nm is contributed by the development of C/M-S-H phases (Pan et al., 2015). That means the
 360 increase of salinity promotes the hydration process in this work.

361



362

363 Fig. 6. The pore size distribution of treated HKMD powder at 56 curing days.

364

365 3.5 SEM-EDX analysis

366 3.5.1 Hydration products in the treated sample

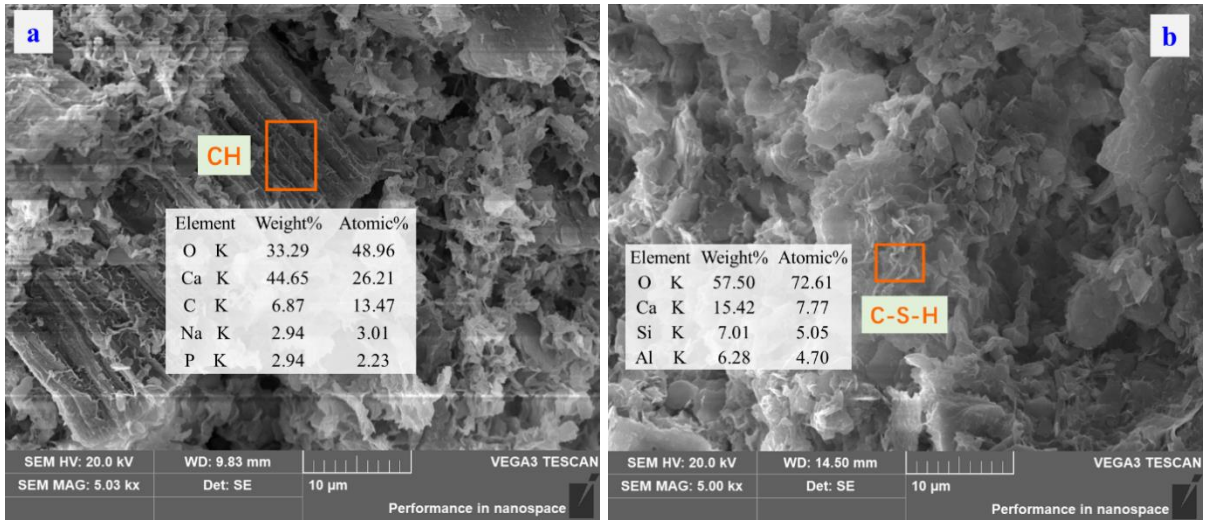
367 The SEM photos of six types of hydration products or neof ormation in the S/S HKMD samples with its
368 element compositions analysed by EDS presented in Fig. 7. From Fig. 7(a), because the high content of
369 Ca and around twice of O content, it is identified as $\text{Ca}(\text{OH})_2$. In addition, the CH displays morphology
370 of a multilayer flake aggregate. The new products with flocculent microstructure of light-colour are
371 identified as C-S-H gels (see Fig. 7(b)). In Fig. 7(c), it was found a particle with the main element
372 content of O, Si, Al, and Mg, which was considered for M-S-H gels. It is in agreement with the viewpoint
373 of Tonelli et al. (2016) and Dewitte et al. (2022) that the M-S-H particles is homogeneous morphology
374 and has the globular nature. It was also reported that CaO could react with Si and Al forming C-A-S-H
375 gels (Hwang et al., 2014). Besides, the C-A-S-H can be formed from the pozzolanic reactions between
376 kaolinite with CH in the mixture as well (Avet et al., 2019; Sun et al., 2023a). The EDS observations in
377 this work confirm the products of C-(A)-S-H gels, ettringite, and F's salt in the treated HKMD samples.
378 The fibrillar or needle-like morphology showing in Fig. 7(d) with the main element compositions of O,
379 Si, Al, Ca and S is detected as AFt type phase (Bachtiar et al., 2022; Han et al., 2012). The square
380 morphology coated by C-(A)-S-H is identified as F's salt, which is evident that most particles in the
381 treated HKMD coated by new formatting gels. The AFt, F's salt and C-(A)-S-H integrate and attach to
382 the other compositions in the system, which interlocks the particles and reinforces the structure of the
383 S/S HKMD improving the strength and producing a lower porosity (Jha and Sivapullaiah, 2016; Yuan
384 et al., 2009). In Fig. 7(f), the precipitated NaCl granule was found to be spherical in shape., which is
385 confirmed by Zhang et al. (2022b) that in the kaolinite-based geopolymer system the presence of MgCl_2
386 induced halite formation.

387

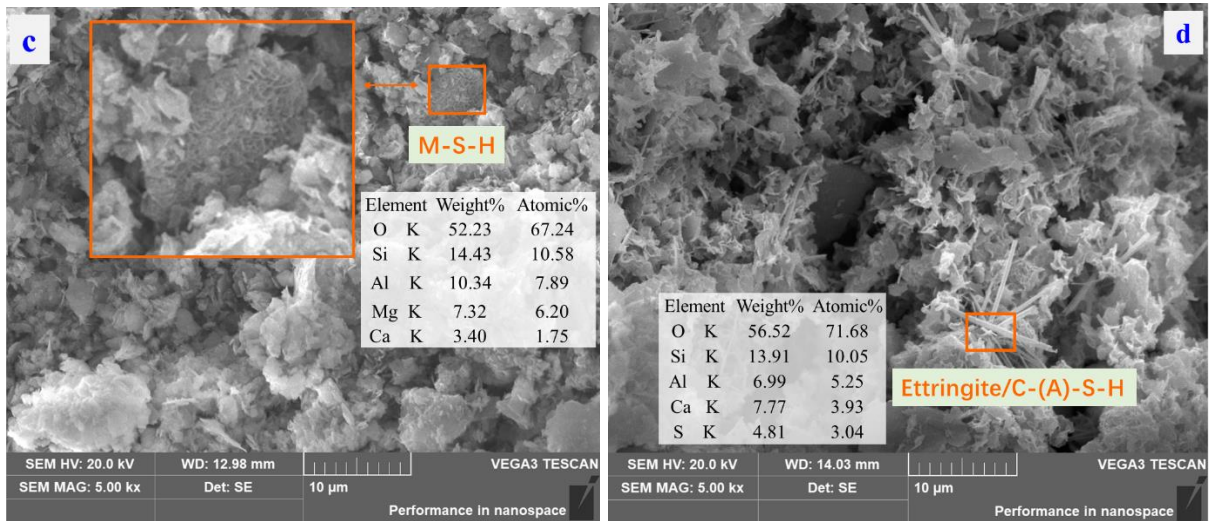
388 The EDS result of the hydration products is mainly composed of Ca, Si, Al and O and other elements
389 such as Na, S, Mg, Fe, Cl and K as well, which demonstrates that the EDS method can detect not only
390 the target component, but also other products or compounds in the mixture (Sun et al., 2023a). In general,
391 ions such as N, K, Mg and Fe presents in extremely low amounts but are negligible in the analysis
392 (Rossen and Scrivener, 2017). According to the EDS data, some of the hydrations such as C-(A)-S-H,
393 AFt phase and F's salt are mixed in Figs. 7(d) and 7(e). It is in agreement with that the C-A-S-H is often

394 used instead of C-S-H considering the incorporation of aluminum (Andersen et al., 2003; Avet et al.,
 395 2019).

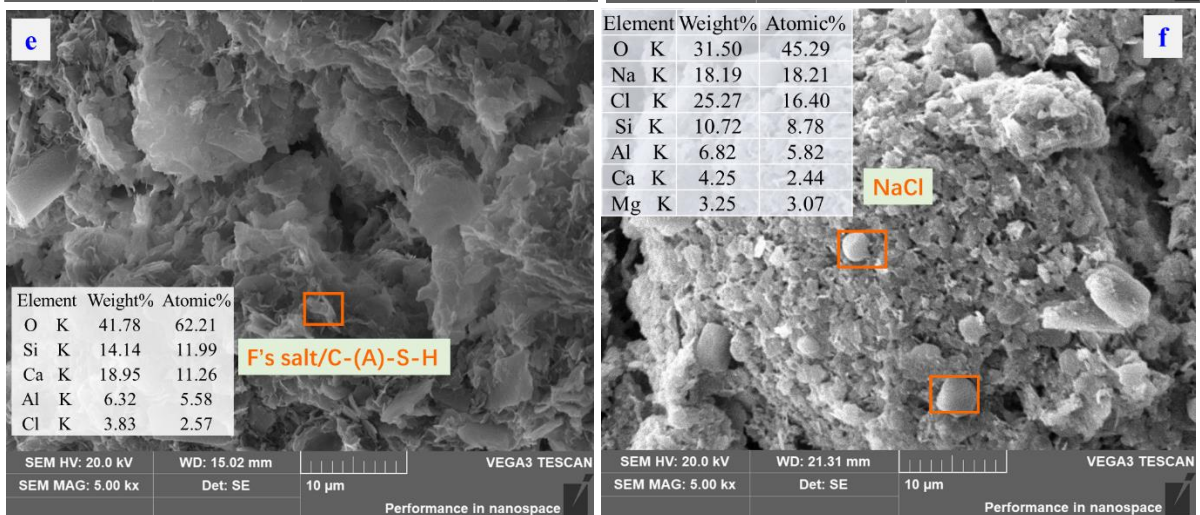
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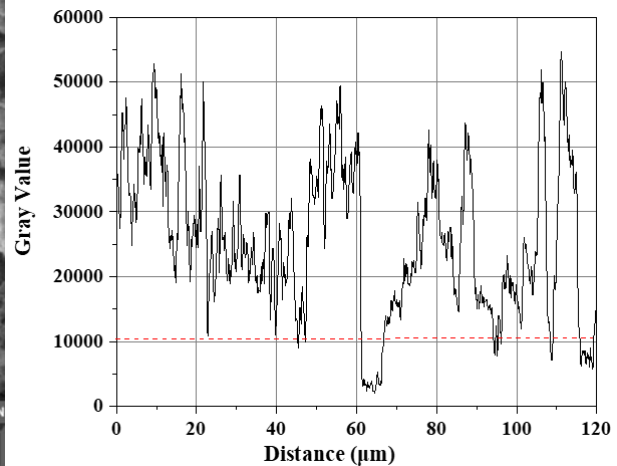
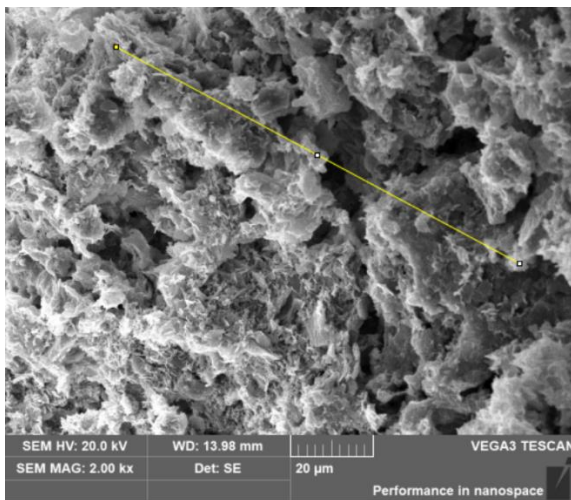


399

Fig. 7. SEM images of treated samples with different hydration products or neoformation.

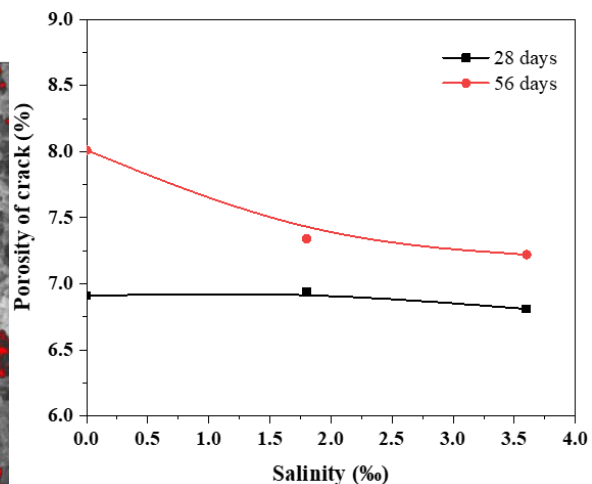
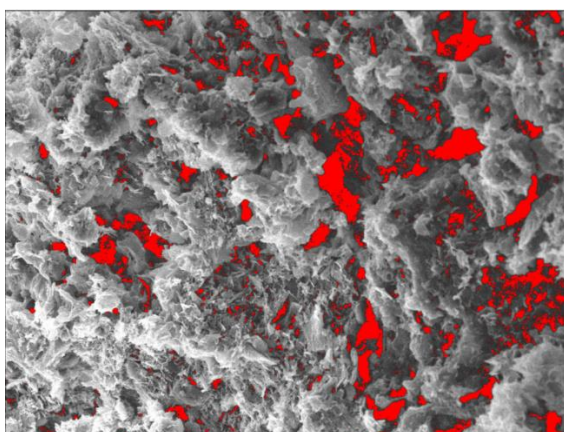
400 **3.5.2 Porosity analysis in the treated sample**

401 Many cracks or air pores ($> 2 \mu\text{m}$) are observed as presenting on the SEM photos in Fig. 8(a). To analyse
402 the porosity of cracks or air pores in S/S HKMD samples mixing with SW, ImageJ software was
403 employed. To define and calculate the porosity, the gray value of the images was firstly extracted with
404 the threshold gray of the cracks determined as shown in Figs. 8(a) and 8(b). Then, the porosity was
405 calculated and shown in Fig. 8(c). The porosity of treated HKMD mixed with different salinity seawater
406 was obtained from the average porosity in SEM images of various magnifications (from 1 kx to 10 kx),
407 which is drawn in Fig. 8(d). It can be seen that the porosity of cracks (or air pores) in S/S samples after
408 curing for 28 days decreases slightly with the salinity of substrate solution increasing and decreases
409 significantly at 56 curing days. Therefore, increasing salinity of seawater (from 0 to 3.6%) is conducive
410 to inhibit the development of cracks in the samples in this work. In addition, when the curing age
411 increases from 28 to 56 days, the porosity of cracks increases. It is consistent with the results of the MIP
412 test in Fig. 5.



(a) SEM images of treated sample

(b) Gray threshold for cracks



415

416

(c) Porosity determination

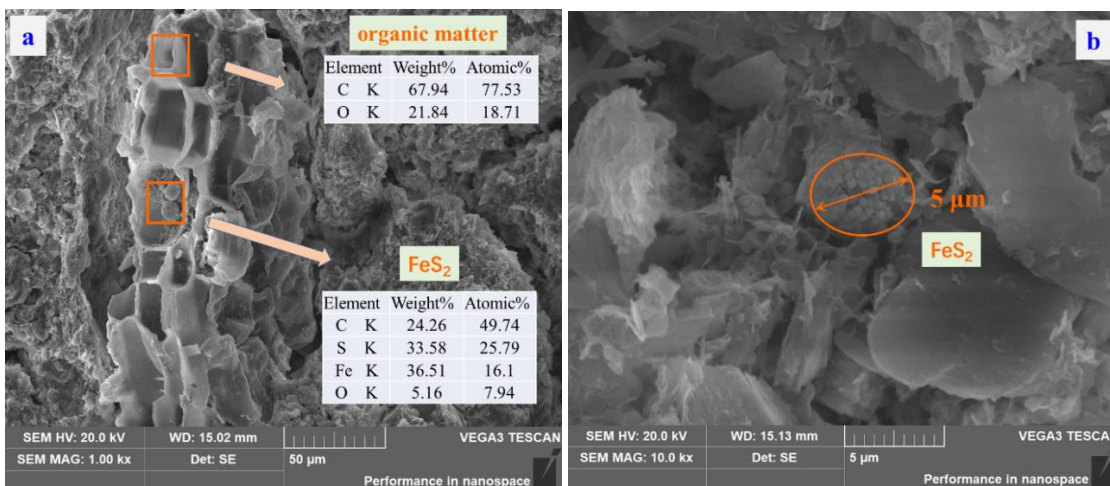
(d) Porosity of cracks in S/S samples

417

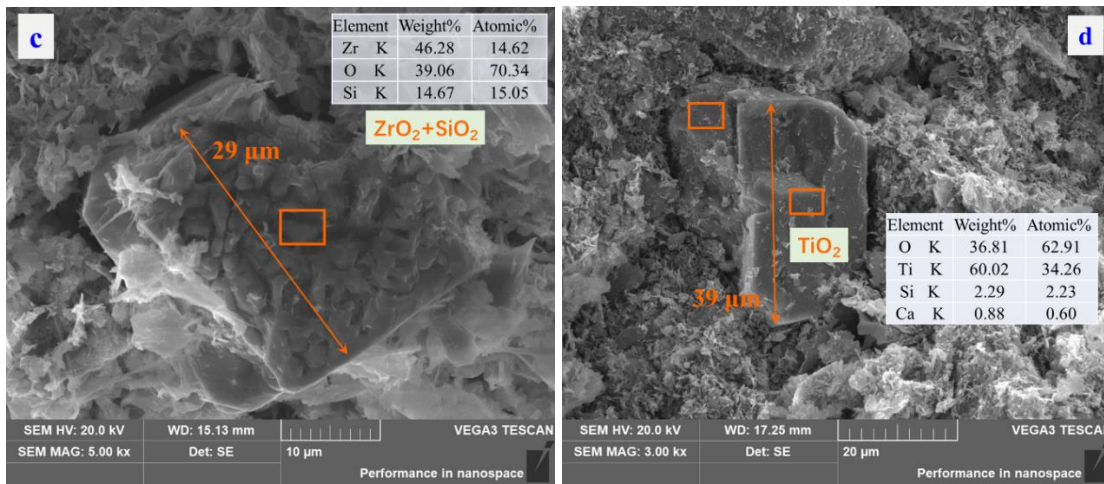
Fig. 8. Porosity of cracks and air pores analysed by using ImageJ.

418 3.5.3 Statement of heavy metals in the treated sample

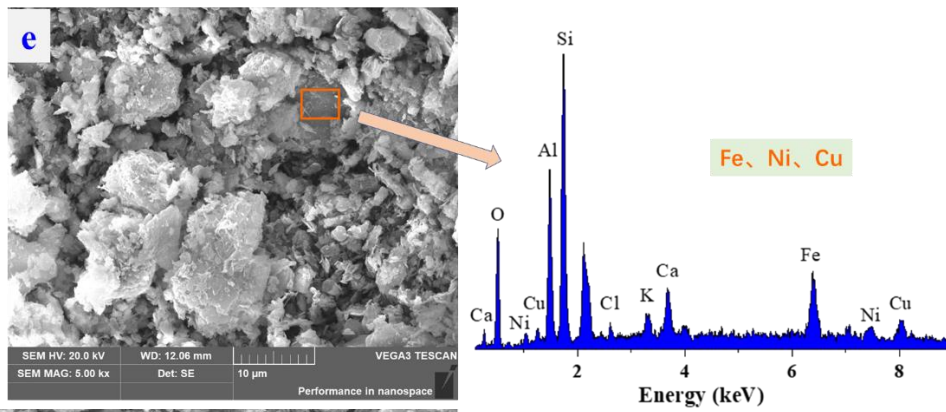
419 In addition to the main hydration products and marine deposit components, some organic matter,
 420 metallic minerals or some heavy metal elements are discovered by SEM-EDS test in both the block and
 421 powder S/S samples as shown in Fig. 9. The FeS₂ aggregate is found with the element compositions as
 422 shown in Fig. 9(a). It presents some small granular FeS₂, and its spherical aggregate has a diameter of
 423 5-10 μm in Figs. 9(a) and 9(b). From Fig. 9(c), some ellipsoidal particles on the surface of the cube with
 424 long side around 29 μm are identified as ZrO₂ by EDS methods. In addition, an incomplete compact
 425 cube with a 39 mm long side was detected as TiO₂ in Fig. 9(d). As the basic compositions in the treated
 426 HKMD system are O, Si, Ca and Al, the EDS analysis of treated HKMD in powder state displays high
 427 intensity of these elements in Fig. 9(e-g). Apart from these elements, it is investigated that the elements
 428 such as Fe, Ni, Cu, Cr, Mo, Ti, etc., are also detected in the system coating by the other compounds,
 429 which suggests that the heavy metal (loid)s or toxic elements are well solidified and stabilized in the
 430 treated HKMD.



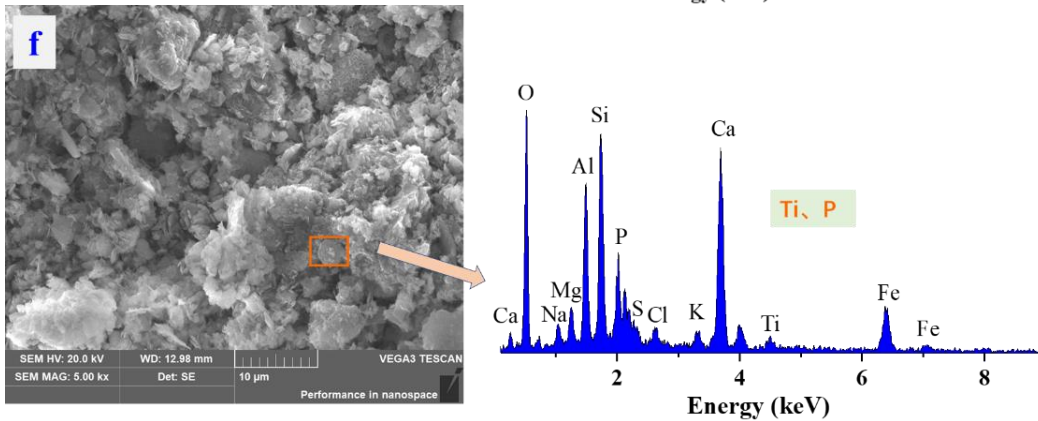
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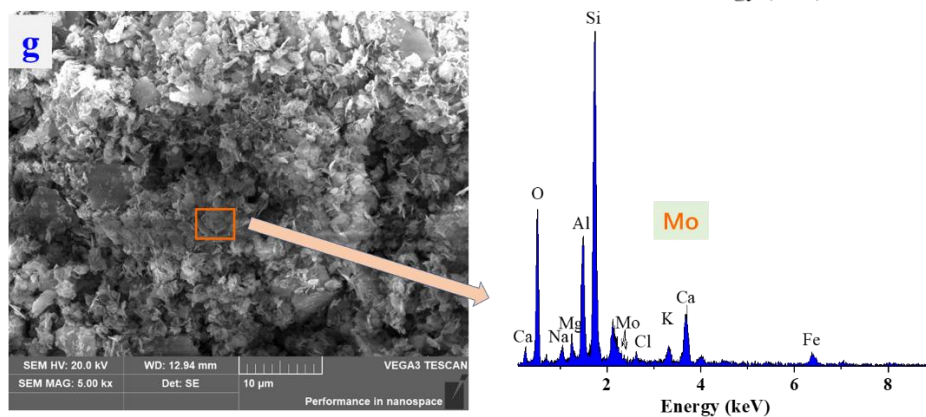
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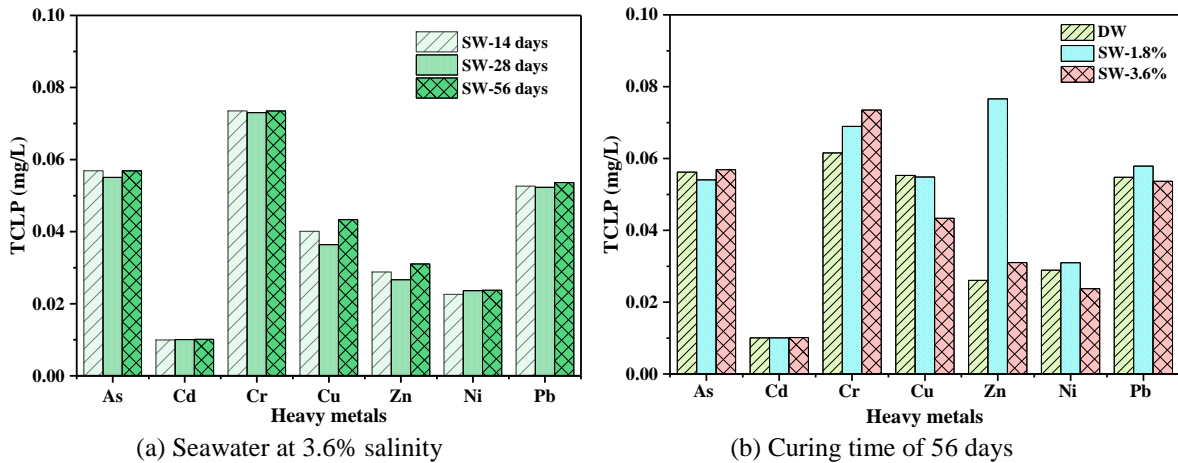
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Fig. 9. SEM-EDS spectrum of the aggregation form of heavy metal-bearing minerals and minor elements in treated samples.

440

441 3.6 TCLP analysis

442 Because the deposits and industrial wastes contain a certain amount of metal(loid)s, the environmental
443 risks of the S/S sediment materials are of great concern (Li et al. 2017). According to the TCLP test
444 results, seven heavy metals from the lime-activated ISSA-GGBS treated Hong Kong marine deposits
445 including As, Cd, Cr, Cu, Zn, Ni and Pb were analysed with the results shown in Fig. 10. As shown in
446 Table 2, the theoretical leachability of Cd (0.30 mg/L), Cr (1.80 mg/L), Pb (1.07 mg/L), and Zn (6.56
447 mg/L) from the freshly mixed materials does not meet the reuse soil standard of Hong Kong (HK EPD
448 2011). Thus, the environmental risk of the cured soil should be evaluated before the engineering use.
449 From Fig. 10, it is clear that all the leachability are environmentally acceptable with a low level of < 0.1
450 mg/L, which is far below the standard requirements of reuse treated soil (HK EPD, 2011). The results
451 confirm that in this work the S/S process in the system could effectively stabilize the contaminants
452 regardless of the curing time and the salinity of the mixing solution. Negligible toxic metals leaching
453 from the S/S HKMD samples suggest that the lime-activated ISSA-GGBS treated Hong Kong marine
454 deposits can be used for engineering applications directly without the leaching concerns.



455
456 (a) Seawater at 3.6% salinity
457 Fig.10. Speciation of heavy metals in S/S treatment samples based on TCLP test.
458

459 4 Discussions

460 4.1 Seawater salinity effect on the porosity of treated soil

461 Pore structure is a main factor that affects the mechanical property of cement-based materials (Sun et
462 al., 2022). The porosities of the S/S HKMD hydrated with seawater of different salinities were evaluated

463 from SEM observations and MIP methods with the result presented in Table 6. In order to intuitively
464 understand the accuracy of the MIP testing results, the void ratio calculated ($e_{cal.}$) from water content
465 and testing densities (see Section 3.1.1) are compared with the void ratio determined by MIP methods
466 (e_{MIP}) in Table 6. It is obvious that $e_{cal.}$ is very stable with its value around 2. Taken $e_{cal.}$ as a reference
467 value of theory void ratio, e_{MIP} are in its 90-107% range. Therefore, it can be thought that the MIP testing
468 porosity of the treated HKMD is reasonable and credible in this work. It also means that the e_{MIP} of
469 samples mixed with SW-1.8% at 28 curing days is smaller than the real one and that of samples with
470 DW at 56 curing days is larger.

471 The porosity was evaluated from SEM observations by using imageJ software with the same method
472 described in Section 3.5.2. Its value is statistically obtained after calculating and analysing SEM images
473 with magnification from 1000 to 10000. From Table 6, it is observed that the average porosity is in the
474 range of 50-65% and decreases with curing time, which is in consistent with the decreasing trend of
475 void ratio in Fig. 2(b). It has been reported that AFt is the main products impacting the pore structure of
476 S/S HKMD (Sun et al., 2023a). Thus, the decrease in porosity over time can be attributed to the increase
477 content of AFt with time during the reaction processes in the system. From Table 6, which is obtained
478 based on MIP test; the porosity is in the range of 60-70%, corresponding to the SEM results. According
479 to the analysis, there are errors in the MIP test results, especially for samples mixed with SW-1.8% at
480 28 curing days and samples with DW at 56 curing days, so there is no clear decreasing trend over time.
481 In addition, the minimum porosity is also found in the treated samples mixed with SW-1.8% from both
482 methods. It is accepted that pore structure is one of the main factors affecting mechanical performance
483 of cement-based materials because a lower porosity could enhance the strength of the block samples
484 (Liu et al., 2019). Combing with the highest UCS and AFt content (see Fig. 4(a), Table 5), it can be
485 concluded that it is feasible to use seawater in S/S treatment of HKMD, and the salinity of 1.8% is more
486 conducive to improve its strength.

487 Besides the total void ratio and the statistical porosity, the crack porosity would also influence the
488 properties of treated soil. In this work, the crack porosity estimated from SEM images decreases with
489 the increasing of salinity from Fig. 8(d), which is confirmed by the MIP results in Fig. 5(a), especially
490 for the pores with diameter $> 40 \mu\text{m}$. It also suggests that because of the lowest crack porosity the

491 intrusion pore volume of treated HKMD mixed with SW-3.6% is the smallest. Thus, higher salinity of
 492 seawater (0 - 3.6%) could inhibit the development of air pores or cracks in the S/S treated HKMD.

493 Table 6. Comparison of the void ratio and porosity calculated by different methods.

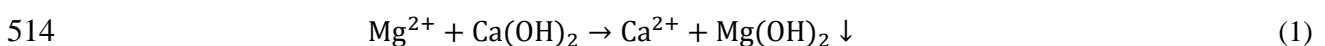
Conditions	e_{MIP}		$e_{cal.}$		Porosity-SEM (%)		Porosity-MIP (%)	
	28 days	56 days	28 days	56 days	28 days	56 days	28 days	56 days
DW	1.98	2.13	1.99	1.96	63.87	57.40	61.83	70.99
SW-1.8%	1.80	2.10	1.99	2.00	57.02	52.85	61.00	60.83
SW-3.6%	2.05	1.88	1.97	1.97	66.40	58.32	69.76	64.10

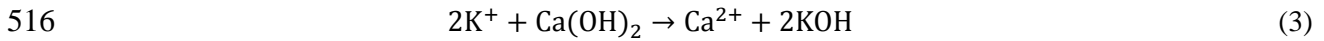
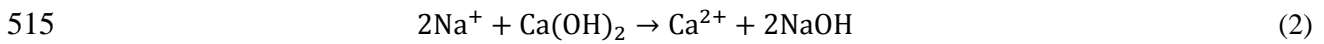
494

495 4.2 Seawater salinity effect on treated process

496 A mixture of multiple materials in the S/S HKMD makes the reactions within the system very complex.
 497 The basic reactions are the hydration and pozzolanic reactions. Among them, the hydration reactions
 498 include the hydration of lime and then the hydration of ISSA and GGBS under the alkaline conditions.
 499 The pozzolanic reactions involve the reaction of calcium hydroxide or magnesium hydroxide with clay
 500 minerals, in which clay minerals would provide amorphous silica or aluminum generating the products
 501 such as C/M-S-H, C-A-H or C-A-S-H (Gruyaert et al., 2012; Dauzeres et al., 2016).

502 It was pointed that many factors such as the clay content, organics, heavy metals, and salts contents in
 503 the sediment will affect the hydration process and strength development in the cement treated soil system
 504 (Wang et al., 2015). Additionally, the alkalinity of the mixture is also a key factor affecting the strength
 505 development of the treated clayey soil. A higher pH means a more alkaline environment, which is
 506 conducive to the hydration and pozzolanic reactions and thus the improvement of strength (Cai et al.,
 507 2022; Li and Yi, 2023). The pH of treated HKMD was measured by a pH meter with the dry soil/water
 508 ratio of 1:5 (GB LY/T 1251-199), which is 11.34, 11.38, and 11.31 under the conditions of DW, SW-
 509 1.8% and SW-3.6%, respectively. That is to say, the pH of treated HKMD with SW at 1.8% salinity is
 510 the highest. While the seawater was adopted as the mixed water, its main compositions (e.g., Na⁺, Mg²⁺,
 511 K⁺, Cl⁻ and SO₄²⁻) would have impacts on the reactions in the materials (Li et al., 2015). The presence
 512 of Mg²⁺, Na⁺ and K⁺ in the seawater will result in the change of pH in the system due to the reactions as
 513 follows:





517 Where $\text{Mg}(\text{OH})_2$ is less soluble than $\text{Ca}(\text{OH})_2$, which causes a lower alkalinity in the mixture. On the
518 contrary, NaOH and KOH will raise the alkalinity and thus the pH in the system.

519 According to Sun et al. (2023b), the strength of treated HKMD hydrated with MgCl_2 solution is the
520 lowest because of Mg^{2+} reacting with OH^- forming $\text{Mg}(\text{OH})_2$ and M-S-H phases (Sun et al., 2022),
521 which has limited contribution in the strength improvement (Wang et al., 2019; Sun et al., 2023b).
522 Besides, Song et al. (2023) reported that the strength of S/S marine deposits slurry by GGBS-MgO at
523 the initial water content of 300% is lower than that by GGBS-CaO. The main reason may be that the pH
524 of cases using GGBS-CaO is higher than that of cases using GGBS-MgO due to the difference between
525 alkalinities of hydrated CaO and MgO (Jin and Al-Tabbaa, 2014). It was also confirmed by Zhang et al.
526 (2022c) that the Na_2SiO_4 -activated geopolymer exhibited a lower compressive strength after the addition
527 of MgCl_2 . Thus, the maximum Mg^{2+} in the system under the conditions of SW-3.6% can explain the
528 lowest pH of treated soil in this case. Similarly, it can be deduced that under the conditions of SW-1.8%
529 the positive impact of Na^+ , K^+ on pH value is superior to the negative impact of Mg^{2+} , so the pH is
530 highest, and the strength of the treated HKMD is the highest. The effect of seawater salinity on the soil
531 system is the result of the synergistic action of the elements in seawater, especially the cations.

532 **5 Conclusions**

533 This study evaluated the stabilisation/solidification performance of HKMD mixed with seawater of
534 different salinity from the aspects of engineering strength and environmental risk assessment. The lime-
535 activated industrial wastes--ISSA and GGBS were adopted as binders. The following conclusions can
536 be drawn:

- 537 1) Seawater could help S/S treated HKMD gain strength, and increasing the salinity of seawater is
538 conducive to inhibit the development of cracks in the S/S samples.
- 539 2) Seawater of 1.8% salinity is better than distilled water and seawater of 3.6% salinity as a
540 substrate solution in the S/S process, because of the highest UCS and lowest porosity in treated
541 samples.

542 3) Based on the TCLP results, the S/S process could effectively stabilize the contaminants in the
543 treated clayey soils with the leachates meeting the reuse standard in Hong Kong.
544 Therefore, this study proves that the S/S treated dredged clayey deposit by industrial wastes-ISSA and
545 GGBS is technically feasible, economically sustainable and environmentally clean, which is possible to
546 use it as geomaterials in reclamation projects.

547

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554 Polytechnic University.

555

556 **Declaration of interests**

557 The authors declare that they have no known competing financial interests or personal relationships that
558 could have appeared to influence the work reported in this paper.

559

560 **Authors' contributions**

561 **Zhao SUN**: Testing, analysis, investigation, and writing; **Wen-Bo CHEN**: Conceptualization, funding
562 acquisition, review, and editing; **Run-Dong ZHAO**: Testing, and investigation; **Jiang-Shan LI and**
563 **Zhenyu YIN**: Review and editing; **Jian-Hua YIN** and **Yong-Gui CHEN**: Supervision, project
564 administration, review.

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