

1 **Personal exposure to fine particles (PM_{2.5}) and respiratory inflammation of common**
2 **residents in Hong Kong**

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27 **Abstract**

28 **Background:** Given the lack of research on the personal exposure to fine particles (PM_{2.5}) in
29 Hong Kong, we examined the association between short-term personal exposure to PM_{2.5} and
30 their constituents and inflammation in exhaled breath in a sample of healthy adult residents.

31 **Method:** Forty-six participants underwent personal PM_{2.5} monitoring for averagely 6 days to
32 obtain 276 samples. Fractional exhaled nitric oxide (FeNO), a biomarker of inflammation in
33 exhaled breath, was measured at the end of each 24-hour personal monitoring. PM_{2.5} chemical
34 constituents, including organic carbon, elemental carbon, 16 polycyclic aromatic hydrocarbons
35 (PAHs), and 6 phthalate esters, were speciated from the personal samples collected. A mixed-
36 effects model was used to estimate the association of PM_{2.5} and their constituents with FeNO.
37 The comparison was also made with parallel analyses using ambient concentrations.

38 **Results:** Personal exposures to PM_{2.5} ($28.1 \pm 23.3 \mu\text{g}/\text{m}^3$) were higher than the ambient levels
39 ($13.3 \pm 6.4 \mu\text{g}/\text{m}^3$) monitored by stations. The composition profile and personal-to-ambient
40 concentration ratio varied among subjects with different occupations. An interquartile range
41 (IQR) change in personal exposure to PM_{2.5} was positively associated with 12.8% increase in
42 FeNO (95% confidence interval, CI: 5.5-20.7%), while nil association was found for ambient
43 PM_{2.5}. Among the constituents measured, only the carcinogenic PAHs were significantly
44 associated with 12% increase in FeNO responses (95% CI, 0.0-25.6%).

45 **Conclusion:** In conclusion, our study provides the first understanding about personal exposure
46 to PM_{2.5} and possible sources in Hong Kong. The results also showed that personal exposure to
47 PM_{2.5} and c-PAHs were linked to increased FeNO levels among healthy adults.

48 **Keywords:** Fine particles; Personal exposure; Respiratory inflammation; Carbonaceous
49 materials; Polycyclic aromatic hydrocarbons

50 **1. Introduction**

51 Numerous epidemiologic studies have documented that fine particulate matter (PM_{2.5}) is
52 associated with inflammation-related diseases such as asthma and chronic bronchitis (Kunzli et
53 al., 2009; Pope and Dockery, 2006). PM_{2.5} is a complex mixture of various organic and
54 inorganic chemical substances and its toxicity changes with its composition (Osornio-Vargas
55 et al., 2003). Therefore, the identification of hazardous components to health is crucial for the
56 implementation of efficient air pollution control strategies. Elemental carbon (EC) and organic
57 carbon (OC), which is frequently measured in epidemiologic studies, are important PM
58 compositions (Jansen et al., 2005; Lin et al., 2011). Polycyclic aromatic hydrocarbons (PAHs)
59 are known for their carcinogenicity; both experimental and epidemiological evidence of PAHs
60 indicated proinflammatory effects on airways (Delfino et al., 2010). Phthalates are common
61 industrial chemicals used in cosmetics, personal care products, plastics, and building materials.
62 Their occurrence in PM_{2.5} have been proved by previous papers (Rakkestad et al., 2007; Tran
63 and Kannan, 2015). Serial investigations done by the US National Health and Nutrition
64 Examination Survey (NHANES) have provided effective evidence for the relationships
65 between total PAEs exposure and airway inflammation, deteriorated lung functions and allergic
66 symptoms (Ferguson et al., 2011; Hoppin et al., 2004). However, there is still lack of reports
67 on the contribution and potential effect of phthalates exposure through inhalation. Most
68 studies have used ambient measure of PM_{2.5} in the assessment of the association between
69 particulate air pollution and health. However, such ambient measurement tends to reflect the
70 urban background of PM_{2.5}, rather than the actual personal exposure, which can be significantly
71 affected by different individual activities and the time spent in various microenvironments (Lee
72 et al., 2010; Lim et al., 2012). Currently, a limited number of studies have examined the adverse
73 effect of personal exposure to PM (Auger et al., 2006; Commodore et al., 2013; Huang et al.,
74 2012; Meier et al., 2014), especially in the general healthy population. Means of measuring
75 inflammatory biomarkers made it possible to assess adverse health effects of PM_{2.5} on the
76 general population. Fractional exhaled nitric oxide (FeNO) is a sensitive noninvasive biomarker
77 of airway inflammation that has been used in many epidemiologic studies of the impact of air

78 pollutants on healthy and asthmatic subjects (Jansen et al., 2005; Koenig et al., 2003). The
79 American Thoracic Society (ATS) and the European Respiratory Society (ERS) now
80 recommended FeNO to be a clinical surrogate marker of eosinophilic airway inflammation
81 (Reddel et al., 2009).

82 In recent years, few researchers have focused on the association between personal exposure
83 to PM_{2.5} and FeNO levels in healthy adults, and the findings are inconsistent (Adar et al., 2007;
84 Boogaard et al., 2013a; Kubesch et al., 2015; Strak et al., 2010). One panel study conducted in
85 USA 2002 estimated personal exposure to PM_{2.5} among healthy non-smokers based on the
86 concentrations measured in microenvironments and found significant effects on FeNO levels
87 (Adar et al., 2007). A cross-sectional study reported a null association between ambient PM_{2.5}
88 and FeNO levels among its 661 adult residents in Netherlands (Boogaard et al., 2013b). In Hong
89 Kong, while numerous epidemiological studies have linked PM measured from central
90 monitors with adverse respiratory outcomes, hospital admission, and mortality, no study has
91 examined levels of personal exposure to PM_{2.5} and its health association among common Hong
92 Kong residents. Thus, this study aimed to fill the data gap and evaluate the association between
93 exposure to PM_{2.5} (and their constituents) and respiratory inflammation in healthy adults in
94 Hong Kong.

95

96 **2. Materials and methods**

97 **2.1 Subjects**

98 The target study population was designed to be non-smoking healthy adults aged 18-45 years
99 old, with no known allergies and other chronic diseases, and with regular living lifestyles. On-
100 site and online advertisements were produced for three months; seventy-nine residents
101 responded to the advertisements with informed consent and subsequently completed a self-
102 administrated questionnaire about demographics, health status, smoking and symptoms related
103 to asthma, rhinitis, and eczema. Among them, 46 met the inclusion criteria and agreed to
104 participate.

105 This longitudinal study spanned across two sampling sessions: June 23, 2014 - September 7,
106 2014 and June 23, 2015. Each participant was required to complete one sampling session, with
107 daily active personal monitoring (24 hours) to measure personal exposure to air pollutants for
108 six consecutive days in each session. FeNO, the biomarker of respiratory inflammation, was
109 measured at the end of each sampling day (5-7 pm) and at least two hours after meal. Every
110 subject was instructed to avoid taking anti-inflammatory medication and vitamin
111 supplementation during the sampling period, and sampling would be stopped and rescheduled
112 if participants developed acute infectious illnesses. Daily activities and respiratory symptoms
113 of the participants were recorded hourly on a self-administered diary and checked by our
114 research assistants. The time duration each participant spent at different locations (indoor,
115 outdoor or on transportation) was estimated according to the diary and used for analysis.
116 Meteorological parameters, including temperature and relative humidity (RH), were obtained
117 from the website of Hong Kong Observatory (Zhu et al., 2010). In this study, the impact of
118 confounding by between-subject characteristics was limited, as each subject acted as his/her
119 control over time in this kind of longitudinal study with repeated measurements.

120

121 2.2 Personal exposure to PM_{2.5}

122 Personal PM_{2.5} were collected with a sampler operated by battery-powered Leland Legacy
123 pump (SKC Inc., PA) at a flow rate of 10 liters per minute (L/min). Each participant was
124 equipped with a suitcase containing the pump connected to an impactor loaded with a 37-mm
125 quartz filter (Whatman Ltd, Maidstone, UK). They were asked to carry, or keep the personal
126 suitcases near them and attach sampler inlets near the breathing zone as they underwent their
127 daily activities. Five participants in a batch were evaluated in parallel with the other five
128 samplers stored for replacement and temporary use. After sampling, the exposed quartz filters
129 were collected by our research assistants and stored in a refrigerator at about -20 °C until
130 chemical analysis. All the filters were cut into two sections for chemical analysis. The first
131 section was analyzed for OC and EC using thermal optical reflectance (TOR) and the second
132 section was analyzed for PAHs and PAEs by thermal desorption-gas chromatography/mass

133 spectrometer (TD/GC/MS) method. 16 PAHs, including acenaphthylene (ACN), acenaphthene
134 (ACE), fluorene (FLU), phenanthrene (PHE), anthracene (ANT), fluoranthene (FLUT), pyrene
135 (PY), benzo[a]anthracene (BaA), chrysene (CHR), benzo[b]fluoranthene (BbF),
136 benzo[k]fluoranthene (BkF), benzo[a]pyrene (BaP), indeno[1,2,3-cd]pyrene (IND),
137 dibenzo[a,h]anthracene (DBA), benzo[ghi]perylene (BP), and six PAEs, including dimethyl
138 phthalate (DMP), diethyl phthalate (DEP), di-n-butyl phthalate (DnBP), butyl benzyl phthalate
139 (BBP), bis(2-ethylhexyl)phthalate (DEHP), di-n-octyl phthalate (DnOP) were selected as the
140 targeted constituents. The detail analytical procedures and performance characterization were
141 described in previous papers (Cao et al., 2005; Ho and Yu, 2004; Ho et al., 2008).

142 The ambient concentrations of PM_{2.5} and other pollutants, i.e., SO₂, NO₂, O₃ and PM₁₀, from
143 seven general air quality monitoring stations (i.e., Central Western, Eastern, Kwun Tong, Sha
144 Tin, Tai Po, Tap Mun and Yuen Long) were downloaded from the website of Hong Kong
145 Environmental Protection Department (HKEPD) (Zhao et al., 2006). Figure S1 displays the
146 locations of the seven stations on map. The ambient data points were matched with personal
147 exposure data points according to the location information provided by each subject's daily
148 activity diary. Ambient concentrations of coarse particles (PM_c) were calculated as the
149 difference between ambient PM₁₀ and PM_{2.5}. In our previous studies, PM_c were found to be
150 associated with hospital admissions of cardiovascular and respiratory diseases in Hong Kong
151 (Qiu et al., 2014). Therefore, ambient PM_c were included in this study as a potential confounder.
152 Data from different stations was used for analysis according to the sampling date and living
153 district of the various subjects.

154

155 2.3 FENO data

156 FeNO was measured at the end of each sampling day with a handheld FeNO device (NIOX
157 MINO Airway Inflammation Monitor; Aerocrine AB, Solna, Sweden) in accordance with the
158 ATS guidelines (2005). Subjects were instructed to exhale and then inhale to total lung capacity
159 through the device, which provides nitric oxide-scrubbed air. Scrubbed air is used for the zero-
160 reference comparison performed in the instrument during every measurement cycle.

161 Subsequent exhalation at a steady rate for 10 seconds at a flow of 50 ± 5 mL/s was aided by a
162 built-in flow control unit of the device, consisting of a mechanical pressure-flow regulator
163 establishing a constant flow when applying an exhalation pressure of 10 to 20 cm H₂O. The
164 lowest detection limit is 5 ppb. The accuracy range of NIOX MINO is ± 5 ppb for measured
165 values less than 50 ppb and $\pm 10\%$ for 50 ppb or greater. Data were stored electronically in the
166 device and written down by our research assistants on the record form of each subject for the
167 whole study period.

168

169 2.4 Quality Assurance and Quality Control (QA/QC)

170 Appropriate quality assurance and quality control were implemented in handling of filters,
171 personal sampling, and chemical analysis. During the sampling campaign, the pump flow rate
172 was measured and adjusted to 10 L per minute ($\pm 5\%$) before sampling and measured again at
173 the end of each sampling session with a DryCal Lite flow meter (Bios Int., USA). The pre- and
174 post-sampling flow rates were averaged to calculate the PM_{2.5} air concentrations. Two field
175 blanks were collected per sampling batch for correction. These PEMs were placed in a sealed
176 plastic bag and carried to corresponding workplaces to imitate the transportation process of real
177 monitors.

178 Other QA/QC procedures used in the chemical analysis were as the same as those have been
179 previously presented by Ho et al. (Ho et al., 2006; Ho et al., 2011). In brief, the analyzer was
180 calibrated with known quantities of CH₄ every day. Twenty blank filters collected were also
181 analyzed, and the sample results were corrected by the average of the blank concentrations,
182 which were 1.0 $\mu\text{g}/\text{m}^3$, 0.1 $\mu\text{g}/\text{m}^3$, 3.1 ng/m^3 and 10.5 ng/m^3 for OC, EC, total PAHs and total
183 PAEs, respectively. The targeted PM_{2.5} constituents and their detection limits are listed in Table
184 S1.

185

186 2.5 Statistical analysis

187 We analyzed the association between repeated measures (within-subject) of FeNO and personal
188 exposures to air pollutants using linear mixed effects models with random subject effects. A

189 compound symmetry structure was preferable for the covariance matrix to model the correlation
190 between repeated measures for each subject based on the Akaike's Information Criterion (Zhao
191 et al., 2013). Since FeNO was right skewed, it was log-transformed to fulfill the assumption of
192 residual normality for linear mixed models. Since the samples with concentrations below the
193 detection limit (5 ppb) only accounted for 8% (22 out of 276) of the total observation points
194 (Lubin et al., 2004), and these values were substituted by half of the detection limit (2.5 ppb)
195 for statistical analysis. Personal exposures to PM_{2.5} were characterized by calculating the
196 individual 24-h average concentration immediately preceding the FeNO monitoring. Personal
197 exposures to PM_{2.5} constituents were later characterized by chemical analysis.

198 Time-dependent variables including day of the week and relative humidity (RH) were
199 controlled in the crude model as covariates. Additional time-independent variables, including
200 age, gender, BMI, education level, occupation, and household income, were also added in the
201 adjusted model to examine their confounding effects. Furthermore, two-pollutant models were
202 conducted to examine whether the association between PM_{2.5} and FeNO was consistent while
203 controlling for ambient gas phase pollutants (SO₂, NO₂, and O₃) or ambient PMc. The
204 relationship between PM_{2.5} constituents and FeNO was also adjusted for personal PM_{2.5} to
205 examine the possible confounding effects. β was the estimated coefficient of a pollutant from
206 the mixed-model. In order to allow hazards risk for different pollutants to be compared by
207 limiting differences due to units of measurement or concentration range, magnitudes of
208 association are also expressed at pollutant interquartile ranges (IQR; 25th–75th percentile)
209 following calculation: $(\exp^{\beta \times \text{IQR}} - 1) \times 100\%$ (Wu et al., 2010).

210 Residual analyses were performed to examine deviations from standard linear mixed model
211 assumptions and the presence of influential observations. Sensitivity analysis was conducted to
212 explore the model robustness (1) by using a more parsimonious model and an extended one
213 (covariates and ambient gaseous pollutants included), (2) by removing FeNO levels lower than
214 the detection limit and (3) by removing FeNO levels higher than 50 ppb since it is the cut point
215 for high FeNO level in adults suggested by ATS (Dweik et al., 2011). All the data analyses
216 were implemented in the R software version 3.1.3 with package '*nlme*'.

217

218 2.6 Ethics statement

219 This study was carried out after obtaining approval from Joint Chinese University of Hong
220 Kong-New Territories East Cluster Clinical Research Ethics Committee (Ref. No. CRE-
221 2014.154). All the data and sample collections were conducted after informed consent was
222 obtained. The consent form included a general description of the study. It assured the subject
223 of the confidentiality of information and his/her right to opt out of the study with no
224 consequence. All questions regarding the study were answered prior to the interview.

225

226 3. Results

227 A total of 276 observations were obtained for the 46 subjects recruited. Five subjects
228 participated for more or less than six days due to unforeseen conditions or logistical reasons.
229 The detailed participating schedule is listed in Table S2. Of the 276 exposure observations,
230 three (from different subjects) had no corresponding FeNO concentration because of
231 emergencies of subjects or unexpected mistakes during field sampling, yielding only 273 FeNO
232 values. As shown in Table 1, FeNO levels were averagely higher in males (Mean: 20.6 ± 19.4
233 ppb) than in females (Mean: 7.7 ± 3.7 ppb). The age of 46 healthy adults participated in this
234 study ranged from 18 to 30 years of age, and the average BMI was 21.4 ± 3.1 kg/m². According
235 to the classification of World Health Organization (WHO) (Zhang et al., 2014), most subjects
236 had normal weight, four subjects were underweight (BMI < 18.5 kg/m²) and one subject was
237 obese (BMI ≥ 30 kg/m²). The subjects' occupations include teacher, student, office worker and
238 unemployment. As displayed in Figure 1, subjects in this study spent $14.9 \pm 17.3\%$ of their time
239 outdoors, $4.3 \pm 6.0\%$ of their time on transportation and $80.7 \pm 18.0\%$ of their time indoors.
240 Among all the subjects with different occupations, unemployed subjects averagely spent the
241 longest time indoors ($84.2 \pm 11.4\%$), while the average time spent by office workers indoors is
242 shortest ($76.5 \pm 26.5\%$).

243 The 24-hour average concentrations of ambient PM_{2.5}, personal PM_{2.5} and their constituents
244 are listed in Table 2. Corresponding concentration information of PMc and ambient gaseous

245 pollutants (NO₂, SO₂, O₃) are shown in Table S3. About 7% of the personal exposure data were
246 missing due to filter damages during sampling. The average concentrations of ambient PM_{2.5}
247 (13.3 ± 6.4 µg/m³) was significantly lower ($p < 0.001$) than that of personal exposures (28.1 ±
248 23.3 µg/m³). The Spearman's correlation coefficient was 0.236 ($p < 0.001$), indicating a
249 relatively weak correlation between personal and ambient PM_{2.5} concentrations. Personal to
250 ambient PM_{2.5} ratio is a good indicator of the exposure differences caused by individual-level
251 factors rather than the ambient concentrations. It can be seen from the Figure 2 that
252 personal/ambient ratios for subjects with different occupation and different gender varied. The
253 average ratios for office workers (3.9 ± 2.7) were higher than those for students (2.2 ± 1.5),
254 teachers (2.1 ± 1.4) and unemployed subjects (2.3 ± 1.5) with p -value < 0.01. Comparable
255 average ratios were found for males (2.4 ± 1.8) and females (2.3 ± 1.8).

256 The mean concentrations of OC, EC, sum of eight c-PAHs, sum of six PAEs, and sum of 16
257 PAHs were 10.1 ± 14.7 µg/m³, 2.4 ± 2.1 µg/m³, 0.3 ± 0.2 ng/m³, 0.7 ± 0.3 ng/m³, and 471 ±
258 603 ng/m³, respectively. The composition profiles of personal samples are shown in Figure 3.
259 On average, OC, EC, sum of six PAEs and sum of 16 PAHs accounted for 35.9%, 8.5%, 1.7%
260 and 0.0023% of personal PM_{2.5}, respectively. Whereas, the composition profiles of personal
261 PM_{2.5} varied among subjects since different constituents have different sources including
262 biomass combustion, vehicle diesel, cleaning products, etc. and the contribution percentages of
263 different sources were influenced by personal activities, living, and working environments and
264 ambient levels simultaneously. It can be seen from Figure 3 that the average concentration
265 percentage of total OC was highest for the subject group of unemployment (47.2 ± 17.3%),
266 while the percentages were 42.0 ± 17.4%, 34.3 ± 12.6% and 24.1 ± 10.2% for office workers,
267 students, and teachers, respectively. The obvious difference between the composition profiles
268 of PAHs/PAEs for different occupation groups was not found. On average, c-PAHs accounted
269 for 46.8 ± 13.4% of the 16 PAHs monitored in this study and DEHP was the predominant PAE
270 (79.3 ± 16.3%). The concentrations of personal PM_{2.5} were highly correlated with OC ($r = 0.56$,
271 $p < 0.05$) and EC ($r = 0.54$, $p < 0.05$), moderately correlated with PAHs ($r = 0.35$, $p < 0.05$) and
272 weakly correlated with PAEs ($r = 0.17$, $p < 0.05$) (Table 3). We observed significant correlations

273 between personal PM_{2.5} and all the other variables in Table 3. Ambient PM_{2.5} levels were
274 significantly associated all the other pollutants except for PAEs. Ambient RH has a negative
275 relationship with all the air pollutants listed. Concentrations of total PAHs and c-PAHs were
276 highly correlated ($r = 0.90, p < 0.05$).

277 The regression results for the association between FeNO and air pollutants are listed in Table
278 3. FeNO showed increases of 11.1% (95% CI: 3.9-18.8%) in the crude model and 12.8% (95%
279 CI: 5.5-20.7%) in the multivariable-adjusted model per 16.4 $\mu\text{g}/\text{m}^3$ increment of personal PM_{2.5}.
280 The IQRs were 3.2 $\mu\text{g}/\text{m}^3$, 1.5 $\mu\text{g}/\text{m}^3$, 0.43 ng/m^3 , 0.277 ng/m^3 and 597 ng/m^3 for OC, EC,
281 PAHs, c-PAHs and PAEs, respectively. As displayed in Table 4, the highest estimated effect
282 was observed for c-PAHs (12.0%, 95% CI: 0.0%-25.6%), followed by total PAHs (8.5%, 95%
283 CI: -3.2, 21.7%), EC (4.5%, 95% CI: -3.3, 12.8%), OC (1.8%, 95% CI: -0.7, 4.4%) and PAEs
284 (1.5%, 95% CI: -6.1, 9.6%) according to adjusted models. The regression results for the
285 associations between FeNO and different PAHs are shown in Figure 4 and Table S4. Effect
286 estimates of the 16 monitored PAHs varies from -5.6% to 12.6% in the crude model and from
287 -4.8% to 12.8% in the adjusted model. Significant associations were observed for BbF and BkF
288 according to both crude and adjusted models. The effect estimates for other c-PAHs (BaA, CHR,
289 BaP, INP, DBA and BP) are in the range of 0.9-9.5% according to adjusted models. In
290 comparison with personal PM_{2.5}, ambient PM_{2.5} concentrations were weakly and insignificantly
291 associated FeNO in both crude and adjusted models (0.8%, 95% CI: -7.5, 9.9% for the adjusted
292 model). Adjusting for confounders led to increases in the effect estimates, and the significance
293 of the association between exposures and FeNO in this study.

294 Sensitivity analyses were performed to test the robustness of the associations. Table S5
295 lists the association between personal PM_{2.5} and FeNO from the adjusted model based on all
296 data points, a dataset with FeNO > 50 ppb removed and dataset with FeNO < 5 ppb removed.
297 Neither of removing FeNO values higher than 50 ppb and removing FeNO values lower than 5
298 ppb showed significant influence on the association between personal PM_{2.5} and FeNO.
299 Removing values less than 5 ppb slightly increased the effect estimate and significance of the
300 association. Table S6 displays all the associations between ambient or personal PM_{2.5} (and their

301 constituents) and FeNO determined by the crude model, adjusted model and two-pollutant
302 models. A significant association between ambient PMc/NO₂/SO₂/O₃ and FeNO was not found
303 in any of the models used in this study. The inclusion of ambient PMc, gaseous pollutant or
304 covariates into the crude model led to small changes (< 20% of the effect estimate) in the
305 associations between FeNO and personal exposures to PM_{2.5}, OC, EC, PAHs, and c-PAHs,
306 respectively. The only exception was that the effect of EC on FeNO decreased from 3.8% to
307 0.8% after being adjusted for ambient PMc. Since all the constituents were significantly
308 correlated with personal PM_{2.5} in concentration (Table 3), their associations with FeNO were
309 also adjusted for personal PM_{2.5} in Two-Pollutant models. As displayed in Table S4 and S6,
310 The inclusion of personal PM_{2.5} into the adjusted model generally led to obvious decrease in
311 the effect estimates of different constituents, while the significance of the association between
312 personal PM_{2.5} and FeNO was barely influenced.

313

314 **4. Discussion**

315 This is the first study to examine the association between personal exposure to PM_{2.5}, their
316 constituents, and respiratory inflammation among healthy adults in Hong Kong. One of the
317 hypotheses of underlying biologic mechanisms responsible for the association is that inhaled
318 particles can rapidly react with extracellular macromolecules or cell constituents in the airway
319 epithelium to generate reactive oxygen species and lipid peroxidation products (Auger et al.,
320 2006). These products further induce local and systemic oxidative or nitrosative stress and
321 subsequent inflammation. NO in human body is generated from the oxidation of L-arginine to
322 L-citrulline by nitric oxide synthase (NOS), which is released by many cells in the lung and up-
323 regulated by cytokines (Redington et al., 2001). In this study, statistically significant and
324 positive association between FeNO and personal PM_{2.5} was found in all the models adjusted for
325 potential confounders. Previous studies using FeNO as an outcome in healthy adults are quite
326 limited. Two previous panel studies with similar sample sizes reported significant associations,
327 and higher effect estimates on FeNO (29% in non-smoking seniors in the USA; 40.7% in non-
328 smoking adults in China (Adar et al., 2007; Zhang et al., 2013). Since FeNO levels may

329 influenced by a lot of time-dependent factors, several earlier studies using cross-sectional study
330 design found insignificant results. Other possible explanations for the discrepancy include but
331 not limited to data scarcity, population susceptibility and different chemical profiles (Boogaard
332 et al., 2013b; Kubesch et al., 2015; Meier et al., 2014).

333 In comparison with ambient concentrations, personal exposures to PM_{2.5} yielded much more
334 significant and robust association with inflammatory biomarker in this study, suggesting that
335 ambient PM_{2.5} concentrations from monitoring stations may not be an appropriate proxy for
336 actual PM_{2.5} exposures. On the other hand, the higher effect estimates associated with personal
337 PM_{2.5} exposure could also be attributed to the different sources and composition profile of
338 personal PM_{2.5} compared with ambient PM_{2.5}. Indoor sources of PM_{2.5} can increase the
339 percentage of OC, EC, and the toxicity of exposure. According to the results of this study,
340 unemployed subjects and office workers were exposed to relatively higher average percentages
341 of OC and EC, which suggests the existence of indoor sources. Cooking, smoking and incense
342 burnings are typical indoor sources related to Chinese culture and living styles (Lung et al.,
343 2007). Previous studies also reported that printing could significantly increase the concentration
344 of PM_{2.5} in the offices at Guangzhou, China (Zhang et al., 2017). On the other hand, poor
345 ventilation is another factor contributing to the moderate correlation coefficient between
346 personal exposures and ambient levels. It is suspected that poor ventilation of the subjects'
347 offices or potential indoor sources (e.g. office printer, second hand smoke) led to the
348 significantly higher personal-to-ambient PM_{2.5} concentration ration for office workers in this
349 study. Therefore, the correlations between personal exposure and ambient levels on weekdays
350 and weekends for all the subjects were compared (Figure 5). It was found that personal and
351 ambient PM_{2.5} levels were better correlated on weekends ($r = 0.52, p < 0.001$) than on weekdays
352 ($r = 0.37, p < 0.001$) and that the correlation on weekdays for office workers was especially
353 weak ($r = 0.18, p = 0.23$) compared with that for students, teachers, and unemployed subjects
354 ($r = 0.623, p < 0.001$ when treated as one group). This result supports our speculation and the
355 health risks caused by poor ventilation or potential indoor sources from office should not be
356 neglected.

357 The associations between FeNO and several PM_{2.5} constituents were also examined in this
358 study. Although c-PAHs only accounted for less than 0.01% of the personal PM_{2.5} mass and
359 42.6% of all the PAHs measured, they showed significant association with FeNO levels.
360 According to adjusted models, the effect estimate of total c-PAHs was comparable with that of
361 personal PM_{2.5} and higher than those of other constituents. Among c-PAHs, BbF and BkF had
362 the highest effect estimates, while FLU had the lowest effect estimate. Significant association
363 between EC and FeNO was reported in an earlier study, but it was not found in this one, possibly
364 attributed to our low statistical power for constituents. Total exposure to PAEs had been linked
365 with pulmonary function (Hoppin et al., 2004). However, for PAEs in PM_{2.5}, the effect estimate
366 observed in this study was close to null.

367 Various other kinds of possible confounders were considered and controlled for in the study
368 design procedures or by using mixed-effects models. In a multivariate linear regression analysis,
369 FeNO levels were positively associated with male gender ($p = 0.01$) after adjusting for age,
370 BMI, RH, occupation, income, weekday and PM_{2.5}. The difference between the FeNO levels in
371 healthy males and females was also observed by previous studies and should be paid attention
372 in the future (Bayram et al., 2014; Kim et al., 2010). Confounding by other covariates (age,
373 BMI, education level, occupation, household income, ambient gaseous pollutants and ambient
374 PM_c) on the association between personal PM_{2.5} and FeNO was minimal, suggesting the
375 robustness of our model, and that changes in biomarker levels were unlikely driven by other
376 time-dependent factors. However, residual confounding remains in this study, which could be
377 attributed to the fact that ambient levels of PM_c and gaseous pollutants cannot reflect the actual
378 personal exposures to these pollutants. If enough resources are accessible, pollutants in gas-
379 phase (such as PAHs and formaldehyde) should be measured at individual level and included
380 in the regression model.

381 Several strengths make this study unique and meaningful. 1) The two methods of exposure
382 assessments were used and compared in examining their associations with FeNO levels in
383 residents in Hong Kong. 2) Several kinds of constituents in personal PM_{2.5} were considered
384 simultaneously for comparison. 3) Repeated measures based on a robust number of samples

385 decreased the chance of confounding by time-independent variables as each subject served as
386 his or her control. 4) Comprehensive confounder adjustment was also made by including
387 meteorological variables, demographic and socioeconomic covariates, and gaseous pollutants.
388 5) Estimates for personal exposures to all the pollutants except PAEs were robust to different
389 combinations of covariates and when extreme FeNO values were removed.

390 Limitations of our study include that certain daily activities may have been suppressed by
391 carrying a PM_{2.5} sampler, which may lead to bias in exposure measurement. Nonetheless, all
392 participants were encouraged to conduct their regular study or working activities during the
393 sampling periods. Colinearity is a major issue when studying the effects of different PM_{2.5}
394 constituents. In this study, the correlations between constituents were examined and the results
395 of Two-Pollutant models indicate that the effect of personal PM_{2.5} mass is more robust than that
396 of the constituents monitored. The small number of observation points for constituents
397 measured limited the statistical power to detect meaningful findings, as well as the
398 generalizability of the study results. FeNO is the only biomarker measured in the study, so we
399 cannot directly assess the association between targeted air pollutants and respiratory
400 inflammation. Also, since PM_{2.5} constituents are collinear variables, which might be
401 confounded by each other, the effect estimates could be meaningful for comparison but must
402 be interpreted with caution.

403 In summary, this study offered valuable new information about personal exposure to PM_{2.5}
404 and their constituents as well as their short-term effects on FeNO levels in healthy adults of
405 Hong Kong, in comparison with ambient concentrations. Considering the inconsistent
406 associations found both in our study, and when compared to other published studies, it is
407 important to validate our findings with future research to reach meaningful conclusions.

408

409 **Competing interests**

410 The authors declare that they have no competing interests.

411

412 **Abbreviations:** PM_{2.5}, Fine particle; FeNO, Fractional exhaled nitric oxide; PAHs, Polycyclic
413 aromatic hydrocarbons; IQR, Interquartile range; CI, Confidence interval; c-PAHs,
414 Carcinogenic PAHs

415

416 **Authors' contributions**

417 ZF conceived the study and was involved in the statistical analyses and preparation of the
418 manuscript. VCP was involved in the statistical analyses and revised the manuscript. XCC was
419 involved in the exposure assessment and chemical analysis. QH and LT assisted in the
420 interpretation of the results. SCL and SSH helped to perform the chemical analysis. SLT and
421 KH conceived the study, participated in its design and revise the manuscript. All authors have
422 read and approved the final manuscript.

423

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427

428 **Additional file available**

429 Tables are addressing target species of PAHs and PAEs, subjects' participating schedule,
430 correlations between meteorological variables and air pollutants and sensitivity analysis.

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