The effects of lateral entrainment on pollutant dispersion inside a street canyon 1

- and the corresponding optimal urban design strategies 2
- 3

Zhengtong Li ^a, Hao Zhang ^a, Chih-Yung Wen ^a, An-Shik Yang ^{b,*}, Yu-Hsuan Juan ^b 4

5 ^a Department of Mechanical Engineering and Interdisciplinary Division of Aeronautical and Aviation Engineering, The Hong Kong Polytechnic University, 6 7 Kowloon, Hong Kong

^b Department of Energy and Refrigerating Air-Conditioning Engineering, National Taipei University of Technology, Taipei 106, Taiwan, ROC

- 8 * corresponding author
- 9

Abstract: 10

11 Intensive traffic emissions have caused many environmental problems and have a negative effect on public health. With the aim of mitigating these problems, it is 12 essential to figure out how the flow structure affects the pollutant dispersion within the 13 14 urban canopy. Most previous studies focus on the canopy vortex caused by top 15 entrainment, but few previous studies are aware of the importance of lateral entrainment. 16 By conducting computational fluid dynamic (CFD) simulations validated by wind tunnel data, we investigate the effects of lateral entrainment on pollutant dispersion 17 inside a street canyon. Eight three-dimensional street canyons with various building 18 19 heights and street lengths are considered. Besides, three optimal design strategies are proposed to improve the air quality by enhancing the lateral entrainment. The results of 20 21 this analysis demonstrate that lateral entrainment could conditionally reduce the pollutant concentration of low-rise canyons. This reduction, which is affected by lateral 22 entrainment, is confined in a range of approximately 2.5 times the street width from the 23 street ends. In contrast, the lateral entrainment causes a more pronounced reduction in 24 the pollutant concentrations of the high-rise canyons. Besides, all three strategies can 25 considerably facilitate the lateral entrainment, leading to a significant reduction in the 26 27 cross-section pollutant concentrations (by up to 76%) and therefore a significant reduction in the personal intake fraction *P IF* of the residents (by up to 81%). 28

29

Keywords: Lateral entrainment; urban geometry; air quality; computational fluid 30 dynamics; urban design strategies 31

1 1. Introduction

The ongoing global urbanization and accompanying intensive traffic emissions 2 3 have given rise to many climatic and environmental problems [1], which significantly affect public health (e.g., respiratory and lung diseases) [2] and even cause considerable 4 economic loss [3]. Evidently, air pollution has become one of the main concerns in the 5 world, especially in metropolises. To alleviate the urban air pollution problem, one of 6 the effective methods is to control the flow structure within urban areas [4,5]. This is 7 8 because once these traffic emissions are discharged into the atmosphere, their dispersion is notably affected by the flow structure [6,7]. 9

In most of the previous studies dealing with pollutant dispersion in canyons, the 10 length of the street canyon was assumed to be infinite when the ambient wind was 11 perpendicular to the street axis [8–11]. Consequently, the flow structure is mainly 12 influenced by the top entrainment at the roof level of the infinite-long street canyon. At 13 the roof level, a strong shear layer is developed due to the flow separation at the 14 windward building edge [12]. From the building roof, the fresh air is entrained into the 15 16 street canyon to form a clockwise recirculation with a horizontal (spanwise) axis (canyon vortex), which occupies the entire space of the street canyon [13]. Once the 17 pollutants are emitted from street-level vehicles, most of the pollutants follow this 18 canyon vortex, hence causing a higher concentration at the leeward side of the street 19 canyons [14]. Then, the upward flow near the leeward surface will transfer to the 20 external flow at the roof level. Since the length of the street canyon is assumed to be 21 22 infinite, this clockwise recirculation in any cross-section of the street canyon is identical. In effect, the length of the street canyon is finite [15,16]; thus, the lateral 23 24 entrainment exists at the street ends. The two-dimensional (2D) simulations that

consider only the top entrainment could not completely reflect the flow topology and pollutant dispersion processes in the entire street canyon [17]. At the same time, the influence of the lateral entrainment on the pollutant dispersion inside the street canyon has been confirmed in early studies. In a finite-long 3D street canyon, as seen in Fig. 1, the canyon vortex caused by the top entrainment usually appears nearby the center-

plane of the street canyon [18]. However, at the ends of the street of a regular street 1 canyon with H/W = 1, Hunter et al. [19] and Leitl and Meroney [20] found that there 2 are double-eddy circulations (corner vortexs) with a vertical axis, entraining fresh air 3 from the lateral shear layer (Fig. 1). Accordingly, the developed flow regime consists 4 of a canyon vortex (caused by top entrainment) in the inner area and of two corner 5 vortices (caused by lateral entrainment) at the street ends [21,22]. A superposition of 6 the canyon vortex and the corner eddies results in a helical flow structure within the 7 8 street canyon. Furthermore, Tsai et al. [23] reported that this helical flow structure 9 causes along-street channeling flows toward the symmetry plane (the mid-plane in the spanwise direction) of the street canyons. Hence, the highest pollutant concentrations 10 are on the symmetry plane. From this plane, the pollutant concentration decays 11 12 symmetrically toward the street ends. Additionally, Gromke et al. [21] explained that a decrease in the pollutant concentration at the street ends is related to the enhanced 13 ventilation by laterally entrained air. Accordingly, lateral entrainment can significantly 14 affect the spanwise distribution of pollutants along a low-rise street canyon. 15



16

Fig. 1 Schematic illustration of the canyon vortex caused by the top entrainment and corner vortex caused by the lateral entrainment within a 3-D regular street canyon (H/W=1) and subjected to perpendicular approaching wind.

Interestingly, the effectiveness of lateral entrainment on the flow in the street canyon has been found to vary with different configurations of street canyons. In the low-rise street canyon, Hunter et al. [19] found that the corner eddies caused by lateral entrainment were always maintained at the street ends, despite the increase in the street length. Similarly, Soulhac et al. [24] confirmed that these two corner eddies always

penetrate toward the inner street-canyon area at approximately the order of the street 1 width. Accordingly, the importance of the canyon vortex and the corner eddies changes 2 with the building length (L) when the building height (H) and street width (W) are fixed. 3 For example, Vardoulakis et al. [25] reported that in relatively low-rise short canyons 4 (L/W=3), the influence of the corner vortices might be strong enough to inhibit a stable 5 vortex that is perpendicular to the street in the mid-plane of the spanwise direction. 6 However, Mei et al. [26] showed that the influence of the corner eddies could be 7 8 neglected when L/W is larger than 20 for a low-rise street canyon (H/W= 1). Furthermore, in a high-rise street canyon (H/W=2), the threshold of L/W is 70 to neglect 9 the corner eddies. 10

There is a large body of literature on the effects of top entrainment on the pollutant 11 12 concentrations within the street canyon, while only a few studies address the effects of lateral entrainment. Most of these existing studies attempted to determine the situations 13 when the lateral entrainment can be neglected in numerical simulations to simplify the 14 3D model to the 2D model [26]. Few of them could unravel the range and degree of the 15 16 influence of lateral entrainment on the air quality when the geometry of the street canyons changes. Therefore, there is still a lack of sufficient understanding to fully 17 utilize the positive effects of lateral entrainment on the air quality inside the canyons 18 for practical urban planning. 19

20 On the other hand, the fact that enhanced ventilation by laterally entrained air causes lower pollutant concentrations at the street ends has been confirmed before. It 21 22 can be inferred that enhancing the influence of lateral entrainment by certain optimal urban design strategies can effectively improve the dilution potential of air pollutants 23 24 within the street canyon. However, as reviewed by previous literature, best urban design strategies focused on the enhancement of canyon vortexes by the top entrainment, such 25 as a step-up street canyon [15,27], "lift-up" design (podium) [28], and arcade design 26 [29,30]. Less attention has been paid to fully utilizing the positive effects of lateral 27 entrainment on pollutant concentrations in terms of optimal urban design strategies. 28

29 In general, few previous studies are aware of the importance of lateral entrainment

on the pollutant dispersion within street canyons, especially for the deep street canyon. 1 Indeed, the distribution of pollutant inside canyons can be very sensitive to the lateral 2 entrainment. Moreover, so far, the quantitative analysis of the influence of lateral 3 entrainment is rare. Besides, previous studies have not determined how to effectively 4 utilize the lateral entrainment to improve the air quality within the urban canopy. All 5 these impose the need for investigating the effects of lateral entrainment on pollutant 6 dispersion inside a street canyon and the corresponding optimal urban design strategies. 7 8 Given this background, the objectives of this study are (1) to elucidate the mechanisms for how lateral entrainment affects the pollutant concentrations in the canyons with 9 different geometries (different building heights and lengths), (2) to quantify the 10 influence of lateral entrainment on the reduction of pollutant concentrations, for the 11 12 canyons with different geometries (different building heights and lengths), compared with the infinite-long canyons alternative, and (3) to explore several optimal design 13 strategies for improving the air quality within the street canyons by enhancing the 14 lateral entrainment. 15

In section 2, a description of wind-tunnel experiments is presented. In section 3, the simulation details of the CFD setup are described, including the case study, model description, boundary conditions, numerical method, and grid sensitivity analysis. In section 4, we validate the present computational model with the turbulence modeling tested. In section 5, the mechanisms for how lateral entrainment affects the pollutant concentrations are discussed. Then, several optimal design strategies are proposed by utilizing lateral entrainment. Finally, the conclusions are given in section 6.

23

24 **2. Description of wind tunnel experiments**

Validation is obligatory to determine the accuracy and reliability of the results of CFD simulations [31]. The current computational model to reproduce the flow and concentration fields within street canyons was justified by a wind tunnel experiment conducted earlier at the Laboratory of Building and Environmental Aerodynamics, University of Karlsruhe [32,33]. The wind tunnel had a test section of 2 m long, 2 m

wide, and 1 m high (Fig. 2(a)), in which a scaled model (1:150) of a three-dimensional 1 isolated street canyon constructed by two parallel model-buildings with the dimension 2 of $H \times W_b \times L = 0.12 \text{ m} \times 0.12 \text{ m} \times 1.2 \text{ m}$ (Fig. 2(b)) was tested. Meanwhile, the street 3 width W is equal to the building width W_b . This isolated street canyon was simulated in 4 a neutral atmospheric boundary layer (ABL) by using the vortex generators and a 5m 5 long fetch covered with roughness elements (Fig. 2(a)). This combination produced a 6 simulated boundary layer with a power-law exponent α of 0.30 and a friction velocity 7 u_{ABL}^* of 0.52 m/s. The mean streamwise velocity profile of the approaching flow in the 8 9 upstream can be approximated by using the following power-law form

10 $U(z) = U_{ref} \times (z/H)^{0.3},$ (1)

where $U_{ref} = 4.7$ m/s is the reference velocity of the incoming flow at z = H with a 11 Reynolds number of approximately 37,600, based on the building height H and the 12 reference velocity U_{ref} . Besides, Sulfur hexafluoride (SF6) was used as a tracer gas for 13 simulating the release of traffic exhaust fumes and was emitted homogenously by four 14 15 line-sources mounted at the bottom of the model. To account for the traffic exhaust fumes released on the street intersections, each line source exceeded the street canyon 16 by approximately 10% on each side. For more information related to the wind-tunnel 17 experiments, the reader is referred to [32,33]. Besides, it should be mentioned that the 18 19 aforementioned wind tunnel experiment mainly offers concentration data within street canyons, including the canyon with trees and the canyon without trees. Herein, the free-20 tree case was chosen for the validation study. 21



Fig. 2 Schematics of (a) test section of the wind tunnel, and (b) wind tunnel model of
the urban street canyon (scale 1:150) [32,33]

3 4

3. Description of CFD simulations

5 **3.1. Description of case studies, computational geometry and grid**

The street canyon configurations used in the CFD simulations were constructed 6 7 based on the scaled model (1:150) of an isolated street canyon adopted in the wind tunnel experiment mentioned above. Besides the configuration studied by the wind 8 9 tunnel experiment, seven more configurations with various height and length aspect ratios, which are defined as H/W (= 1 and 3) and L/W (= 1, 5, 10, and ∞), were 10 considered to investigate the effects of the lateral entrainment (Fig. 3(b)). These eight 11 12 street canyons were first divided into two groups according to the aspect ratio of the building height to the street width (H/W), namely, the low-rise street canyons (H/W)13 1) and high-rise street canyons (H/W=3). Additionally, in each group, four aspect ratios 14 of the building length to the street width (L/W) were considered, namely, the short street 15 canyon (L/W = 1), the medium street canyon (L/W = 5), the long street canyon (L/W = 5)16

1 10), and the infinite-long street canyon $(L/W = \infty)$, according to the classification of 2 Oke et al. [34].

3 The size and discretization of the computational domain were referred from the practice guidelines by Tominaga et al. [35], thus the distances between the building and 4 the inlet boundary, lateral boundaries, top boundary, and outflow boundary were 5 H, 5 5 H, 5 H, and 15 H, respectively, as shown in Fig. 3(a). The computational domain was 6 discretized into approximately 2.8 million hexahedral cells for the low-rise medium 7 street canyon (H/W = 1 and L/W = 5), as shown in Fig. 4. Considering the relatively 8 large gradients of the velocity near the ground and building surfaces, the finest grids 9 were deployed around these two types of walls. In this study, a grid-sensitivity analysis 10 was performed based on two additional grids: a coarser grid and a finer grid for the low-11 rise medium street canyon case (Fig. A2). For the coarse, basic, and fine grids, the 12 minimum sizes were set to be 0.006 m, 0.003 m, and 0.0015 m, respectively. The total 13 cell numbers for the coarse, basic, and fine grids are 0.74 million, 2.83 million, and 14 9.66 million, respectively. Therefore, the ratios of the two consecutive cell numbers for 15 16 the grid refinement meet the criterion of 3.4 in the mesh-independent study [35]. Then, the results of grid-sensitivity analysis discussed later indicate that the basic grid 17 provides nearly grid-independent results, which can be further used for the remainder 18 of this study. 19





1 Momentum equation:

2
$$\frac{\partial u_i u_j}{\partial x_j} = -\frac{1}{\rho} \left(\frac{\partial p}{\partial x_i} \right) + \frac{\partial \tau_{ij}}{\partial x_j}, \qquad (3)$$

3 where the stress tensor τ_{ij} is defined as:

4
$$\tau_{ij} = \rho \left[v_i \left(\frac{\partial u_i}{\partial x_j} + \frac{\partial u_j}{\partial x_i} \right) \right] - \frac{2}{3} \rho k \delta_{ij}, \qquad (4)$$

where the term u_i denotes the *i*-axis component of the air velocity; p and ρ represent the
pressure and density; v_i is the turbulent kinematic viscosity; δ_{ij} is the Kronecker delta;
k is the turbulence kinetic energy.

8 The species transport equation was solved to probe the pollutant dispersion in an 9 urban environment, as follows:

10
$$\frac{\partial u_i Y}{\partial x_i} - \frac{\partial}{\partial x_i} \left[\left(D + D_i \right) \frac{\partial Y}{\partial x_i} \right] = S_p, \qquad (5)$$

where S_p is the pollutant source term (kg/(m³·s)); D and D_t (= v_t/Sc_t) denote the molecular and turbulent diffusion coefficients of the pollutant, respectively. Sc_t is the turbulent Schmidt number, which was set to 0.4 to account for the underestimation of the turbulent mass diffusion from the RANS models [36,37]. *Y* is the mass fraction of the pollutants. This dispersion of pollutants was simulated with the User Defined Scalar (UDS) option in ANSYS Fluent.

Moreover, the renormalization group (RNG) k- ε model [38] is chosen because of 17 its generally good performance in predicting the flow separation by buildings and 18 reversed flow [39], which is essential for the analysis of lateral entrainment in the 19 20 present study. Also, the RNG k- ε model complements the disadvantage of a standard k- ε model, which overestimates turbulent kinetic energy near the edges of buildings where 21 ambient flow impinges and separate [40]. Thus, the RNG k- ε model was used to solve 22 this steady-state isothermal flow field. The conservation equations of the RNG k- ε 23 turbulence model for the turbulence kinetic energy (k) and dissipation rate (ε) are as 24 follows: 25

1
$$\frac{\partial u_j k}{\partial x_j} = \frac{\partial}{\partial x_j} \left(\frac{v_t}{\sigma_k} \frac{\partial k}{\partial x_j} \right) + P_k - \varepsilon$$
(6)

$$\frac{\partial u_j \varepsilon}{\partial x_j} = \frac{\partial}{\partial x_j} \left(\frac{v_t}{\sigma_{\varepsilon}} \frac{\partial \varepsilon}{\partial x_j} \right) + \frac{\varepsilon}{k} (C_{\varepsilon_1}^* P_k - C_{\varepsilon_2} \varepsilon)$$
(7)

3 In this equation,
$$P_k = v_t S^2$$
, $S = \sqrt{2S_{ij}S_{ij}}$, $S_{ij} = \frac{1}{2} \left(\frac{\partial u_i}{\partial x_j} + \frac{\partial u_j}{\partial x_i} \right)$, $v_t = C_\mu \frac{k^2}{\varepsilon}$, $\sigma_k = 1$,

4
$$C_{\varepsilon_1}^* = 1.42 - \frac{\eta(1 - \eta/4.38)}{1 + 0.012\eta^3}, \ \eta = \frac{k}{\varepsilon}S, \ C_{\varepsilon_2} = 1.68, \text{ and } \sigma_{\varepsilon} = 0.719$$

5 **3.3. Boundary conditions**

6 The measured inlet velocity profile from the wind tunnel experiments [32], which
7 is given in Eq. (1), was used to characterize a neutral ABL. The turbulent kinetic energy
8 *k* and turbulence dissipation rate *ε* profiles were calculated using Eqs. (8) and (9) [41]:

$$k = \frac{(u_{ABL}^{*})^{2}}{\sqrt{C_{\mu}}},$$
(8)

9

10

2

$$\mathcal{E} = \frac{(u_{ABL})^3}{\kappa(z+z_0)},\tag{9}$$

11 where u_{ABL}^* is the ABL friction velocity (= 0.52 m/s), κ is the von Karman's constant 12 (= 0.42), z_0 is the aerodynamic roughness (= 0.0015 m), and C_{μ} is the model constant 13 (= 0.085).

Besides, as seen in Fig. 3(a), the top and lateral boundaries of the domain were set 14 as symmetry boundaries, namely setting normal velocity and normal gradients of all 15 variables to zero. On the outlet of the domain, a zero diffusive flux was imposed for all 16 17 flow variables in the direction normal to the outflow plane since the domain downstream was long enough to ensure a fully developed outlet flow. The standard wall 18 functions by Launder and Spalding [42] with and without roughness modification by 19 Cebeci and Bradshaw [43] were applied at the ground surface and building surface, 20 21 respectively. To reduce horizontal inhomogeneity, the sand grain roughness height k_s is calculated by the roughness constant C_s (= 9.9) and the aerodynamic roughness z_0 (= 22 0.0015 m) in Eq. (10) [44]. 23

$$k_s = \frac{9.793z_0}{C_s} , \tag{10}$$

2 Besides, CO was used as the pollutant representative. To calculate the CO concentration, a uniform volume source (width $W_p = 0.8$ W and length L_p = street length 3 L) of CO was specified near the ground with a depth of 0.1 H to represent the traffic 4 5 lanes, as shown in Fig. 3(a). The constant emission rate per hour and unit street length (36.1 g/h/m, i.e., total mass release rate of $L_p \times 1.0 \times 10^{-5}$ kg/s) was adopted for each 6 7 CO source with reference to Ng and Chau [45]. Considering the type and number of vehicles passing by a realistic street per hour in Mongkok, Hong Kong, Ng and Chau 8 [45] summarized the pollutant release rate above. 9

10 **3.4. Solver settings**

1

11 The commercial software ANSYS/Fluent[®] CFD software (Release 15.0) [46] was 12 used to simulate the airflow of ambient wind over this isolated street canyon. This study 13 utilized the pressure-linked equations-consistent (SIMPLEC) numerical method for the 14 pressure-velocity coupling. The second-order upwind scheme [47] was used to 15 discretize both the convective terms and the diffusion terms. A double-precision solver 16 was also selected for the CFD calculations. The convergence criterion of the normalized 17 residual errors was set to 10^{-6} for the governing equations.

18 **3.5 Grid sensitivity analysis**

19 Three types of meshes were tested for the low-rise medium canyon under the same environmental conditions ($U_{ref} = 4.7 \text{ m/s}$). Fig. 5 and 6 (a)-(c) depict a comparison of 20 the results for the dimensionless streamwise mean velocity (u/U_{ref}) and dimensionless 21 pollutant concentration (K) (defined in section 3.6) on the three grids along three 22 23 vertical lines (x = -0.25H, 0, and 0.25H) in the vertical center plane at y/H = 0. Along these lines, the fine and the basic grid provide almost identical results, while some 24 deviations are found between the coarse and the basic grid. Then, the grid convergence 25 Index (GCI) proposed by Roache [48] (Eqs. (11) and (12)) is used to estimate the error 26 of u/U_{ref} and K on the basic grid. 27

$$GCI_{u} = F_{s} \left| \frac{r^{p} (u_{basic} - u_{fine}) / U_{ref}}{1 - r^{p}} \right|$$

$$\tag{11}$$

$$GCI_{K} = F_{s} \left| \frac{r^{p} (K_{basic} - K_{fine}) / K_{fine}}{1 - r^{p}} \right|$$

$$(12)$$

1

Where F_s is the safety factor taken as 1.25 when three or more grids are compared, 3 r is the linear grid refinement (= $\sqrt{2}$), p is the former order of accuracy (=2), u and K 4 are streamwise mean velocity and normalized concentration in one of the two grids 5 (basic and fine), and U_{ref} is the reference wind speed of 4.7 m/s. The values of the GCI_u 6 averaged along each vertical line are 0.04% for x/H = -0.25, 0.06% for x/H = 0, and 7 0.08% for x/H = 0.25 (Fig. 5 (d)-(f)). Similarly, the values of the GCI_K averaged along 8 each vertical line are 1.80% for x/H = -0.25, 1.85% for x/H = 0, and 4.04% for x/H =9 0.25 (Fig. 6 (d)-(f)). By analyzing the discrepancy in wind speed and pollutant 10 concentration of the three grids as well as comparing GCI values of the Fine and Basic 11 grids, it can be concluded that the basic grid provides nearly grid-independent results, 12 13 which can be further used for the remainder of this study. Besides, the near-wall area was resolved by the standard wall functions directly on the condition that the v+ (a 14 dimensionless wall distance to judge the applicability of wall functions [49], 15 $y + = u_T y / v$, where u_T is the friction velocity, y is the absolute distance from the wall, 16 17 and ν is the kinematic viscosity) of the first near-wall mesh for building surfaces and ground was 167.7 on average, which was in the log-law layer 30 < y + < 300 [46,50]. 18





Fig. 5 (a-c) Comparison of dimensionless streamwise mean velocity (u/U_{ref}) along three vertical lines inside the street canyon in the vertical center plane in coarse, basic, and fine grids; (d-f) grid-convergence index (GCI) along the same three vertical lines.



9

Fig. 6 (a-c) Comparison of dimensionless pollutant concentration *K* along three
vertical lines inside the street canyon in the vertical center plane in coarse, basic, and
fine grids; (d-f) grid-convergence index (GCI) along the same three vertical lines.

5 **3.6 Air quality indices**

6 **3.6.1 Dimensionless pollutant concentration**

7 In this study, the concentration of pollutants is mainly presented as dimensionless
8 concentration *K*,

$$K = \frac{CU_{ref} HL}{S_p V_p}$$
(13)

10 Where *C* is the local pollutant concentration (kg/m³); V_p is the volume of pollutant 11 source (m³).

12 **3.6.2** Average dimensionless pollutant concentration

13 The lateral entrainment can significantly affect the spanwise distribution of 14 pollutant concentrations along the street length. Therefore, the pedestrian-level and 15 cross-section average dimensionless pollutant concentrations were introduced to better 1 evaluate the effects of the lateral entrainment.

2 The average dimensionless pollutant concentration at the pedestrian-level (z = 1.5
3 m at full scale) along the street width was calculated by Eq. (14),

$$K_{ped} = \frac{\int_0^W K dx}{W} \tag{14}$$

5 The cross-section average dimensionless pollutant concentration along the street 6 length was calculated in Eq. (15),

7
$$K_{cross} = \frac{\int_0^W \int_0^H K dx dz}{W \times H}$$
(15)

8 **3.6.3** Personal intake fraction (*P_IF*)

4

9 This study utilizes the personal intake fraction (P_IF) as the air quality index, 10 which stands for a fraction of the total traffic exhaust inhaled by each person on average, 11 which was first introduced by Hang et al. [51] into CFD simulations to quantify the 12 average personal exposure.

13 It is defined and calculated as follows:

14
$$P_{IF} = \frac{\sum_{i=j}^{N} \sum_{j=1}^{M} P_i \times Br_{i,j} \times \Delta t_{i,j} \times Ce_j / m}{\sum_{j=1}^{M} P_i}$$
(16)

where N is the number of population groups (children, adults, elders, N = 3, i = 1 to 3), 15 M is the number of different microenvironments (indoors at home, other indoor 16 locations, near-vehicle locations, and other outdoor locations away from vehicles, M= 17 4, j=1 to 4). Moreover, we assumed the following: the near-road buildings were 18 residential, and only a microenvironment of j=1 (indoor at home) was considered to 19 20 assess the personal intake fraction for the local residents. $Br_{i,j}$ and $\Delta t_{i,j}$ are the average volumetric breathing rate (m^3/s) [51] (Table A1) and time spent (s) for individuals in 21 the *i*th population group in the *j*th microenvironment [52] (Fig. A1(a)), respectively. P_i 22 is the total number of people exposed in the *i*th population group, which can be further 23 calculated by the demographic structure (herein, taking Shenzhen, China, as an example 24 for this study [52], Fig. A1(b)). Ce_j is the pollutant concentration in the *j*th 25

1 microenvironment (kg/m³), which could be calculated from the average concentration 2 at each floor (3 m). In this instance, *m* is the total pollutant emissions (kg).

3 4. Validation study

Before validation study and case study, a simulation was conducted with an empty 4 computational domain to check the achievement of the horizontal homogeneity of ABL, 5 6 since it is a prerequisite to a reliable prediction of pollutant dispersion within street canyons [53]. First, the inlet boundary conditions of the CFD simulation based on the 7 8 experimental data (described in section 2) fit the inflow wind profile of the wind tunnel. 9 Fig. A3 then shows a check of horizontal homogeneity for the present CFD simulation, which compared the dimensionless streamwise velocity and dimensionless turbulence 10 kinetic energy of the inlet profile and incident profile (at the building position). The 11 comparison indicates that the development of horizontal inhomogeneity is insignificant. 12

Besides, a solid model that included the street canyon (H/W = 1 and L/W = 10) was 13 created by replicating the details of the geometrical shape from the wind tunnel 14 experimental set-up of the tree-free case [32,33]. The computational domain was in line 15 16 with the CFD set-up for the case study and pollutant sources were consistent with the wind tunnel setting. Moreover, the computational grid resolution resulted from a grid-17 sensitivity analysis, which yielded a fully structured hexahedral grid with 4.68 million 18 19 cells. Then, a cross-comparison of the dimensionless vertical velocity at the y/L = 0 and the dimensionless pollutant concentration at the walls of the street canyon between the 20 numerical and experimental results was presented in Fig. 7(a) and (b). The 21 concentration value was calculated in the non-dimensional form as $C_{+} = \frac{CU_{ref}H}{O/l}$, 22

where *C* is the measured concentration (g/m³), and *Q/l* is the tracer gas source strength per unit length (g/m/s). Generally, the experimental and numerical distributions of the dimensionless vertical wind speed were consistent (Fig. 7(a)). Only on the windward side, the RNG *k*- ε turbulence model predicted slightly higher flow velocities. Then, two *Sc_t* (*Sc_t* = 0.4 and 0.7) were tested. As seen in Fig. 7 (b), the predicted dimensionless concentrations are similar to those obtained in the wind tunnel for both *Sc_t*. Nevertheless,

- 1 the numerical results agree better with the wind tunnel data when $Sc_t = 0.4$. The RNG
- 2 k- ε model with $Sc_t = 0.4$ was consequently adopted for our CFD simulations.



- 5 Fig. 7 Comparison results of (a) dimensionless vertical velocity at y/L = 0, and (b) 6 dimensionless pollutant concentration at the walls of the street canyon.
- 7 5. Results and discussion

4

- 8 5.1 Effects of the lateral entrainment
- 9 5.1.1 Low-rise street canyon (H/W=1)

10 The effects of the lateral entrainment on the low-rise street canyons (H/W=1) are 11 explored in this section. Figures 8 and 9 show the predicted dimensionless pollutant 12 concentration K and dimensionless wind velocity (U/U_{ref}) contours for different street configurations at the pedestrian level and cross-section, respectively. To quantitatively
 estimate the effects, Fig. 10 compares the average pollutant concentrations at the
 pedestrian level and at various cross-sections.

In a low-rise infinite-long street canyon (H/W=1 and $L/W=\infty$), the flow structure 4 within the canyon was affected only by the top entrainment at the roof level. Thus, as 5 seen in Fig. 9(a), the whole street canyon was occupied by the y-axis vortex. Evidently, 6 this y-axis vortex in any cross-section of the infinite-long canyon would be identical, 7 8 namely, a clockwise canyon vortex. Moreover, as evidenced in Fig. 8(a) and Fig. 9(a), there was relatively strong ventilation within this low-rise infinite-long street canyon 9 since the top entrainment can readily penetrate into the ground level, thus leading to a 10 lower dimensionless pollutant concentration at the pedestrian level ($K \le 66.5$). In effect, 11 12 in addition to the influence of the top entrainment, the airflow within the finite-long street canyon is also significantly affected by the lateral entrainment, which will be 13 discussed later. 14

In a low-rise short street canyon (H/W=1 and L/W=1) in Fig. 9(b), the flow 15 16 structure (3D streamlines) was still mostly dominated by the y-axis vortex caused by the top entrainment, which was similar to the infinite-long street canyon in Fig. 9(a). In 17 contrast, the pedestrian level was occupied by the outward airflows (the along-street 18 channeling flows toward the street ends) (Fig. 8(b)). The possible reason was that the 19 lateral entrainment caused a pair of corner vortices at the pedestrian level of the street 20 ends. Following these corner vortices, the wind flowed outward along the leeward side 21 22 of the upwind building. As a result of this marked outward airflow, the maximum pedestrian level concentration was notably reduced by almost 65% (Fig. 8(b)), and the 23 24 average pedestrian level concentration decreased by 65-73% along the street length, as well (Fig. 10(a)), compared with the infinite-long street canyon. At the same time, as 25 seen in Fig. 9(b), the wind velocity was enhanced at all of the cross-sections. The 26 upward transportation of the pollutants was also improved, which led to a significant 27 reduction in the concentration at all of the cross-sections. Figure 10(b) further 28 29 confirmed that the average cross-section concentration remarkably decreased by almost

1 73-77% along the street length due to the lateral entrainment.

In a low-rise medium street canyon (H/W=1 and L/W=5), there existed two evident 2 corner vortices at the street ends (Fig. 8(c)), although the 3D streamlines were still 3 dominated by the y-axis vortex (Fig. 9(c)). On the other hand, except for the region 4 covered by the corner vortices, as shown in Fig. 8(c), the whole pedestrian level was 5 mainly occupied by the inward flows (along-street channeling flows toward the 6 symmetry plane). In effect, these inward channeling flows can be attributed to the 7 8 superposition of the canyon vortex (caused by the top entrainment) and the corner vortex (caused by the lateral entrainment). These inward channeling flows enhanced 9 the pedestrian level dimensionless wind velocity (up to nearly 0.1) and then transported 10 the pollutants toward the symmetry plane (Fig. 8(c)). Consequently, as seen in Fig. 10 11 (a), in almost 65% of the region of the street canyon, the concentration significantly 12 decreased, especially at the street ends (by up to 70%). However, the concentration 13 increased in the remaining 35% region of the street canyon near the symmetry plane 14 (by up to 86%). This trend occurred because the accumulating pollutants in the canyon-15 16 center region caused by the inward channeling flow could not be dispersed upward effectively along with the canyon vortex, which was further confirmed in Fig. 9(c). 17 Clearly, the wind velocity in most inner section of this medium street canyon was lower 18 than that in the infinite-long case, thus leading to a significant increase in the pollutant 19 concentration, especially near the ground. Therefore, Fig. 10 (b) reported that the cross-20 section average concentration declined by up to 78% from y/L=0.17 to 0.5, while it 21 increased by up to nearly 117% from v/L=0 to 0.17. 22

In a low-rise long street canyon (H/W= 1 and L/W= 10), similar to the medium street canyon, the corner vortices and the inward channeling flow were clearly observed in Fig. 8(d). In contrast, the inward channeling flow only penetrated approximately 2.5 times the street width from the street ends (Fig. 8(d)). Furthermore, in the inner region of the canyon (from y/L= 0.3 to 0), the airflow was almost dominated by the *y*-axis vortex (canyon vortex), which was similar to the infinite-long street canyon (from the windward side of the downwind building to the leeward side of the upwind building).

Accordingly, the pedestrian level dimensionless velocity was significantly enhanced by 1 up to 0.17 from only y/L=0.5 to 0.25, but it decreased by approximately 0.1 in the inner 2 region of the canyons (Fig. 8(d)), compared with the infinite-long street canyon. As a 3 result, the average pedestrian level concentration decreased by up to almost 78% in the 4 outer region of the canyon, but it significantly increased in the inner region of the 5 canyon, especially from y/L=0.28 to 0 (even by almost 50%) (Fig. 10(a)). On the other 6 hand, as seen in Fig. 9(d), a lower wind velocity was found in the inner three sections, 7 8 which further deterred the upward dispersion of the pollutants. Obviously, as shown in Fig. 9(d) and Fig. 10(b), the cross-section average concentration increased by up to 34%, 9 from y/L = 0.27 to 0. 10

Overall, it is concluded that in the low-rise street canyon (H/W=1), the lateral 11 12 entrainment can partially improve the air quality, depending on the street length. This finding occurred because the positive effects of lateral entrainment on the pollutant 13 concentration inside the street canyon were confined in a range of approximately 2.5 14 times the street width from the street ends (Fig. 8(d)). Therefore, the lateral entrainment 15 16 was of great importance in reducing the pollutant concentration of the short and medium street canyon. However, the air quality of the low-rise long canyon could only be 17 improved in the outer half of the street length by the lateral entrainment. 18







Fig. 9 Predicted 3D streamlines, *x-z* cross-section pollutant concentration and wind velocity contours for the low-rise street canyons (H/W=1) with various street

lengths: (a) $L/W=\infty$, infinite-long street canyon, (b) L/W=1, short street canyon,



(c) L/W= 5, medium street canyon, and (d) L/W= 10, long street canyon.



1

2



Fig. 10 Average pollutant concentration along half of the street length for the low-rise
street canyon (*H/W*= 1) with various street lengths: (a) average pedestrian level
pollutant concentration, and (b) average cross-section pollutant concentration

7

5.1.2 High-rise street canyon (H/W=3)

8 To discuss the influence of the lateral entrainment on high-rise street canyons, the 9 numerical results on the dimensionless wind velocity and pollutant concentration are 10 presented at the pedestrian level in Fig. 11 and for various *x-z* cross-sections in Fig. 12; 11 at the same time, the average pollutant concentrations for different street lengths are 12 compared in Fig. 13.



Within a high-rise infinite-long street canyon (H/W=3 and $L/W=\infty$), as seen in Fig.

12(a), the top entrainment from the roof level induced two vertically aligned vortices 1 (y-axis vortex). Consequently, Fig. 12(a) showed that it was pretty difficult for the top 2 3 entrainment to penetrate downward into the pedestrian level where the traffic emission sources were located. The airflow within the lower two y-axis canyon vortices was too 4 slow (dimensionless wind velocity < 0.02) to generate any upward pollutant dispersion 5 (Fig. 12(a)). In other words, the upward advective transport of the airflow had little 6 contribution to the dispersion process of the pollutants (transporting the pollutant from 7 8 the lower recirculation to the upper recirculation and eventually toward the roof level). Accordingly, it was evident that there existed a substantial pollutant accumulation in 9 the lower part of the high-rise canyon (Fig. 11(a) and Fig. 12(a)). The highest pollutant 10 concentration in the lower space was almost one order higher than that at the roof level. 11 12 Interestingly, these results were inconsistent with the field measurements by Zhang et al. [54] in a similar deep street canyon ($H/W \approx 2.7$ and $L/W \approx 10$). As reported by Zhang 13 et al. [54], the highest low-level concentration was only two times higher than the roof-14 level concentration. Thus, it could be deduced that the airflow was possibly sensitive to 15 16 the lateral entrainment within a finite-long high-rise street canyon, which might extensively promote the pollutant dispersion in the lower space. 17

In the high-rise short street canyon (H/W=3 and L/W=1) in Fig. 12(b), the top 18 entrainment produced two separated y-axis vortices compared with the infinite-long 19 counterpart. Interestingly, at the pedestrian level, there were noticeable divergent and 20 outward airflows caused by lateral entrainment (Fig. 11(b)). Although the pedestrian 21 22 level wind speed was still relatively low, the pollutants could be effectively transported outward along with these divergent flows. In consequence, as seen in Fig. 13(a), the 23 24 average pedestrian level concentration was reduced by almost up to 98% along the street length, compared with the infinite-long counterpart. In other words, the lateral 25 entrainment could affect the pollutant dispersion of whole street canyon. Furthermore, 26 the wind velocity and flow patterns of the various x-z cross-sections showed that the 27 canyon was almost occupied by the strong downward airflows (Fig. 12(b)). At the same 28 29 time, the dimensionless wind velocity at the cross-section was enhanced by up to 0.5 (Fig. 12(b)). Accordingly, the pollutants only slightly accumulated near the ground level
 (Fig. 12(b)). Figure 13(b) reported that the average concentration of the *x-z* section
 decreased by up to 99% along the street length.

3

In a high-rise medium street canyon (H/W=3 and L/W=5), similar to the short 4 canyon, the top entrainment still caused a y-axis vortex near the roof level (Fig. 12(c)). 5 In contrast, the lateral entrainment produced two symmetric spanwise recirculation in 6 the lower space. As evidenced in Fig. 11(c), the outer regions (0.35 < y/L < 0.5) were 7 8 dominated by the outward airflows at the pedestrian level. Instead, in the inner regions (0 < y/L < 0.35), there existed the inward channeling airflows. Therefore, a higher 9 pollutant concentration was found near the symmetry plane. Compared with the 10 infinite-long case (Fig. 11(a)), the medium street still had a markedly smaller magnitude 11 12 of the concentration with the same level of wind velocity (Fig. 11(c)). As also shown in Fig. 13(a), the average pedestrian level concentration was reduced by 81 - 98%. A 13 possible explanation lies in the stronger advective transport of pollutants provided by 14 the x-axis recirculation, which could be substantiated in the flow patterns and wind 15 16 velocity contours of various x-z cross-sections (Fig. 12(c)). In the inner two sections, there was an upward airflow with a higher wind velocity; thus, the pollutants within the 17 street center region could be more substantially transported out across the roof level. 18 Compared to the results of the infinite-long case, the pollutant concentration decreased 19 remarkably due to the lateral entrainment for all of the x-z cross-sections (Fig. 12(c)), 20 and the average concentration along the street length was reduced by 75 - 98% (Fig. 21 22 13(b)).

As shown in Fig. 12(d), in a high-rise long street canyon (H/W= 3 and L/W= 10), the flow patterns were slightly different from those in the short and medium street canyons. Nevertheless, the *y*-axis and *x*-axis vortex dominated the pollutant transport in the upper and lower spaces, respectively. Furthermore, the *x*-axis vortex was elongated. This *x*-axis vortex caused clear inward channeling flows at the pedestrian level, transporting most of the pollutants toward the symmetry plane and leading to a more significant pollutant accumulation in the street center region, compared with the

shorter canyons (short and medium street canyons). As shown in Fig. 13(a), although 1 the maximum pedestrian level concentration of the long street canyon was up to 10 2 3 times higher than its shorter counterparts, this maximum value was still much lower than the infinite-long street result. On the other hand, as seen in Fig. 12(d), although 4 the wind velocity in the lower part of most inner section began to be stagnant, the x-5 axis vortex could still reinforce the upward advective transportation of the pollutants. 6 Therefore, the maximum average cross-section concentration was almost three times 7 8 lower than the infinite-long counterpart (Fig. 13(b)). Besides, this high-rise long street canyon (H/W=3 and L/W=10) shares a similar configuration with the study of Zhang 9 et al. [54]. The measurement position by Zhang et al. [54] is nearly 0.3L away from the 10 street ends. The low-level concentration was two times higher than the roof-level 11 12 concentration. In the present study (Fig. 12 (c)), the low-level concentration of the second section (y/L=1/4) from the street ends was nearly 3-4 times higher than the roof-13 level concentration. The difference between this CFD simulation and field 14 measurement is reasonable, and they are in the same order. Notably, the realistic traffic-15 16 induced turbulence [55], solar radiation [56], and building separation [45] (were not considered in the present study) can also improve the pollutant dispersion, especially 17 for the low-space of street canyon. 18

In general, in the high-rise street canyons, the lateral entrainment can reduce the 19 pollutant concentration more significantly, compared with the low-rise street canyons. 20 The reason is that the lateral entrainment can entirely affect the x-axis vortex/ 21 22 recirculation in the lower part of the canyon; hence, it can increase the vertical advective transportation of the pollutants in the canyon's center region. With an increase in the 23 24 street length, the flow patterns remained unchanged, with the dominated y-axis vortex 25 in the most upper space and x-axis vortex/recirculation in the lower space, respectively, but the influence of lateral entrainment on the pollutant concentration became weaker. 26 Despite this effect, the concentrations for the short, medium, and long canyons were 27 still far lower than that of the infinite canyon. In consequence, these phenomena 28 29 demonstrated that the lateral entrainment significantly contributed to the pollutant dispersion for these high-rise street canyons. Moreover, compared with the low-rise canyon, the high-rise canyon has a significantly higher concentration, especially for the longer street length. Taking the infinite-long canyon as examples, the pedestrian level average pollutant concentration of high-rise canyon (=769.3) is about 20 times that of the low-rise canyon (=36.7), which is in line with the study of Assimakopoulos et al. [11]. Accordingly, it indicates that weaker top entrainment in high-rise canyon greatly limits the dilution of pollutants.



9 Fig. 11 Predicted 3D streamlines, pedestrian level pollutant concentration contours,
10 and pedestrian level wind velocity contours for the high-rise street canyons
11 (*H/W*= 3) with various street lengths: (a) *L/W*= ∞, infinite-long street canyon,
12 (b) *L/W*= 1, short street canyon, (c) *L/W*= 5, medium street canyon, and (d)
13 *L/W*= 10, long street canyon.



Fig. 12 Predicted pollutant concentration and wind velocity contours at different *x-z* cross-sections for the high-rise street canyons (*H/W*= 3) with various street
 lengths: (a) *L/W*= ∞, infinite-long street canyon, (b) *L/W*= 1, short street canyon,
 (c) *L/W*= 5, medium street canyon, and (d) *L/W*= 10, long street canyon.





8

Fig. 13 Average pollutant concentration along half of the street length for the high-rise street canyon (H/W= 3) with various street lengths: (a) average pedestrian level pollutant concentration and (b) average cross-section pollutant concentration

9

5.2 Optimal urban design strategies for lateral entrainment

10 As discussed in the last section, the lateral entrainment can effectively improve the 11 dilution potential of the pollutants inside the street canyon, especially for the deep 12 canyons. In the low-rise medium street canyon (H/W= 1 and L/W= 5), the lateral 13 entrainment caused the corner vortex at the street ends, and then, it contributed to the

dilution of the pollutants near the ground level. In the high-rise street canyon (H/W=31 and L/W=5), the lateral entrainment had a more profound impact on the flow structure 2 compared with the low-rise canyon, thus creating the x-axis recirculation in the lower 3 space of the canyons. As discussed above, this x-axis recirculation can effectively 4 improve the advective transport of the pollutants in the lower space. In summary, it 5 might be useful to further improve the dilution potential of the pollutants by enhancing 6 the intensity of the corner vortex in the low-rise street canyon or the x-axis recirculation 7 8 in the high-rise street canyon. Therefore, three attempts have been made to enhance the influence of the lateral entrainments, i.e., the corner-trim of the downwind building, the 9 short upwind building, and the lower height at the ends of the upwind building. In this 10 section, the low-rise (H/W=1) and high-rise (H/W=3) canyons with the medium-long 11 12 street (L/W=5) were considered to be the base cases to enhance the improvement on the pollutant concentration reduction. Also, the influence of dimensions of the corner-trim 13 of downwind building D_{trim}, intended length of upwind building L_{intended}, and reduced 14 height of upwind building Hreduced has been examined in Fig. A4 to A6. It is suggested 15 16 that even the relatively minor optimal design can effectively improve the ventilation and the potential of pollutant dilution inside street canyons. As length limits, only the 17 cases of $D_{trim} = 0.5 W$, $L_{intended} = 0.5 W$, and $H_{reduced} = 0.5 H$ were discussed in detail. 18

19

5.2.1 Design I: Corner-trim of the downwind building

The first attempt is to trim the corner of the downwind building, thus creating a 20 "venturi effect" at the street ends. The dimensions of the trimmed corner are shown in 21 Fig. 14 (a). A comparison of the results of the corner-trim and base cases in the low-rise 22 street canyon is also presented. Notably, the maximum pedestrian level dimensionless 23 24 wind velocity increased by approximately 0.2, although the flow structure changed only slightly. Hence, as illustrated in Fig. 14 (a), the concentration decreased in most of the 25 canyons. This corner-trim design also significantly reduced the highest concentration 26 of the base case in the canyon center region by up to almost 36% (Fig. 15(a)). On the 27 other hand, in the vertical direction, this design also caused a significant reduction in 28 29 the leeward side (leeward side P IF reduced by almost 11%, Fig. 15(b)) since this design significantly enhanced the vertical ventilation on this side. Also, it slightly led
 to slightly lower windward *P_IF* by about 30- 100.

3 As seen in Fig. 14 (b), the corner-trim design had more significant implications for the reduction of the concentration in the high-rise case (H/W=3). The "venturi effect" 4 at the street ends caused a strong inward channeling flow toward the symmetry plane. 5 6 Further, the intensity of the x-axis recirculation was also enhanced. The pedestrian level dimensionless wind velocity increased substantially by 0.5 (Fig. 14 (b)). Consequently, 7 8 the pedestrian level concentration decreased in most of the canyons as a result (Fig. 14 9 (b)). At the same time, as shown in Fig. 14 (b), the base case had a higher concentration 10 at both the street ends and symmetry plane. This corner-trim design can effectively reduce the pollutant concentration in these two regions (by up to almost 63%) (Fig. 14 11 12 (b) and Fig. 15 (a)). Additionally, as shown in Fig. 15 (c), the lower-story residents (level 2 to level 6) suffered high P IF in the base case. The corner-trim design relieved 13 this situation and effectively reduced leeward P IF by up to 78% for the lower-story 14 residents, although it slightly increased windward P IF by 80-150. 15





- Fig. 14 Cross-comparison of the pollutant concentration at the pedestrian level between
 the base case and the corner-cut design case: (a) low-rise street canyon and (b)









(b) *P_IF*, low-rise street canyon
Fig. 15 Cross-comparison of the pollutant concentration and personal intake fraction *P_IF* along the street length between the base case and corner-trim design
case: (a) Average cross-section pollutant concentration along the street length,
(b) *P_IF*, low-rise street canyon, and (c) *P_IF*, high-rise street canyon

1

To enhance the influence of the lateral entrainment, the second attempt is to shorten 8 9 the length of the upwind building by 1/2 W. Upon shortening the upwind building of the low-rise street canyon, as seen in the 3D streamlines of Fig. 16 (a), the lateral 10 incoming wind flowed over the side of the upwind building, and it hit the windward 11 surface of the downwind building. Then, the incoming wind flowed toward the 12 symmetry plane, hence leading to a considerable increase in the dimensionless wind 13 velocity at the pedestrian level (by up to 0.3). Correspondingly, the pedestrian level 14 concentration was reduced significantly, especially in the canyon center region. In 15 addition, as evidently shown in Fig. 17 (a), there was a large decrease in the cross-16 section pollutant concentration from y/L=0.35 to 0 (by up to 45%). In addition, this 17 18 design substantially reduced both leeward and windward P IF from level 1 to level 6 (up to 49%) (Fig. 17(b)). 19

For the high-rise street canyon (H/W=3) with the short upwind building, as Fig. 16 (b) shows, the lateral incoming wind also hit the ends of the windward surface of the downwind building, and then, it enhanced the flow intensity of the *x*-axis recirculation. Therefore, there existed a strong upward airflow at the symmetry plane and a strong

outward airflow at the pedestrian level (Fig. 16 (b)). The pedestrian level dimensionless 1 2 wind velocity was remarkably improved (by up to 0.2). As a result, the accumulated pollutants in the canyon center region can easily escape from the street canyon across 3 the street lateral boundaries. As shown in Fig. 17 (a), the concentration reduced in 4 almost all of the street's length, especially in the canyon center region (by up to 76%). 5 Additionally, by enhancing the intensity of the x-axis recirculation, this design 6 appreciably reduced the windward P IF by up to 44-69% and the leeward P IF by 71-7 8 81%, especially for the lower-story residents, who were always suffering the worst air 9 quality.









(a) Average cross-section pollutant concentration along the street length





7 5.2.3 Design III: Lower height at the ends of the upwind building

1

As discussed in section 5.2.2, design II successfully introduced the lateral incoming flow into the street canyon from the street ends, hence increasing the flow strength of the *x*-axis recirculation and reducing the pollutant concentration. However, design II will be at the cost of a lower building coverage ratio. Thus, the third attempt is to explore whether only lowering the building height at the ends of the upwind building can also improve the ventilation in the same way (Fig. 18(a) and (b)).

For the low-rise street canyon with design III (the indented length = 1/2 W and the 14 reduced height = 0.5 H), as seen in the 3D streamlines of Fig. 18 (a), the pedestrian 15 16 level wind velocity was markedly improved by introducing fresh air from the upper part of the lateral street boundaries, the same as in design II. Therefore, the cross-section 17 pollutant concentration also reduced substantially from y/L=0.4 to 0 (by up to 34%) 18 (Fig. 19 (a)). Clearly, for the low-rise canyons, the reduction in the pedestrian level 19 concentration due to design III was only slightly lower than in the design II counterpart. 20 In terms of P IF (Fig. 17(b)), this design also reduced leeward P IF from level 1 to 21 level 6 (up to 39%), but the windward P IF had little change. 22

For the high-rise street canyon with design III (the indented length = 1/2 W and the

reduced height = 1/3 H), similar to design II, the wind velocity was also improved (Fig. 18 (b)), although the increment in the wind velocity is less than that of design II (Fig. (b)). Therefore, as shown in Fig. 19 (a), there was a considerable decrease in the pedestrian level concentration along the street length, especially in the canyon center region (by up to 71%). Furthermore, the *P_IF* noticeably declined at both the leeward side (by 55-73%) and the windward side (by up to 56-73%).



Lower height at the ends of the upwind building













8 6.1 Conclusions

9 This paper has presented numerical simulations with Computational Fluid Dynamics (CFD) to investigate the influence of the lateral entrainment on the pollutant 10 concentration within the street canyons based on eight 3D street canyons with different 11 12 aspect ratios of building height and length to street width (H/W) and L/W under the perpendicular wind. The simulations were based on grid-sensitivity analysis and 13 validation of the CFD results from the literature. Based on the CFD results, the 14 importance of the lateral entrainment was confirmed. Further, three designs were 15 proposed to improve air quality by enhancing the influence of lateral entrainment. The 16 17 major results are summarized as follows:

(1) In a low-rise street canyon, the flow structure was mainly dominated by the top
entrainment. The lateral entrainment slightly altered the flow, except for the
appearance of the corner vortex and the inward channeling flow near the ground
level. Thus, the positive effect of the lateral entrainment on the pollutant
concentration was limited (approximately 2.5 times the street width from the
street end). For example, the lateral entrainment can significantly reduce both

the cross-section and pedestrian level pollutant concentrations of the short and
 medium canyon by up to 78%. However, for the long canyon, these two
 concentrations declined only for the outer half of the street in the length
 direction due to the lateral entrainment.

- 5 (2) In a high-rise street canyon, the top entrainment caused only one canyon vortex 6 (the axis was parallel to the street length) in the upper space. In contrast, the 7 lateral entrainment dominated the lower space by the two symmetric 8 vortices/recirculation with an axis perpendicular to the street length. Thus, the 9 pollutant concentrations markedly decreased by up to almost 99% for the short, 10 medium, and long street canyons, due to the lateral entrainment, compared with 11 the case of only considering the top entrainment.
- (3) All of the three optimal designs were considerably useful in reducing the 12 13 pollutant concentration by enhancing the lateral entrainment, especially for the high-rise street canyons. First, the corner-trim of the downwind building 14 created a "venturi effect" at the street ends, thus significantly reducing the 15 cross-section concentration in most regions of the street canyon (by up to 16 almost 36% and 63% for the low-rise and high-rise canyons) and the personal 17 intake fraction P IF (by up to almost 11% and 78% for the low-rise and high-18 rise canyons). Second, the short upwind building notably introduced the 19 incoming wind impinging at the ends of the downwind building; thus, the cross-20 section concentration greatly decreased in both the low-rise canyons (by up to 21 45%) and the high-rise canyons (by up to 76%). In addition, it reduced the P IF 22 of the low-rise canyons by up to 49%, and the P IF of the high-rise canyons 23 by up to 81%. Third, the lower height at the ends of the upwind building also 24 introduced fresh air in the same way as in the setup of the short upwind building, 25 alleviating the cost of a lower building coverage ratio (the design III is less 26 27 expensive than the design II). The reduction in the concentrations caused by 28 this design was just slightly lower than the setup of a short upwind building. By discussing those results, two suggestions can be proposed for sustainable street 29

design to reduce pollutant concentration inside street canyons. First, the importance of 1 lateral entrainment should not be neglected, especially for the high-rise street canyon. 2 Therefore, less blockage should be achieved at the street ends. In other words, at the 3 street ends, large-size advertisement boards [57] should not be installed or trees with 4 large-size canopy and high leaf area density [58,59] should not be planted. Second, to 5 utilize the lateral entrainment to improve the ventilation within street canyons, it might 6 be feasible to create a corner trim of downwind buildings at the street ends. Also, the 7 8 upwind buildings of street canyon should be shorter than the downwind buildings.

9 6.2 Limitations

Despite the obtained findings, the present study had several limitations. First, only 10 one boundary condition (perpendicular wind with constant wind speed) was considered, 11 12 but in actual practice, ambient wind could flow into the urban configuration from all directions with different wind speeds. It is expected that the minor variation of wind 13 direction might significantly alter the lateral entrainment. Second, this study is limited 14 to isothermal conditions. In effect, the buoyancy effect due to temperature differences 15 16 between heated building surfaces/ground and ambient air should not be ignored, which is an essential driving force for natural ventilation, especially when the wind is 17 relatively weak. For instance, the heated leeward surface by solar radiation can cause a 18 stronger upward airflow in the center of the canyon [15], which might influence lateral 19 entrainment. Third, this study only used a generic street canyon configuration. However, 20 real urban areas consist of irregular arrangements of buildings of various heights. 21 Irregular arrangements of buildings may modify the influence of lateral entrainment. 22 Finally, only the RANS model and the RNG k-ɛ model were used to predict the urban 23 24 wind flow under steady-state conditions. It is known that the RANS models provide effective time-averaged flow solutions. The dispersion contribution of transient 25 fluctuations, however, are compromised. Additionally, even with a lower turbulent 26 Schmidt number (= 0.4) set to account for the underestimation of the turbulent mass 27 diffusion from the RANS model, the accumulation of pollutants near the ground level 28 29 could be overestimated, especially for the high-rise street canyon. In contrast, the large

eddy simulation (LES) model can provide more accurate results due to its better capture 1 of flow intermittencies and separations around the street canyon. However, due to the 2 high computation cost and challenges in selecting sub-grid scale models with 3 appropriate boundary conditions by using the LES model, many studies still select the 4 RANS approach instead [58]. In this regard, a more accurate RANS turbulence model, 5 such as Shear-Stress Transport (SST) k-w model or Reynolds Stress Model (RSM), 6 should be adopted to evaluate the influence of lateral entrainment under unsteady-state 7 8 conditions. Generally, all these factors could influence the lateral entrainment 9 investigated in this study, and this will be explored in our future studies.

10

11 **REFERENCE**

- 12 [1] S. Grimmond, Urbanization and global environmental change: local effects of urban warming, Geogr. J. 13 173 (2007) 83-88. 14 R. Rückerl, A. Schneider, S. Breitner, J. Cyrys, A. Peters, Health effects of particulate air pollution: a [2] 15 review of epidemiological evidence, Inhal. Toxicol. 23 (2011) 555-592. 16 [3] Y. Xia, D. Guan, X. Jiang, L. Peng, H. Schroeder, Q. Zhang, Assessment of socioeconomic costs to 17 China's air pollution, Atmos. Environ. 139 (2016) 147-156. 18 doi:https://doi.org/10.1016/j.atmosenv.2016.05.036.
- [4] A.-S. Yang, Y.-H. Juan, C.-Y. Wen, C.-J. Chang, Numerical simulation of cooling effect of vegetation
 enhancement in a subtropical urban park, Appl. Energy. 192 (2017) 178–200.
 doi:https://doi.org/10.1016/j.apenergy.2017.01.079.
- [5] A.-S. Yang, C.-Y. Wen, Y.-H. Juan, Y.-M. Su, J.-H. Wu, Using the central ventilation shaft design within
 public buildings for natural aeration enhancement, Appl. Therm. Eng. 70 (2014) 219–230.
 doi:https://doi.org/10.1016/j.applthermaleng.2014.05.017.
- P. Edussuriya, A. Chan, A. Ye, Urban morphology and air quality in dense residential environments in
 Hong Kong. Part I: District-level analysis, Atmos. Environ. 45 (2011) 4789–4803.
 doi:https://doi.org/10.1016/j.atmosenv.2009.07.061.
- [7] Z. Li, T. Ming, S. Liu, C. Peng, R. de Richter, W. Li, H. Zhang, C.-Y. Wen, Review on pollutant
 dispersion in urban areas-part A: Effects of mechanical factors and urban morphology, Build. Environ.
 (2020) 107534.
- S.-J. Mei, J.-T. Hu, D. Liu, F.-Y. Zhao, Y. Li, H.-Q. Wang, Airborne pollutant dilution inside the deep
 street canyons subjecting to thermal buoyancy driven flows: Effects of representative urban skylines,
 Build. Environ. 149 (2019) 592–606. doi:https://doi.org/10.1016/j.buildenv.2018.12.050.
- W.C. Cheng, C.-H. Liu, D.Y.C. Leung, Computational formulation for the evaluation of street canyon
 ventilation and pollutant removal performance, Atmos. Environ. 42 (2008) 9041–9051.
- 36 [10] D. Marucci, M. Carpentieri, Effect of local and upwind stratification on flow and dispersion inside and
 37 above a bi-dimensional street canyon, Build. Environ. 156 (2019) 74–88.
- [11] V.D. Assimakopoulos, H.M. ApSimon, N. Moussiopoulos, A numerical study of atmospheric pollutant
 dispersion in different two-dimensional street canyon configurations, Atmos. Environ. 37 (2003) 4037–
 40

1 2	[12]	S.E. Belcher, Mixing and transport in urban areas, Philos. Trans. R. Soc. A Math. Phys. Eng. Sci. 363 (2005) 2947–2968.			
3	[13]	F. Caton, R.E. Britter, S. Dalziel, Dispersion mechanisms in a street canyon, Atmos. Environ. 37 (2003			
4	F1 41	693-702.			
5	[14]	A. Di Bernardino, P. Monti, G. Leuzzi, G. Querzoli, Pollutant fluxes in two-dimensional street canyons,			
6		Urban Clim. 24 (2018) 80–93.			
1	[15]	Z. Li, H. Zhang, CY. Wen, AS. Yang, YH. Juan, Effects of height-asymmetric street canyon			
8		configurations on outdoor air temperature and air quality, Build. Environ. (2020) 107195.			
9	[16]	Z. Li, H. Zhang, CY. Wen, AS. Yang, YH. Juan, Effects of frontal area density on outdoor thermal			
10		comfort and air quality, Build. Environ. 180 (2020) 107028.			
11		doi:https://doi.org/10.1016/j.buildenv.2020.107028.			
12	[17]	W.G. Hoydysh, W.F. Dabberdt, Kinematics and dispersion characteristics of flows in asymmetric street			
13		canyons, Atmos. Environ. 22 (1988) 2677–2689.			
14	[18]	C. Gromke, B. Ruck, Influence of trees on the dispersion of pollutants in an urban street canyon-			
15		experimental investigation of the flow and concentration field, Atmos. Environ. 41 (2007) 3287-3302.			
16	[19]	L.J. Hunter, G.T. Johnson, I.D. Watson, An investigation of three-dimensional characteristics of flow			
17		regimes within the urban canyon, Atmos. Environ. Part B. Urban Atmos. 26 (1992) 425-432.			
18		doi:https://doi.org/10.1016/0957-1272(92)90049-X.			
19	[20]	B.M. Leitl, R.N. Meroney, Car exhaust dispersion in a street canyon. Numerical critique of a wind tunnel			
20		experiment, J. Wind Eng. Ind. Aerodyn. 67 (1997) 293-304.			
21	[21]	C. Gromke, B. Ruck, Pollutant concentrations in street canyons of different aspect ratio with avenues of			
22		trees for various wind directions, Boundary-Layer Meteorol. 144 (2012) 41-64.			
23	[22]	C. Gromke, N. Jamarkattel, B. Ruck, Influence of roadside hedgerows on air quality in urban street			
24		canyons, Atmos. Environ. 139 (2016) 75-86. doi:https://doi.org/10.1016/j.atmosenv.2016.05.014.			
25	[23]	M.Y. Tsai, K.S. Chen, Measurements and three-dimensional modeling of air pollutant dispersion in an			
26		Urban Street Canyon, Atmos. Environ. 38 (2004) 5911–5924.			
27	[24]	L. Soulhac, V. Garbero, P. Salizzoni, P. Mejean, R.J. Perkins, Flow and dispersion in street intersections,			
28		Atmos. Environ. 43 (2009) 2981–2996.			
29	[25]	S. Vardoulakis, B.E.A. Fisher, K. Pericleous, N. Gonzalez-Flesca, Modelling air quality in street canyons:			
30		a review, Atmos. Environ. 37 (2003) 155–182. doi:https://doi.org/10.1016/S1352-2310(02)00857-9.			
31	[26]	SJ. Mei, Z. Luo, FY. Zhao, HQ. Wang, Street canyon ventilation and airborne pollutant dispersion: 2-			
32		D versus 3-D CFD simulations, Sustain. Cities Soc. 50 (2019) 101700.			
33		doi:https://doi.org/10.1016/j.scs.2019.101700.			
34	[27]	Z. Li, T. Shi, Y. Wu, H. Zhang, YH. Juan, T. Ming, N. Zhou, Effect of traffic tidal flow on pollutant			
35		dispersion in various street canyons and corresponding mitigation strategies, Energy Built Environ. 1			
36		(2020) 242–253. doi:https://doi.org/10.1016/j.enbeny.2020.02.002.			
37	[28]	Y. Du, C.M. Mak, Y. Li, A multi-stage optimization of pedestrian level wind environment and thermal			
38	[-]	comfort with lift-up design in ideal urban canyons. Sustain Cities Soc. 46 (2019) 101424.			
39		doi:https://doi.org/10.1016/i.scs.2019.101424.			
40	[29]	Y -H. Juan, A -S. Yang, C -Y. Wen, Y -T. Lee, P -C. Wang, Optimization procedures for enhancement of			
41	[=>]	city breathability using arcade design in a realistic high-rise urban area. Ruild Environ 121 (2017) 247–			
42		261. doi:https://doi.org/10.1016/i.buildeny 2017.05.035			
⊥∠ ⊿२	[30]	C-Y Wen Y-H Juan A-S Yang Enhancement of city breathability with half open spaces in ideal			
<u>-</u> -Ο ΔΛ	[30]	urban street canyons Build Environ 112 (2017) 322-336			
		aroun succe canyons, bund. Environ. 112 (2017) 322-330.			

1		doi:https://doi.org/10.1016/j.buildenv.2016.11.048.			
2	[31]	J.I. Perén, T. van Hooff, B.C.C. Leite, B. Blocken, CFD simulation of wind-driven upward cross			
3		ventilation and its enhancement in long buildings: Impact of single-span versus double-span leeward			
4		sawtooth roof and opening ratio, Build. Environ. 96 (2016) 142-156.			
5	[32]	C. Gromke, B. Ruck, On the impact of trees on dispersion processes of traffic emissions in street canyons,			
6		Boundary-Layer Meteorol. 131 (2009) 19-34.			
7	[33]	C. Gromke, CODASC: a database for the validation of street canyon dispersion models, in: 15th Int.			
8		Conf. Harmon. within Atmos. Dispers. Model. Regul. Purp., 2013.			
9	[34]	T.R. Oke, Street design and urban canopy layer climate, Energy Build. 11 (1988) 103-113.			
10	[35]	Y. Tominaga, A. Mochida, R. Yoshie, H. Kataoka, T. Nozu, M. Yoshikawa, T. Shirasawa, AIJ guidelines			
11		for practical applications of CFD to pedestrian wind environment around buildings, J. Wind Eng. Ind.			
12		Aerodyn. 96 (2008) 1749-1761. doi:https://doi.org/10.1016/j.jweia.2008.02.058.			
13	[36]	Y. Tominaga, T. Stathopoulos, Turbulent Schmidt numbers for CFD analysis with various types of			
14		flowfield, Atmos. Environ. 41 (2007) 8091-8099. doi:https://doi.org/10.1016/j.atmosenv.2007.06.054.			
15	[37]	J. Hang, Y. Li, R. Buccolieri, M. Sandberg, S. Di Sabatino, On the contribution of mean flow and			
16		turbulence to city breathability: the case of long streets with tall buildings, Sci. Total Environ. 416 (2012)			
17		362–373.			
18	[38]	V. Yakhot, L.M. Smith, The renormalization group, the $\epsilon\text{-expansion}$ and derivation of turbulence models,			
19		J. Sci. Comput. 7 (1992) 35-61.			
20	[39]	JJ. Kim, JJ. Baik, A numerical study of the effects of ambient wind direction on flow and dispersion in			
21		urban street canyons using the RNG k-ε turbulence model, Atmos. Environ. 38 (2004) 3039–3048.			
22		doi:https://doi.org/10.1016/j.atmosenv.2004.02.047.			
23	[40]	JJ. Kim, DY. Kim, Effects of a building's density on flow in urban areas, Adv. Atmos. Sci. 26 (2009)			
24		45–56.			
25	[41]	P.J. Richards, R.P. Hoxey, Appropriate boundary conditions for computational wind engineering models			
26		using the k-ɛ turbulence model, in: Comput. Wind Eng. 1, Elsevier, 1993: pp. 145-153.			
27	[42]	B.E. Launder, D.B. Spalding, The numerical computation of turbulent flows, in: Numer. Predict. Flow,			
28		Heat Transf. Turbul. Combust., Elsevier, 1983: pp. 96–116.			
29	[43]	T. Cebeci, P. Bradshaw, Momentum transfer in boundary layers, Hemi. (1977).			
30	[44]	B. Blocken, T. Stathopoulos, J. Carmeliet, CFD simulation of the atmospheric boundary layer: wall			
31		function problems, Atmos. Environ. 41 (2007) 238-252.			
32	[45]	WY. Ng, CK. Chau, A modeling investigation of the impact of street and building configurations on			
33		personal air pollutant exposure in isolated deep urban canyons, Sci. Total Environ. 468-469 (2014) 429-			
34		448. doi:https://doi.org/10.1016/j.scitotenv.2013.08.077.			
35	[46]	A. Fluent, ANSYS fluent theory guide 15.0, Inc, Canonsburg, PA. (2013).			
36	[47]	T. Barth, D. Jespersen, The design and application of upwind schemes on unstructured meshes, in: 27th			
37		Aerosp. Sci. Meet., 1989: p. 366.			
38	[48]	P.J. Roache, Quantification of uncertainty in computational fluid dynamics, Annu. Rev. Fluid Mech. 29			
39		(1997) 123–160.			
40	[49]	H. Schlichting, K. Gersten, Boundary-layer theory, Springer, 2016.			
41	[50]	R. Zhang, Y. Zhang, K.P. Lam, D.H. Archer, A prototype mesh generation tool for CFD simulations in			
42		architecture domain, Build. Environ. 45 (2010) 2253-2262.			
43	[51]	J. Hang, Z. Luo, X. Wang, L. He, B. Wang, W. Zhu, The influence of street layouts and viaduct settings			
44		on daily carbon monoxide exposure and intake fraction in idealized urban canyons, Environ. Pollut. 220			

1		(2017) 72-86. doi:https://doi.org/10.1016/j.envpol.2016.09.024.
2	[52]	D. Cui, X. Li, Y. Du, C.M. Mak, K. Kwok, Effects of envelope features on wind flow and pollutant
3		exposure in street canyons, Build. Environ. 176 (2020) 106862.
4		doi:https://doi.org/10.1016/j.buildenv.2020.106862.
5	[53]	H. Yu, J. Thé, Validation and optimization of SST k-ω turbulence model for pollutant dispersion within a
6		building array, Atmos. Environ. 145 (2016) 225-238.
7	[54]	YW. Zhang, ZL. Gu, SC. Lee, TM. Fu, KF. Ho, Numerical Simulation and In Situ Investigation of
8		Fine Particle Dispersion in an Actual Deep Street Canyon in Hong Kong, Indoor Built Environ. 20 (2010)
9		206–216. doi:10.1177/1420326X10387694.
10	[55]	T. Shi, T. Ming, Y. Wu, C. Peng, Y. Fang, R. de_Richter, The effect of exhaust emissions from a group of
11		moving vehicles on pollutant dispersion in the street canyons, Build. Environ. 181 (2020) 107120.
12		doi:https://doi.org/10.1016/j.buildenv.2020.107120.
13	[56]	X. Xie, Z. Huang, J. Wang, Z. Xie, The impact of solar radiation and street layout on pollutant dispersion
14		in street canyon, Build. Environ. 40 (2005) 201-212.
15	[57]	Y. Lin, G. Chen, T. Chen, Z. Luo, C. Yuan, P. Gao, J. Hang, The influence of advertisement boards, street
16		and source layouts on CO dispersion and building intake fraction in three-dimensional urban-like models,
17		Build. Environ. 150 (2019) 297–321.
18	[58]	H. Yang, T. Chen, Y. Lin, R. Buccolieri, M. Mattsson, M. Zhang, J. Hang, Q. Wang, Integrated impacts
19		of tree planting and street aspect ratios on CO dispersion and personal exposure in full-scale street
20		canyons, Build. Environ. 169 (2020) 106529. doi:https://doi.org/10.1016/j.buildenv.2019.106529.
21	[59]	T. Chen, H. Yang, G. Chen, C.K.C. Lam, J. Hang, X. Wang, Y. Liu, H. Ling, Integrated impacts of tree
22		planting and aspect ratios on thermal environment in street canyons by scaled outdoor experiments, Sci.
23		Total Environ. (2020) 142920.
24		



Fig. A1 (a) Time activity patterns for different age groups and microenvironments;
and (b) demographic structure in Shenzhen, China[52].



1 2 3

sensitivity analysis cases



Fig. A3. (a) Comparison of inlet and incident dimensionless streamwise velocity (u/U_{ref}) profiles, (b) Comparison of inlet and incident dimensionless turbulence kinetic energy $(k/(u^*_{ABL})^2)$ profiles, and (c) Schematic cross-section of the domain with location of inlet profile (x/H = 0) and incident profile (x/H = 5).







Fig. A5 Cross-comparison of the dimensionless wind velocity and pollutant concentration at the pedestrian level of a low-rise street canyon (H/W=1 and L/W=5) between the base case and design II case with different indented length Lindented



Fig. A6 Cross-comparison of the dimensionless wind velocity and pollutant concentration at the pedestrian level of a low-rise street canyon (H/W=1 and L/W=5) between the base case and design III case with different reduced height $H_{reduced}$

Table A1 Br	Table A1 Breathing rate for various age groups and microenvironments [51]							
Breathing rate	Indoor at home	Other indoor location	Near vehicle	Other outdoor location				
B_r (m ³ /day)	(j=1)	(<i>j</i> = 2)	(j=3)	(j= 4)				
Children	12.5	14.0	14	18.7				
Adults	13.8	15.5	15.5	20.5				
Elderly	13.1	14.8	14.8	19.5				