

1 **Evaluation of four image fusion NDVI products against in-situ**  
2 **spectral-measurements over a heterogeneous rice paddy**  
3 **landscape**

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## 20 Abstract

21 Satellite image fusion methods that improve spatial and temporal resolution have significant  
22 potential to advance understanding of ecosystem dynamics in space and time. However, systematic  
23 evaluations of image fusion methods against *in situ* spectral data are lacking. Here, we used a suite of  
24 *in situ* spectral data collected at 60 elementary sampling units ( $10 \times 10$  m) covering 15 Landsat pixel  
25 ( $30 \times 30$  m) plots and one Moderate Resolution Imaging Spectroradiometer (MODIS) pixel ( $250 \times$   
26  $250$  m) throughout the entire growing season in a heterogeneous rice paddy landscape to evaluate four  
27 state-of-the-art image fusion NDVI products. They include the Enhanced Spatial and Temporal  
28 Adaptive Reflectance Fusion Model (ESTARFM), Flexible Spatiotemporal DATA Fusion (FSDAF),  
29 SaTellite dATA IntegRation (STAIR), and the CubeSat Enabled Spatio-Temporal Enhancement  
30 Method (CESTEM); the former three blended Landsat and MODIS data, whereas the latter combined  
31 CubeSats, Landsat, and MODIS observations. All fusion products showed strong linear relationships  
32 against *in situ* data when combining all spatial and temporal observations ( $R^2$ : 0.73 to 0.93) although  
33 there were partly negative biases ( $-1\%$  to  $-9\%$ ). These biases resulted from forcing data to image  
34 fusion algorithms, such as Landsat ( $-4\%$ ) and MODIS ( $-7\%$ ). Performance difference between fusion  
35 methods were considerably larger for spatial than for temporal variation. Furthermore, Landsat NDVI  
36 explained only 17–22% of spatial variation against *in situ* spectral data, which can be translated into  
37 weak performance of image fusion products to predict spatial variability in NDVI. Image fusion  
38 products that relied on spatial interpolation showed large biases ( $-15\%$  to  $-30\%$ ) for a vegetation plot  
39 surrounded by mixed land cover plots. Our results highlight key sources of uncertainty and will be  
40 instrumental in improving satellite image fusion methods to monitor land surface phenology in space  
41 and time.

## 42 1. Introduction

43 High-spatiotemporal-resolution remote sensing images have enhanced our ability to monitor  
44 ecosystem dynamics across different scales. However, a fundamental tradeoff between spatial  
45 resolution and temporal revisit frequency limits satellites' ability to observe the Earth's surface at  
46 high spatiotemporal resolution (Gao *et al.* 2006). To overcome this, previous studies developed  
47 satellite image fusion methods that enhanced spatiotemporal resolution by combining temporally  
48 sparse fine-spatial-resolution images with frequent but coarse-spatial-resolution images (Gao *et al.*  
49 2006; Hilker *et al.* 2009a; Luo *et al.* 2018; Zhu *et al.* 2010; Zhu *et al.* 2016). Satellite image fusion  
50 products have been widely used to extract spatial and temporal information on changes in land use  
51 and land cover (Chen *et al.* 2015b; Schmidt *et al.* 2015; Senf *et al.* 2015), classification of  
52 vegetation types (Liu *et al.* 2015), vegetation phenology (Hilker *et al.* 2009b; Walker *et al.* 2014;  
53 Zheng *et al.* 2016), and crop growth situations at field scale (Gao *et al.* 2017a; Kimm *et al.* 2020).  
54 However, the question as to how much we can trust the quality of fusion products over heterogeneous  
55 landscapes remains.

56 The uncertainties associated with fusion products stem from the inconsistency of the input images  
57 and the temporal or spatial interpolation processes (Walker *et al.* 2012). Uncertainties in fusion  
58 products due to input images may be caused by the different spectral response functions of satellite  
59 sensors and sun-target-view geometries (Gao *et al.* 2014; Roy *et al.* 2016; Wang *et al.* 2014).  
60 Moreover, uncertainties surrounding spatial interpolation in fusion products may arise from  
61 assumptions that nearby pixels within the same land cover have similar spectral reflectance patterns  
62 (Zhu *et al.* 2018). Fusion products may average out drastic changes between individual pixels, leading  
63 to blurred images. Furthermore, uncertainties caused by the temporal interpolation process may  
64 emerge due to the time lag between the dates of the input pair and the dates of the predicted fusion  
65 products (hereafter, "time-lag effects"; Fu *et al.* 2015; Olexa and Lawrence 2014). This affects the  
66 performance of temporal interpolation over different land cover types (Emelyanova *et al.* 2013; Olexa  
67 and Lawrence 2014).

68 *In situ* spectral data over space and time allow us to directly quantify the uncertainties of fusion  
69 products as well as assess the uncertainty contribution of input imagery and characterize the time-lag  
70 effects. The quality of satellite image fusion products is typically assessed for a given date by  
71 comparing original fine spatial resolution satellite images with the fusion product developed using  
72 independent data (Emelyanova *et al.* 2013). However, systematically measured *in situ* data—which is  
73 obtained temporally close to the overpass time with accurate and precise geolocation—is essential for  
74 the comprehensive evaluation of image fusion products. To the best of our knowledge, few efforts to  
75 date have tried in this regard (Gao *et al.* 2017a).

76 Ground-based evaluation of fusion products is particularly important in heterogeneous landscapes.

77 In homogeneous landscapes, the development and evaluation of fusion products are relatively  
78 straightforward, as canopy structure and function change only gradually within a subpixel of coarse  
79 spatial resolution imagery. In heterogeneous landscapes, however, abrupt changes in the subpixel over  
80 space and time are a considerable challenge for image fusion methods (Fu *et al.* 2015). For example,  
81 in crop landscapes, various fields may be managed differently for transplanting and harvest dates,  
82 fertilization, or irrigation, which can lead to differences in canopy structure, physiology, and  
83 phenology (Ding *et al.* 2014). Only few studies have evaluated image fusion products over  
84 heterogeneous landscapes such as dryland (Walker *et al.* 2012) and cropland ecosystems (Gao *et al.*  
85 2017b; Zheng *et al.* 2016) by comparing them to independent satellite images or ground-based visual  
86 phenology assessments. To the best of our knowledge, no study has evaluated image fusion products  
87 with *in situ* spectral observations based on a systematic spatio-temporal sampling design over  
88 heterogeneous landscapes.

89 Most fusion products rely on weight function–based methods that combine all input data by  
90 considering spectral differences in input data, time-lag effects, and the distance between the central  
91 pixel in the predicting area and neighboring pixels (Zhu *et al.* 2018). The spatial and temporal  
92 adaptive reflectance fusion model (STARFM) was the original weight function–based method (Gao *et al.*  
93 2006). When each pixel of input data includes only one land cover, STARFM considers changes in  
94 reflectance to be consistent over time. After this initial study, weight function-based methods were  
95 subsequently improved in heterogeneous areas (Zhu *et al.* 2010), in capturing abrupt changes in land  
96 cover (Zhu *et al.* 2016). Moreover, (Luo *et al.* 2018) proposed cloud-free/gap-free daily step data.  
97 Recently, Houborg and McCabe (2018a) introduced a promising spatiotemporal fusion product by  
98 harmonizing the satellite images from CubeSat constellation, albeit with non–weight function–based  
99 methods.

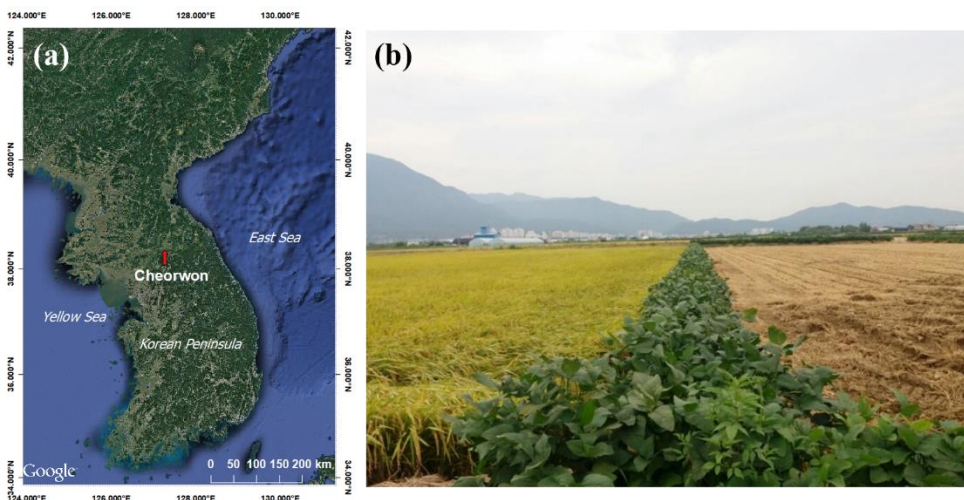
100 The objective of this study was to evaluate the performance of three satellite-based spatiotemporal  
101 fusion products and a novel CubeSat constellation-based fusion product with *in situ* measurements  
102 over a heterogeneous rice paddy landscape. We evaluated the following four image fusion products:  
103 enhanced STARFM (ESTARFM; Zhu *et al.* 2010), Flexible Spatiotemporal DATA Fusion (FSDAF;  
104 Zhu *et al.* 2016), SaTellite dATA Integration (STAIR; Luo *et al.* 2018), and the CubeSat Enabled  
105 Spatio-Temporal Enhancement Method (CESTEM; Houborg and McCabe 2018a). Of these methods,  
106 only CESTEM does not include a spatial gap-filling module. We attempted to answer the following  
107 questions: (i) How do the input data and the fusion algorithm affect the accuracy of fusion products?  
108 (ii) Under which conditions does the spatial or temporal interpolation process affect the performance  
109 of fusion products? (iii) How does heterogeneity affect the performance of fusion products? We report  
110 the results of direct comparisons of each product with *in situ* measurements.

111

112 **2. Methods**

113 **2.1 Study site**

114 The study site was a rice paddy landscape in Cheorwon, Republic of Korea (CRK; 38.2013 N,  
115 127.2507 E), which is part of the Korea Flux Network (KoFlux; Huang *et al.* 2018; Dechant *et al.*  
116 2019; Dechant *et al.* 2019; Hwang *et al.* 2020; Figure 1a). As the site is flat and rice canopies have  
117 relatively low heights, it is suitable for *in situ* spectral measurements at locations determined by  
118 satellite pixel locations. The rice growing season lasts from approximately May to September, and  
119 the predominant species in the study area is *Oryza sativa* L. ssp. *Japonica*. The site has a temperate  
120 monsoon climate with frequent cloud cover and high precipitation from June to August. The size of  
121 the paddy fields ranges from around 2,500 to 4,300 m<sup>2</sup>. Cultivation management includes  
122 irrigation, fertilization, drainage, and harvesting and varies between rice paddies (Figure 1b).

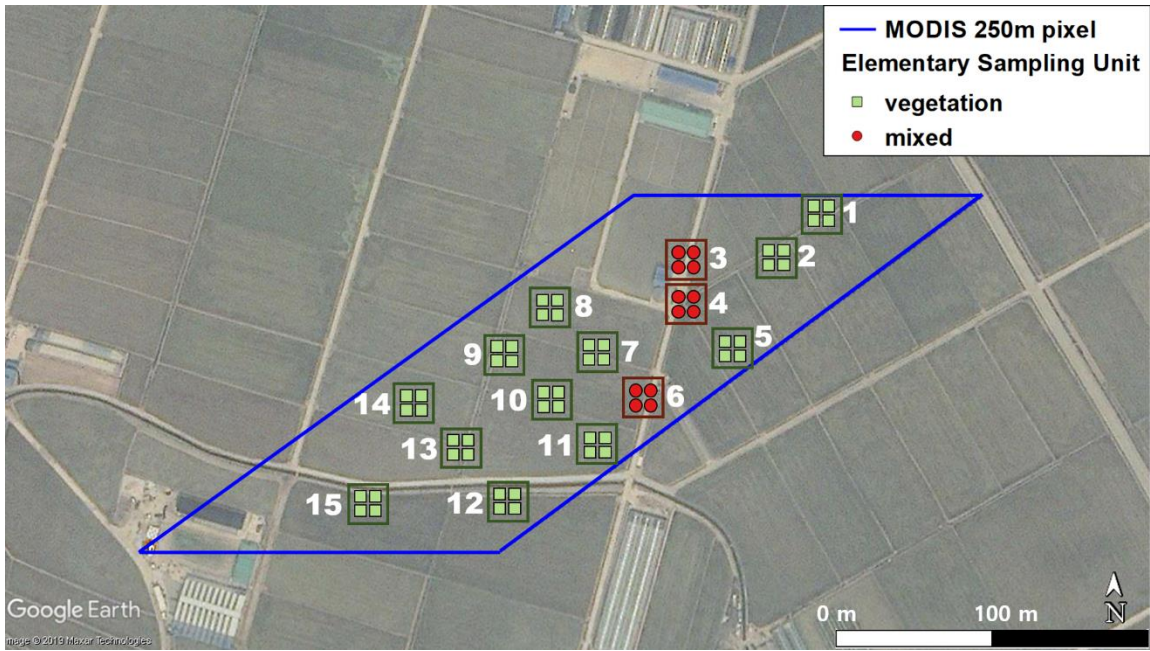


123 **Figure 1 Study site. (a) Map of the Korean Peninsula (image source: Google Earth) with the**  
124 **study site, Cheorwon, marked with a red arrow. (b) This image of the study site was acquired**  
125 **on September 5, 2017, and shows rice paddy fields with different harvest dates bordered by**  
126 **soybean plants.**  
127

128

129 **2.2 In-situ measurements**

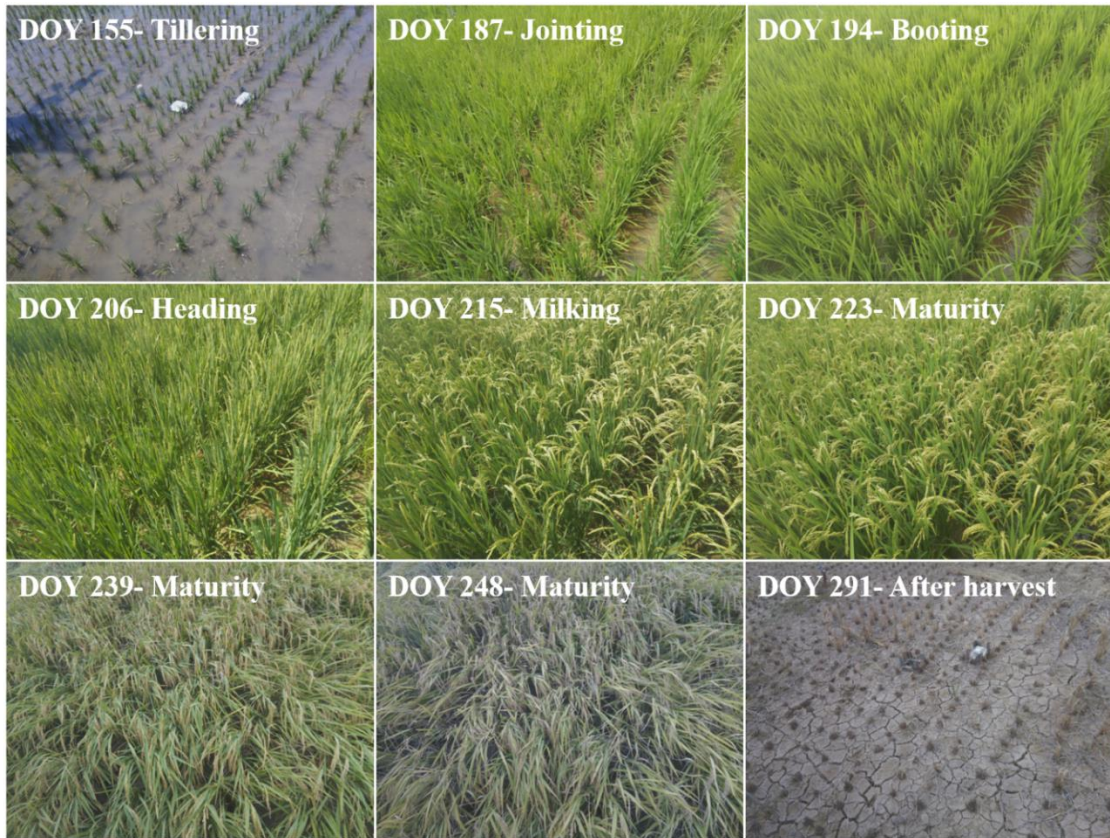
130 To conduct a systematic sampling that represented the study landscape, we relied on the elementary  
131 sampling unit (ESU), which has been used to evaluate moderate spatial resolution satellite images  
132 (Morissette *et al.* 2006). To minimize Global Positioning System (GPS) positioning error, we designed  
133 a 10 × 10 m pixel as an ESU to represent a pixel of an independent satellite image (i.e., Sentinel-2;  
134 Appendix 1). To evaluate the fusion products and their input satellite products (i.e., Landsat 8;  
135 Moderate Resolution Imaging Spectroradiometer [MODIS]), we designed the resolution of the ESUs  
136 to represent a pixel of each satellite imagery type: a 30 × 30 m Landsat 8 pixel by 4 ESUs and a 250 ×  
137 250 m MODIS pixel by 60 ESUs (Figure 2; Appendix 1).



138

139 **Figure 2 Schematic overview of the sampling design. The elementary sampling units (ESUs; 10**  
 140 **× 10 m) are represented by light green squares for vegetation land class and red circles for**  
 141 **mixed land class. Correspondingly colored squares around each ESU indicate each plot**  
 142 **(Background image source: Google).**

143 We collected *in situ* measurements using a hyperspectral spectroradiometer (Jaz, Ocean Optics,  
 144 Dunedin, FL, USA) equipped with fiber optics and a cosine corrector (CC-3-UV-T, Ocean Optics) to  
 145 measure bi-hemispheric reflectance. We located the cosine corrector at 3.5 m above ground using a  
 146 bar to cover 90% of the upwelling hemispherical irradiance in an ESU (Liu *et al.* 2017). To minimize  
 147 the uncertainties of *in situ* measurements from changing sky conditions, we collected 6 to 10 spectra  
 148 from the same position of each ESU, removed outliers that showed unstable values, and took the  
 149 average of the remaining spectra. Considering the satellite overpass times of around 10 a.m. to 12:30  
 150 p.m. (Sentinel-2, Landsat 8, and MODIS in this study), we conducted all ground measurements  
 151 between 9 a.m. and 1 p.m. local time (Coordinated Universal Time; UTC +9). To ensure the accurate  
 152 and precise location of the samples, we recorded the GPS coordinates of each point using a  
 153 commercial device that included a 0.5 m resolution base map (MONTANA 650TK, Garmin,  
 154 Switzerland). Moreover, we cross-checked the spectral quality of the *in situ* Jaz spectrometer data  
 155 with an ASD spectrometer (FieldSpec 4 Wide-ResField Spectroradiometer, ASD, Boulder, CO, USA)  
 156 and found good agreement (Appendix 2). The dates on which we performed the *in situ* measurements  
 157 are detailed in Figure 3.



158

159 **Figure 3 Images of the paddy in which the flux tower is located. Each image was labeled with**  
 160 **day-of-year (DOY) *in situ* measurements and the phenological stage of the rice.**

161 At the peak of the growing season, we classified the study landscape within a MODIS pixel into  
 162 two land cover types: vegetation and non-vegetation (e.g., road, building, ditch) with 76% and 24%  
 163 coverage, respectively. We did this by applying K-means classification to high-spatial-resolution  
 164 CubeSat images (i.e., PlanetScope Ortho Tile product, 3 m spatial resolution). Of the 15 plots, Plots 3,  
 165 4, and 6 were located on mixed land cover that included vegetation, roads, buildings, and ditches  
 166 (hereafter, “mixed” land class; Figure 2). Other plots were located on vegetation cover (hereafter,  
 167 “vegetation” land class; Figure 2). In the vegetation land class, Plot 5 was harvested 2 weeks earlier  
 168 than the other plots.

169

### 170 **2.3 Image fusion products**

171 We chose four state-of-the-art image fusion products that have been widely used or recently  
 172 developed (Table 1; Figure 4; Supplementary 1). We defined spatial interpolation in this study as  
 173 enhancing the spatial resolution of a coarse-resolution image (e.g., MODIS) on the predicted date  
 174 using each algorithm with paired input data (Luo *et al.* 2018; Zhu *et al.* 2010; Zhu *et al.* 2016).

175 ESTARFM (Zhu *et al.* 2010) is a weight function–based method that requires two or more pairs of

176 images. Each pair of images consists of fine-spatial-resolution images (e.g., Landsat) and coarse-  
177 spatial-resolution images (e.g., MODIS) acquired on the same date. Based on the pairs, ESTARFM  
178 predicts the fusion product using coarse-resolution images on the desired dates. It first searches for  
179 spectrally similar pixels from fine-resolution images and then weights them according to geographic  
180 distance from the central pixel in a moving window and the purity of their corresponding MODIS  
181 pixels. It then computes the coefficients of linear regression between the two pairs to convert coarse-  
182 resolution images to fine-resolution images. In a heterogeneous landscape, the reflectance changes in  
183 a coarse-resolution pixel across time do not equal the changes in the fine-resolution pixels within it.  
184 Thus, by introducing a conversion coefficient for each pixel, ESTARFM has advantages in  
185 heterogeneous areas (Zhu *et al.* 2010).

186 FSDAF (Zhu *et al.* 2016) is a hybrid method that requires minimum input data of one pair of fine-  
187 and coarse-spatial-resolution images (i.e., base images) on the same day and coarse-spatial-resolution  
188 images acquired on the prediction date. FSDAF uses unsupervised classification for the fine-  
189 resolution images and estimates the temporal change in each class at fine resolution from the coarse-  
190 resolution images. From the estimated temporal changes from the coarse images, FSDAF predicts the  
191 fine-resolution pixel by adding the temporal changes to the base fine-resolution pixel, i.e., temporal  
192 prediction. Then, FSDAF uses thin-plate spline interpolation to distribute residuals of temporal  
193 prediction for capturing abrupt land cover changes. An advantage of FSDAF is that it can capture both  
194 gradual and abrupt changes in land cover (Zhu *et al.* 2016).

195 STAIR (Luo *et al.* 2018) is a weight function-based method that uses all available Landsat and  
196 MODIS images. STAIR applies segmentation to homogeneous pixels to identify missing-value pixels,  
197 such as cloud pixels or data gaps caused by failures in the Landsat 7 Scan Line Corrector. The  
198 imputation of missing-value pixels in each segment is based on an adaptive-average correction, which  
199 assumes that changes in pixel values are approximately identical between neighborhood pixels within  
200 a short time frame (e.g., <2–3 weeks). Finally, STAIR fuses a daily MODIS time series with imputed  
201 Landsat-MODIS pairs to produce a daily time series of predicted surface reflectance images. A time-  
202 series based new cloud masking algorithm was applied for both MODIS and Landsat data to achieve  
203 higher performance in identifying cloud masked compared with the existing MODIS/Landsat cloud  
204 mask products. STAIR's strength is its ability to generate daily time series of fine-spatial-resolution  
205 products by systematically integrating Landsat-MODIS image pairs for missing-pixel imputation and  
206 automatically determining the input pair of fine and coarse images for the target date (Jiang *et al.*  
207 2020; Kimm *et al.* 2020).

208 CESTEM (Houborg and McCabe 2018a) is a machine-learning method that leverages rigorously  
209 calibrated 'gold standard' satellites (e.g., MODIS, Landsat) in concert with lower quality but superior  
210 resolution CubeSats (e.g., PlanetScope products; Planet Labs, San Francisco, CA, USA) to produce a

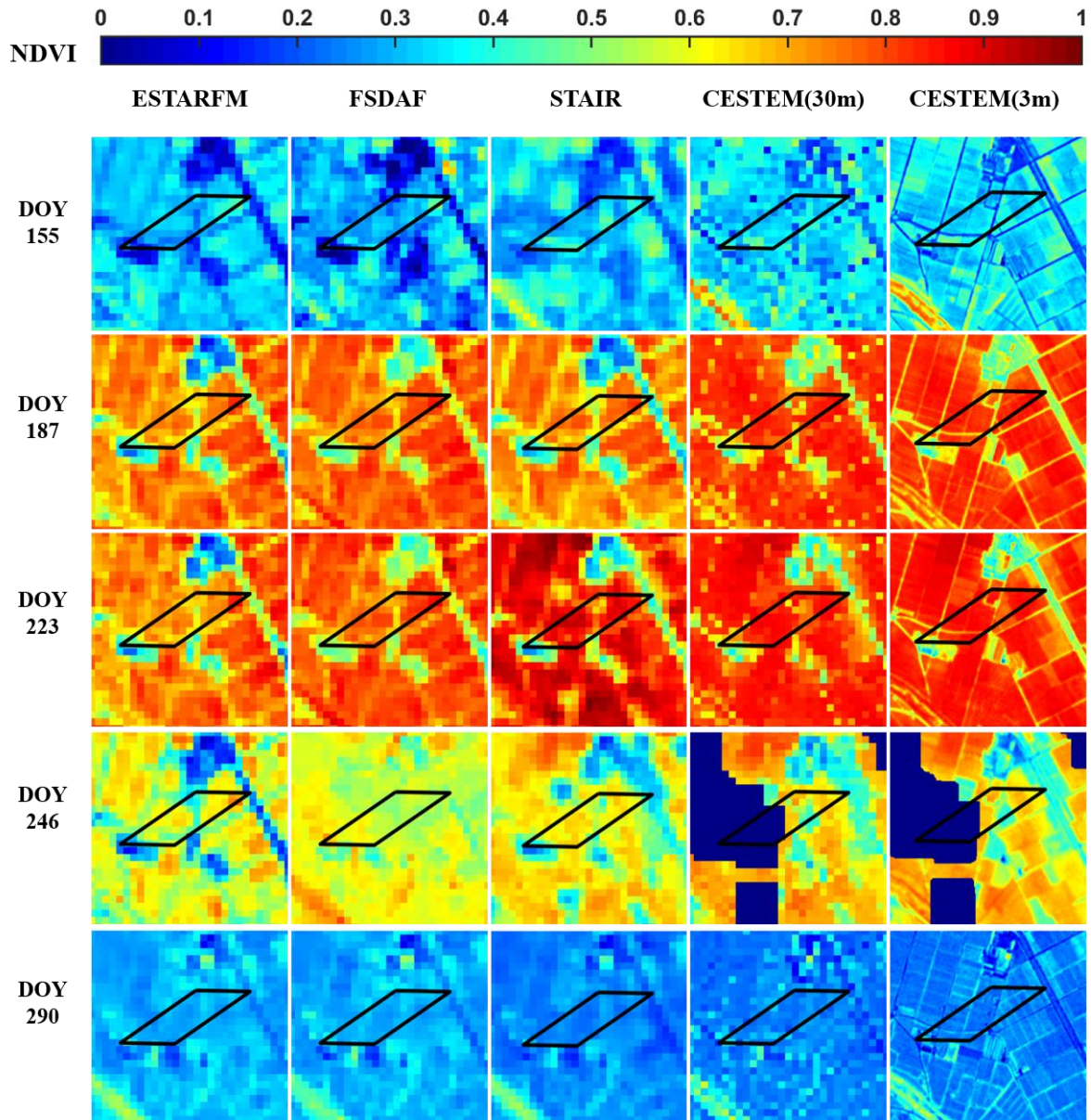


211 radiometrically harmonized and temporally consistent surface reflectance product. To ascertain  
 212 associations between training data of input explanatory variables (CubeSat spectral data) and the  
 213 target variable (Landsat 8 surface reflectance data), CESTEM uses a cubist rule-based regression  
 214 technique that is a nonparametric machine-learning approach belonging to the family of regression  
 215 tree methods. CESTEM harmonizes multi-sensor CubeSat images against Landsat and MODIS  
 216 images for consistent radiometric correction. Its key strength is its ability to ingest data from multiple  
 217 sensors with differing radiometric and spectral responses and generate images that are spectrally  
 218 consistent with Landsat 8 surface reflectance data while inheriting the high spatial resolution of  
 219 PlanetScope data (3 m). In this study, we used CESTEM version 1.0, which does not include a spatial  
 220 and temporal gap-filling process.

221 **Table 1 The image fusion products used in this study. ‘√’ and ‘N/A’ indicate whether the**  
 222 **specific process was included or not, respectively.**

Fusion product	Type	Target resolution	Spatial interpolation	Temporal gap filling	Radiometric correction	Input data (revisit frequency)
<b>ESTARFM</b>	Weight function-based method	30 m	√	N/A	N/A	MODIS (MOD09GQ; daily), Landsat (Landsat 8 OLI C1 Level 2; 16 days)
<b>FSDAF</b>	Hybrid method (weight function-based and unmixing method)	30 m	√	N/A	N/A	MODIS (MOD09GQ; daily), Landsat (Landsat 8 OLI C1 Level 2; 16 days)
<b>STAIR</b>	Weight function-based method	30 m	√	√	N/A	MODIS (MCD43A4; daily), Landsat (Landsat 7 and 8 Level 2; 16 days)
<b>CESTEM</b>	Machine-learning method	3 m	N/A	N/A	√	MODIS (MCD43A4; daily), Landsat (Landsat 8 OLI products Level 1T-6SV; 16 days), Planet (Planet Scope Ortho Tile; daily)

223 For comparisons across products, all fusion products were resampled to 30 m spatial resolution with  
 224 the map projection WGS 84, UTM zone 52N. To reduce the potential errors from resampling, we  
 225 tested three resampling methods (i.e. nearest-neighbor, bilinear, bicubic) with and without antialiasing  
 226 for CESTEM. Among the methods, we chose the nearest-neighbor interpolation method without  
 227 antialiasing as it showed the best performance to in-situ measurements (Supplementary 2).



228

229 **Figure 4** NDVI maps ( $870 \times 870$  m) of fusion products on *in situ* measurement dates, including  
 230 when all fusion products were available on the same date. The study site is located within the  
 231 black polygon that indicates a MODIS 250 m pixel. The dark navy color on CESTEM DOY 246  
 232 indicates masking due to clouds. All image fusion products were 30 m resolution except for  
 233 CESTEM (3 m), which was aggregated to 30 m resolution for consistency.

234

## 235 2.4 Input data for image fusion products

236 Satellite data with different spatial and temporal resolutions were used to generate image fusion  
237 products (Table 1). The ESTARFM and FSDAF products used MODIS surface reflectance daily  
238 products (C6 MOD09GQ) and Landsat 8 Operational Land Imager (OLI; Landsat 8 OLI/TIRS C1  
239 Level 2). The STAIR product used Landsat reflectance products from all platforms (Landsat 7  
240 Enhanced Thematic Mapper and Landsat 8 OLI) and MODIS (C6 MCD43A4 product). The CESTEM  
241 product used the Planet Scope Ortho Tile product (Planet Team 2018), Landsat 8 OLI products Level  
242 1T corrected to surface reflectance by 6SV, and nadir bidirectional reflectance distribution function  
243 (BRDF)-adjusted MODIS daily surface reflectance.

## 244 2.5 Evaluation

245 To evaluate fusion products with *in situ* measurements, we used NDVI (Rouse 1974; Tucker 1979),  
246 which is widely used as a measure of vegetation greenness as plants appear relatively dark in red but  
247 bright in the near-infrared (NIR). NDVI is a band ratio formulation (Eq. (1)) that is less sensitive to  
248 sensor-target-sun geometry than the surface reflectance of individual red and near-infrared channels  
249 (Feng *et al.* 2002; Ryu *et al.* 2010). Moreover, the spectral **discrepancies between MODIS and**  
250 **Landsat did not cause significant differences in NDVI products (Zhou *et al.* 2021).** Our *in situ*  
251 observation setting measured bi-hemispheric reflectance (i.e., blue-sky albedo), whereas satellite data  
252 provided the bidirectional reflectance factor (BRF). We confirmed that the discrepancy in view  
253 geometry between *in situ* and satellite images led to negligible differences in surface reflectance and  
254 NDVI (Appendix 3), which were within ranges reported by Czapla-Myers *et al.* (2015).

$$255 \quad NDVI = \frac{\rho_{NIR} - \rho_{Red}}{\rho_{NIR} + \rho_{Red}} \quad \text{Eq. (1)}$$

where  $\rho$  indicates reflectance. Hence,  $\rho_{NIR}$  is reflectance measurements in near-  
infrared region, and  $\rho_{Red}$  is reflectance measurements red region.

256 To ensure spectral consistency between *in situ* measurements and fusion NDVI products that  
257 emulated Landsat-like images, we converted the in-situ measurements to Landsat 8 OLI band-like  
258 data (red band: 636–673 nm, NIR band: 851–879 nm) using the spectral response function of OLI  
(Barsi *et al.* 2014) and then computed NDVI values.

259 The coefficient of determination ( $R^2$ ) of linear regression models, relative bias (*rbias*; Eq. (2)), and  
260 the relative root mean square error (*rRMSE*; Eq. (3)) were used to evaluate the image fusion  
261 products, as in previous studies (e.g. Chen *et al.* 2015a; Zhu *et al.* 2018). To evaluate the linear  
262 correlation between *in situ* measurement and fusion products,  $R^2$  was used, which equals squared  
263 Pearson correlation for linear regression models. From *rbias*, we can determine whether fusion

264 products overestimate or underestimate compared to *in situ* measurements. *rRMSE* is widely used in  
 265 quantitative assessments of image quality (Zhou and Bovik 2002).

266

$$rbias = \frac{E(\|A - B\|)}{E(B)} \quad \text{Eq. (2)}$$

$$rRMSE(A) = \frac{\sqrt{E((B - A)^2)}}{E(B)} \quad \text{Eq. (3)}$$

where A is fusion NDVI or original satellite NDVI products, B is *in situ* NDVI, and E is the mean operator.

267 To evaluate the overall performance of fusion NDVI products, we compared *in situ* NDVI (NDVI<sub>in situ</sub>) data to the fusion NDVI products (NDVI<sub>fusion</sub>; NDVI<sub>ESTARFM</sub>, NDVI<sub>FSDAF</sub>, NDVI<sub>STAIR</sub>,  
 268 NDVI<sub>CESTEM</sub>) and original satellite NDVI products (NDVI<sub>MODIS</sub>, NDVI<sub>Landsat</sub>). Regarding spatial  
 269 variation, NDVI<sub>fusion</sub> were evaluated on each date when NDVI<sub>in situ</sub> were measured and also evaluated  
 270 separately for the two land cover classes (i.e., vegetation and mixed land cover). To characterize  
 271 changes in NDVI within a MODIS pixel, we compared the NDVI<sub>MODIS</sub> value on each date to the  
 272 average of the NDVI<sub>in situ</sub> over each land cover class. Regarding temporal variation, NDVI<sub>fusion</sub> on each  
 273 plot were evaluated against corresponding NDVI<sub>in situ</sub> throughout the entire growing season.  
 274 Moreover, to evaluate the time-lag effects on NDVI<sub>fusion</sub>, especially in peak growing season, we used  
 275 the relative values that are the differences between mean NDVI<sub>in situ</sub> and mean NDVI<sub>fusion</sub> over the two  
 276 land cover classes. We evaluated the time-lag effect on NDVI<sub>fusion</sub> from DOY 206 (available date for  
 277 both MODIS and Landsat 8) to DOY 246 as NDVI<sub>fusion</sub> after DOY246 can be influenced by DOY 286  
 278 Landsat 8 data, which is near the harvest. For the purpose of comparison, we set all relative values to  
 279 start from zero on DOY 206. Due to data-gap by the cloud contamination or no data, NDVI<sub>MODIS</sub>,  
 280 NDVI<sub>Landsat</sub>, and NDVI<sub>fusion</sub> were not available on several *in situ* data dates. In this case, we used  
 281 satellite data within  $\pm 3$  days centered on *in situ* date (Table 2). The data dates used for the evaluation  
 282 are listed in Supplementary 3.

284

285 **Table 2 Days-of-year used to compare satellite data and fusion products against *in situ* NDVI.**  
 286 **Underline: date within  $\pm 3$  days of in-situ date due to clouds or no-data, \*: no data or cloud**  
 287 **contamination within  $\pm 3$  days, \*\*: only a fraction of data was available for the current date.**  
 288 )

Data	Day-of-year
------	-------------

In-situ	155	187	194	206	215	223	239	248	291
MODIS (MOD09GQ)	155	187	<u>195</u>	206	<u>213</u>	223	<u>243</u>	<u>246</u>	<u>290</u>
Landsat (Landsat8 OLI)	*	*	*	206	*	*	*	*	<u>286</u>
ESTARFM	155	187	<u>195</u>	*	<u>213</u>	223	<u>243</u>	<u>246</u>	<u>290</u>
FSDAF	155	187	<u>195</u>	206	<u>213</u>	223	<u>243</u>	<u>246</u>	<u>290</u>
STAIR	155	187	194	206	215	223	239	248	291
CESTEM	155	187	*	**	*	223	<u>238</u>	<u>247</u>	<u>290</u>

289

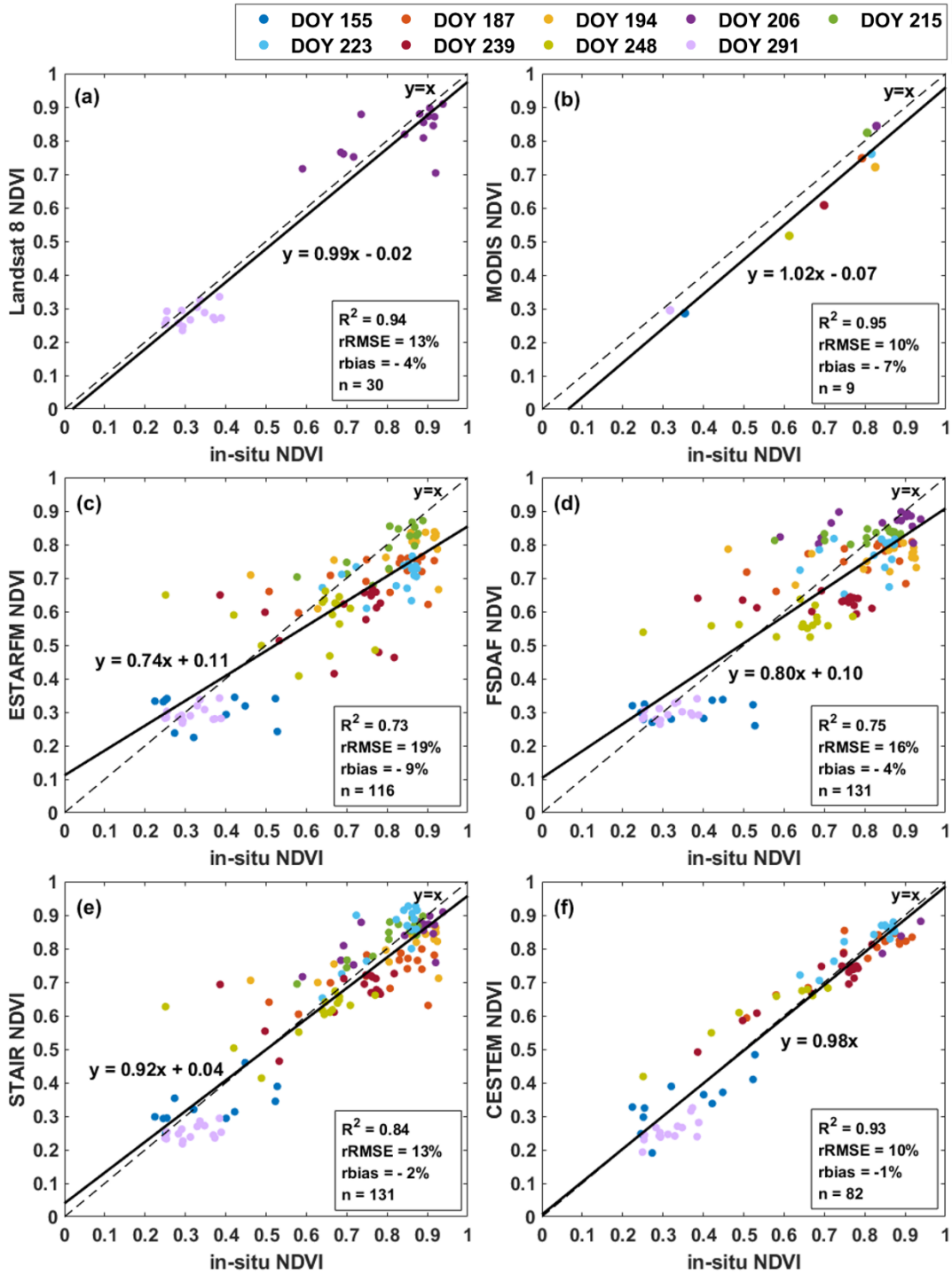
### 290 3. Results

291

#### 292 3.1 Spatio-temporal evaluation of satellite and image fusion NDVI products against *in situ* 293 NDVI

294 Regression analyses of satellite NDVI products against all available *in situ* measurements showed  
295 strong linear relationships and negative bias ( $R^2 > 0.93$ , *rbias* up to  $-7\%$ ).  $NDVI_{Landsat}$  against  $NDVI_{in}$   
296 *situ* was strongly correlated over the growing season ( $R^2 = 0.93$ , *rbias* =  $-4\%$ ).  $NDVI_{MODIS}$  also showed  
297 a strong correlation with the average  $NDVI_{fusion}$  values that covered the 250 m MODIS pixel ( $R^2 =$   
298  $0.95$ , *rbias* =  $-7\%$ ; Figure 5).

299  $NDVI_{fusion}$  products and all available  $NDVI_{in situ}$  showed strong linear relationships ( $R^2 = 0.73-0.93$ )  
300 and negative biases (*rbias* up to  $-9\%$ ). Linear relationships were strong with small biases in CESTEM,  
301 STAIR, FSDAF, and ESTARFM, with  $R^2$  (*rbias*) of  $0.93$  ( $-1\%$ ),  $0.84$  ( $-2\%$ ),  $0.75$  ( $-4\%$ ), and  $0.73$  ( $-$   
302  $9\%$ ), respectively (Figure 5).



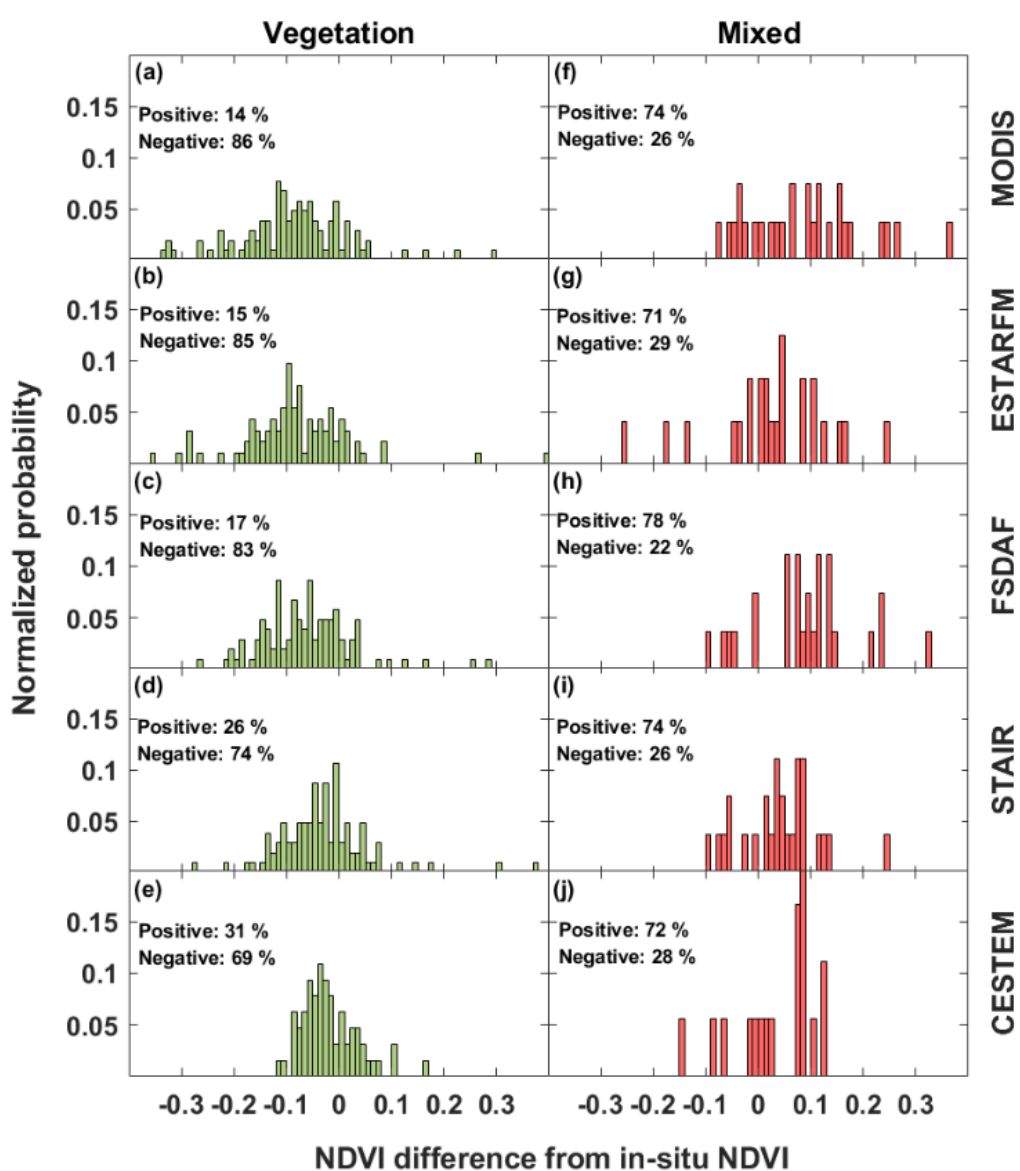
303

304 **Figure 5** Evaluation of Landsat and MODIS NDVI products (a-b) and image fusion NDVI  
 305 products (d-f) against *in situ* NDVI data sets across available dates. (a) Landsat 8, (b) MODIS,  
 306 (c) ESTARFM, (d) FSDAF, (e) STAIR, (f) CESTEM.  $R^2$ , *rRMSE*, *rbias*, and *n* are the coefficient  
 307 of determination, relative root mean square error, relative bias, and number of samples,  
 308 respectively. Colored circles indicate NDVI values for each day. Dashed and thick lines indicate  
 309 one-to-one lines ( $y = x$ ) and trend lines, respectively.

310

311 The difference between  $NDVI_{MODIS}$  values and  $NDVI_{in situ}$  values showed different trends for plots  
 312 of vegetation and mixed land classes. In the vegetation land class,  $NDVI_{MODIS}$  values were negatively  
 313 biased against 86% of  $NDVI_{in situ}$  (Figure 6). In the mixed land class, however,  $NDVI_{MODIS}$  were  
 314 positively biased against 76% of  $NDVI_{in situ}$  (Figure 6).

315 The bias distributions in  $NDVI_{fusion}$  against  $NDVI_{in situ}$  were similar to the biases in  $NDVI_{MODIS}$ .  
 316 Histograms of the difference between  $NDVI_{fusion}$  and  $NDVI_{in situ}$  mainly showed negative bias for the  
 317 vegetation land class (ESTARFM, 85%; FSDAF, 83%; STAIR, 74%; CESTEM, 69%; Figure 6) and  
 318 positive bias for the mixed land class (ESTARFM, 71%; FSDAF, 78%; STAIR, 74%; CESTEM,  
 319 72%; Figure 6).



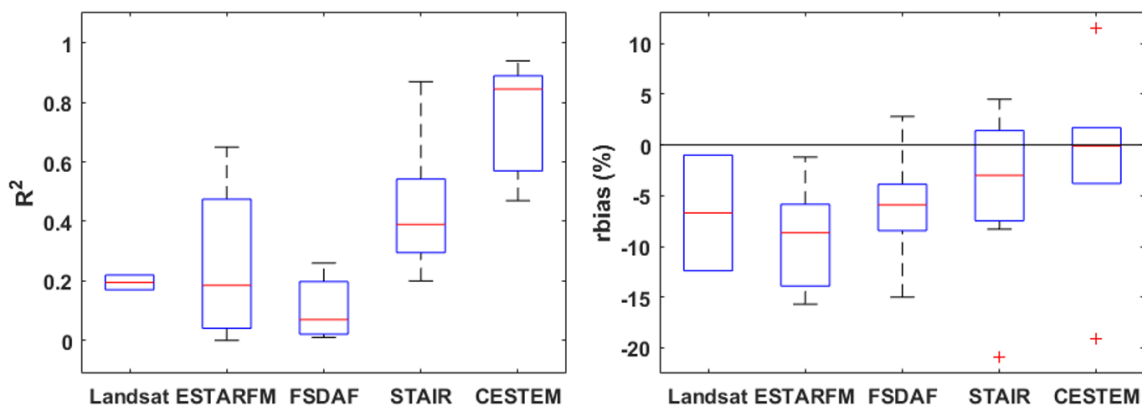
320

321 **Figure 6** Histogram of difference between MODIS or fusion NDVI and *in situ* NDVI for  
 322 vegetation (a–e) and mixed (f–j) land classes. Results shown here are based on data from all  
 323 dates.

324

### 325 3.2 Spatial evaluation of fusion NDVI products against *in situ* NDVI on each date

326 Regression analyses of  $NDVI_{Landsat}$  and  $NDVI_{fusion}$  against  $NDVI_{in situ}$  on each date showed wide  
 327 ranges in  $R^2$  and  $rbias$  (Figure 7; Appendix 4). Spatial relationships between  $NDVI_{Landsat}$  and  $NDVI_{in}$   
 328  $situ$  on individual dates (Figure 7;  $R^2 = 0.17$  and  $0.22$ ,  $rbias = -12.4\%$  and  $-1.0\%$ , Appendix 4) were  
 329 lower than the overall performance that included spatial and temporal variations (Figure 5a). Spatial  
 330 relationships between  $NDVI_{fusion}$  and  $NDVI_{in situ}$  for each date were also considerably scattered (Figure  
 331 7;  $R^2 = 0-0.94$ ,  $rbias -20.9\%$  to  $11.5\%$ , Appendix 4).



332

333 **Figure 7 Evaluation of spatial performance of fusion methods and Landsat 8 against *in situ***  
 334 **NDVI on each date. Boxplots (the 25<sup>th</sup> to 75<sup>th</sup> percentiles) of  $R^2$  and relative bias( $rbias$ ) are**  
 335 **shown. On each box, the red line indicates the median, and the whiskers extend to both**  
 336 **minimum and maximum values that do not consider outliers (+). A table with the corresponding**  
 337 **detailed numbers is shown in Appendix 4**

338

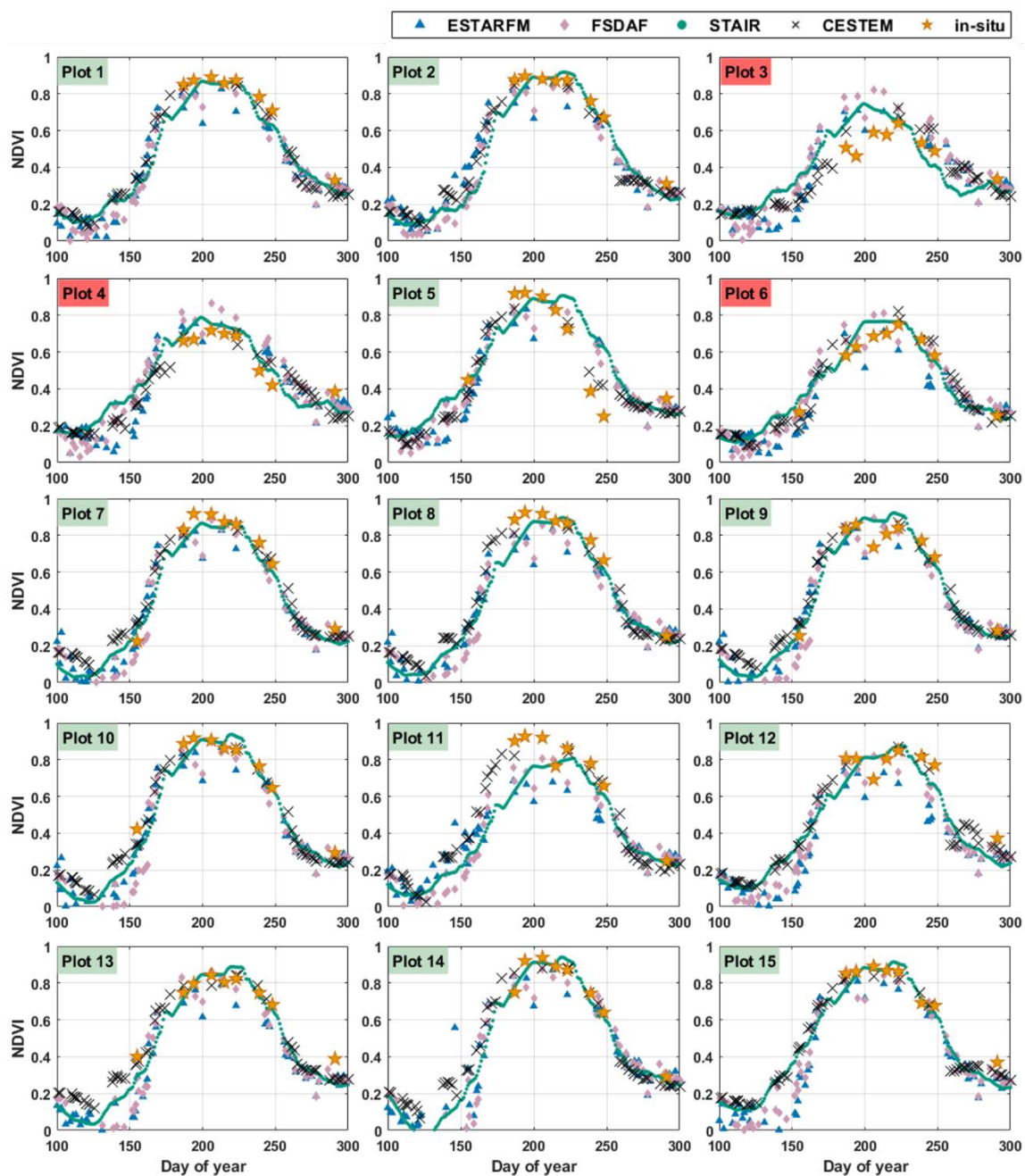
### 339 3.3 Temporal evaluation of fusion NDVI products against on each plot

340 Time series NDVI data sets across the 15 plots provided not only the seasonal change in NDVI but  
 341 also the amplitude of the change in NDVI. Early in the growing season,  $NDVI_{ESTARFM}$  and  $NDVI_{FSDAF}$   
 342 were lower than  $NDVI_{in situ}$ , whereas  $NDVI_{STAIR}$  and  $NDVI_{CESTEM}$  did not show such discrepancies  
 343 (Figure 8). All fusion products had similar values at the end of the growing season. At the peak of the  
 344 growing season,  $NDVI_{ESTARFM}$  and  $NDVI_{FSDAF}$  showed inconsistent NDVI values. In particular, the  
 345 NDVI of these two methods showed a clear drop near DOY 200 (Figure 8).

346 The  $NDVI_{fusion}$  on each plot showed linear relationships to  $NDVI_{in situ}$ , but biases varied from  
 347 negative to positive (Figure 8; Figure 9; Appendix 5). In the vegetation land class,  $NDVI_{fusion}$  generally  
 348 showed a negative bias against  $NDVI_{in situ}$  (Appendix 5), in particular at the peak of the growing  
 349 season (Figure 8). In the mixed land class,  $NDVI_{fusion}$  generally showed a positive bias against  $NDVI_{in}$   
 350  $situ$  on plots (Figure 8; Appendix 5), in particular at the peak of the growing season (Figure 8).

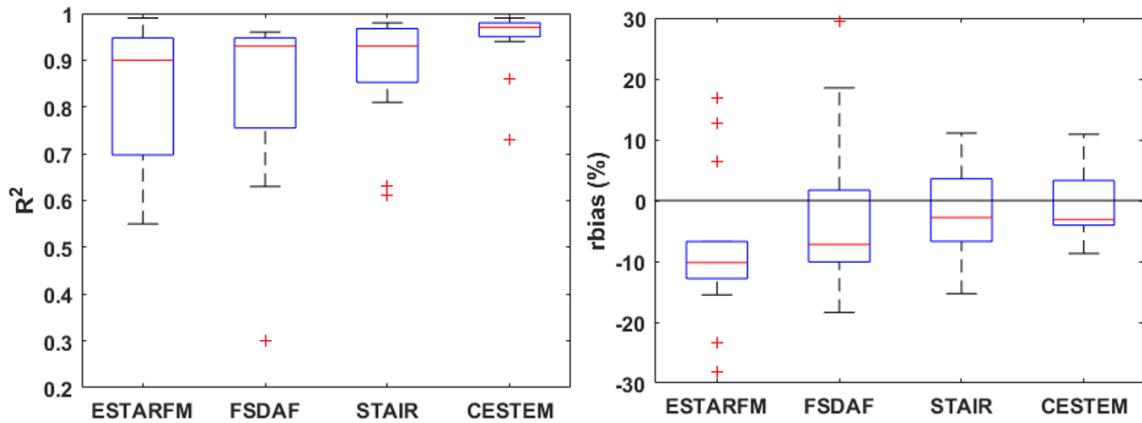


351 Among the vegetation land class plots, NDVI<sub>fusion</sub> on Plot 5 and Plot 11 showed different patterns  
 352 against NDVI<sub>in situ</sub>. NDVI<sub>ESTARFM</sub>, NDVI<sub>FSDAF</sub>, and NDVI<sub>STAIR</sub> did not catch a sharp decline in NDVI<sub>in</sub>  
 353 <sub>situ</sub> on Plot 5 (Figure 8), which led to a positive bias against NDVI<sub>in situ</sub> (Appendix 5). At the peak of  
 354 the growing season, the differences between NDVI<sub>in situ</sub> and NDVI<sub>ESTARFM</sub>, NDVI<sub>FSDAF</sub>, and NDVI<sub>STAIR</sub>  
 355 products were greater for Plot 11 than for the other vegetation plots (Figure 8), which resulted in a  
 356 negative bias against NDVI<sub>in situ</sub> up to -28.2% (Appendix 5).



357  
 358 **Figure 8** Seasonal variation in NDVI from image fusion products and *in situ* measurements.  
 359 Background colors of plot numbers indicate vegetation (green) and mixed (red) land classes.

360



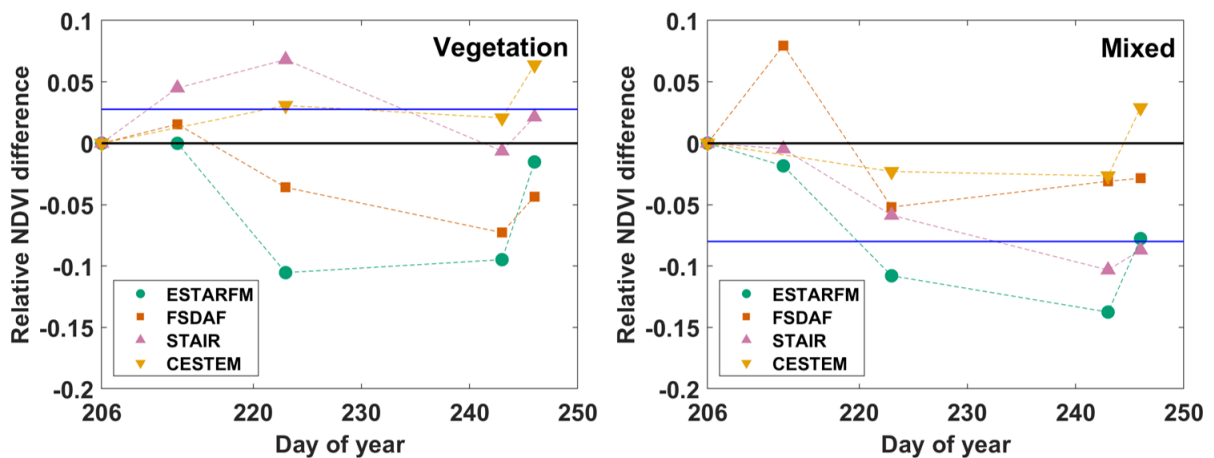
361

362 **Figure 9 Evaluation of temporal performance of fusion methods against *in situ* NDVI across all**  
 363 **15 plots. Boxplots (the 25<sup>th</sup> to 75<sup>th</sup> percentiles) of  $R^2$  and relative bias (*rbias*) are shown. On each**  
 364 **box, the red line indicates the median, and the whiskers extend to minimum and maximum**  
 365 **values that do not consider outliers (+). A table with the corresponding detailed numbers is**  
 366 **shown in Appendix 5.**

367

### 368 3.4 Evaluation of time-lag effects across land classes

369 Differences between  $NDVI_{fusion}$  and  $NDVI_{in\ situ}$  differed from differences between  $NDVI_{Landsat}$  and  
 370  $NDVI_{in\ situ}$  (blue line in Figure 10) over time. In the vegetation land class,  $NDVI_{ESTARFM}$  and  
 371  $NDVI_{FSDAF}$  showed clear time-lag effects as the relative NDVI difference became larger with time.  
 372  $NDVI_{STAIR}$  and  $NDVI_{CESTEM}$  differed from  $NDVI_{Landsat}$  but became closer to  $NDVI_{in\ situ}$  around DOY  
 373 240. In the mixed land class,  $NDVI_{ESTARFM}$  and  $NDVI_{STAIR}$  showed clear time-lag effects.  $NDVI_{FSDAF}$   
 374 and  $NDVI_{CESTEM}$  did not show clear time-lag effects.



375

376 **Figure 10 Evaluation of time-lag effects. Time-lag effects denote uncertainties caused by the**  
 377 **temporal interpolation process due to the time lag between the dates of the input pair and the**  
 378 **dates of predicted fusion products. Relative NDVI differences are the differences between mean**

379 *in situ* NDVI and mean fusion NDVI over the two land cover classes. Relative NDVI differences  
380 were set to start from zero on DOY 206, which was the date of input pair (Landsat and MODIS)  
381 for the fusion products. Blue horizontal lines indicate the difference between Landsat 8 NDVI  
382 and *in situ* NDVI on DOY 206, which is the target quality for fusion NDVI. MODIS was  
383 available on DOY206 and 256

384

## 385 4. Discussion

### 386 4.1 Strengths of image fusion products

387 All four fusion products showed tendencies in performances that are consistent with the design  
388 features of their algorithms. This is discussed in more detail in the following for all four methods.

389 Our results support the ability of ESTARFM and FSDAF to capture spatial and temporal variation,  
390 respectively, in land surface properties. As  $NDVI_{ESTARFM}$  and  $NDVI_{FSDAF}$  products shared identical  
391 input data, the difference in output can be entirely attributed to the difference in algorithms.

392 ESTARFM computes the coefficient of linear regression between two pairs (e.g., two Landsat-  
393 MODIS pairs) over each pixel because reflectance changes in coarse-spatial-resolution pixels across  
394 time do not equal those in fine-resolution pixels (Zhu *et al.* 2010). From the results, ESTARFM was  
395 consistent with design of the algorithm on heterogeneous landscape, as it was generally better than  
396 FSDAF at capturing spatial variation in NDVI on each date (Figure 7; Appendix 4). FSDAF was  
397 developed to capture both gradual and abrupt changes in land cover more effectively than ESTARFM  
398 by distributing residuals of temporal change using thin-plate spline interpolation (Zhu *et al.* 2016).  
399 This design of FSDAF was consistent with the results as  $NDVI_{FSDAF}$  outperformed  $NDVI_{ESTARFM}$  in  
400 explaining temporal variation for each plot (Figure 9; Appendix 5). FSDAF was also less sensitive  
401 than ESTARFM to time-lag effects (Figure 10).

402 STAIR confirmed the strengths of cloud-/gap-free products. The number of retrieved dates ( $n = 214$ )  
403 was highest for STAIR compared to the other products ( $n = 65, 66,$  and  $60$  for ESTARFM, FSDAF,  
404 CESTEM, respectively; Supplementary 3). The spatiotemporal variation in STAIR NDVI agreed  
405 better with  $NDVI_{in situ}$  values than that in  $NDVI_{ESTARFM}$  and  $NDVI_{FSDAF}$  (Figure 5; Figure 7; Figure 9).  
406 As time-lag effects were smaller in  $NDVI_{STAIR}$  than in  $NDVI_{ESTARFM}$  and  $NDVI_{FSDAF}$  products (Figure  
407 10). Moreover, the additional input pairs (e.g., Landsat 7) of  $NDVI_{STAIR}$  were better than  
408  $NDVI_{ESTARFM}$  at capturing spatial variation in  $NDVI_{in situ}$  at the peak of the growing season (Appendix  
409 4). In fact, STAIR even captured spatial NDVI patterns better than the original Landsat images  
410 (Figure 7; Appendix 4). The strong performance of STAIR is likely due to the use a temporal gap-  
411 filling method involving all available Landsat images (e.g., Landsat 7 and Landsat 8) to produce daily  
412 cloud-free products without missing pixels.

413 CESTEM fully utilized the merits of CubeSat constellation data (daily, 3 m resolution) in its

414 harmonization process with Landsat and MODIS. The CESTEM product has already been proven a  
415 promising data set for land surface monitoring, but detailed evaluation of the product quality with  
416 pixel-scale *in situ* measurements has not been attempted before (Aragon *et al.* 2018; Houborg and  
417 McCabe 2018b). Regarding both spatial and temporal resolution, NDVI<sub>CESTEM</sub> showed the strongest  
418 linear relationship with the smallest bias to NDVI<sub>*in situ*</sub> values of all of the products (Figure 5; Figure 7;  
419 Figure 8). Also, spatial and temporal variation in NDVI<sub>CESTEM</sub> were in better agreement with NDVI<sub>*in*</sub>  
420 <sub>*situ*</sub> than that of NDVI<sub>Landsat</sub> against NDVI<sub>*in situ*</sub> despite aggregating the CESTEM product to the coarser  
421 spatial resolution of Landsat (Figure 5, Figure 7; Figure 9). We partly attribute this strong  
422 performance of CESTEM to lower uncertainties in input data and noise reduction via data aggregation  
423 (see section 4.2 below).

424

425

#### 426 **4.2 Impacts of input data and fusion algorithms on uncertainties in fusion NDVI**

427 NDVI<sub>Landsat</sub> is the main driver that controls spatial variation in NDVI<sub>fusion</sub> products, so uncertainties  
428 in NDVI<sub>Landsat</sub> must be quantified systematically. Several studies have evaluated NDVI<sub>Landsat</sub> time  
429 series against ground-based spectral sensors in fixed positions (Ke *et al.* 2015; Kim *et al.* 2019; Ryu *et*  
430 *al.* 2014); however, a comprehensive evaluation of NDVI<sub>Landsat</sub> against *in situ* spectral sensors in terms  
431 of spatial variation is lacking. We found that NDVI<sub>Landsat</sub> explained only 17–22% of spatial variation  
432 over the landscape for each date (DOY 206, the peak of the growing season; DOY 286 after harvest;  
433 Figure 7; Appendix 4). Because we aligned the spectral response, acquisition time, and footprint  
434 between Landsat and *in situ* data, we assume that uncertainties in geolocation in Landsat 8 might have  
435 contributed to the scattered relationships. Earlier studies have reported that the absolute geolocation  
436 accuracy of a Landsat 8 Tier 2 product (i.e., surface reflectance) is greater than 12 m (Dwyer *et al.*  
437 2018; Storey *et al.* 2014). Of the four image fusion products, ESTARFM and FSDAF are particularly  
438 dependent on Landsat 8 for generating sub-MODIS pixel spatial variability. Indeed, scattered  
439 relationships were also shown in the NDVI<sub>fusion</sub> on each date (Figure 5; Figure 7; Appendix 4).  
440 Noteworthy is the fact that NDVI<sub>CESTEM</sub> had a higher R<sup>2</sup> than NDVI<sub>Landsat</sub> on DOY291 (Figure 7;  
441 Appendix 4). During CESTEM pre-processing, Landsat data is geometrically aligned to the CubeSat  
442 data to ensure a near-perfect alignment (Houborg and McCabe 2018a). Unlike the other products,  
443 CESTEM relies on CubeSats, which have a relatively high positional accuracy (<10 m) and therefore  
444 likely to be better correlated with the observed spatial NDVI variability (Planet Team 2018). In  
445 addition, we aggregated 3 m CESTEM pixels to match the Landsat pixel resolution (30 m), which will  
446 reduce random noise. This difference between Landsat and CESTEM might explain the performance  
447 of NDVI<sub>CESTEM</sub> in capturing spatial variation in NDVI<sub>*in situ*</sub> on each date (Figure 7; Appendix 4) which  
448 was superior to that of the other fusion NDVI products.

449 NDVI<sub>MODIS</sub> was negatively biased toward NDVI<sub>in situ</sub> (Figure 5). NDVI<sub>MODIS</sub> played a role in  
450 temporal patterns of NDVI<sub>fusion</sub>. Ideally, NDVI<sub>in situ</sub> that covers 3 × 3 MODIS pixels would be  
451 measured to account for uncertainties in the MODIS geolocation product (Wolfe *et al.* 2002).  
452 However, this is extremely labor intensive and was beyond the scope of our experiment. We  
453 assumed that the negative bias in NDVI<sub>MODIS</sub> would result in the underestimation of NDVI<sub>fusion</sub> in  
454 the vegetation land class at the peak of the growing season (Figure 8). As the vegetation land class  
455 accounted for 76% of the MODIS 250 m pixel (section 2.2), the NDVI<sub>MODIS</sub> values mainly  
456 represented the dynamics of the vegetation land class. However, the NDVI<sub>MODIS</sub> values were higher  
457 than NDVI<sub>in situ</sub> in the mixed land class within a MODIS subpixel (Figure 6). The positive bias of  
458 NDVI<sub>MODIS</sub> against NDVI<sub>in situ</sub> on mixed land class might explain the overestimation of NDVI<sub>fusion</sub>  
459 values than NDVI<sub>in situ</sub> values in the mixed land class, in particular at the peak of the growing  
460 season (Figure 8). These findings suggest that biases in NDVI<sub>MODIS</sub> can lead to different signs in  
461 biases for dominant and minor land cover classes within the MODIS pixel.

462 The coarse spatial resolution of MODIS caused biases in the spatial interpolation process. In the  
463 vegetation land class, Plot 5 was harvested 2 weeks earlier than the other plots (section 2.2), so it had  
464 a lower NDVI<sub>in situ</sub> value than the other vegetation class plots around DOY 230 (Figure 8). The area  
465 harvested on DOY 230 was only 3.7% of the total study area over MODIS 250 m pixels. Therefore,  
466 the harvest event on Plot 5 was not reflected in the NDVI<sub>MODIS</sub> value, which explains the positively  
467 biased NDVI<sub>fusion</sub> values in this plot (Figure 8). NDVI<sub>CESTEM</sub> successfully detected the early harvest in  
468 Plot 5 (Figure 8) and showed a stronger linear relationship against NDVI<sub>in situ</sub> for Plot 5 compared to  
469 the other NDVI<sub>fusion</sub> (Appendix 5). This performance of NDVI<sub>CESTEM</sub> is due to the use of the CubeSat  
470 constellation that has near-daily temporal coverage at fine spatial resolution. (Houborg and McCabe  
471 2018a).

472 Weight function–based fusion NDVI, which includes ESTARFM, FSDAF, and STAIR, revealed  
473 uncertainties across land cover transition zones. Such methods give greater weight to neighboring  
474 pixels for spatial interpolation (Zhu *et al.* 2018). Plot 11 was close to a road and ditch to the east and  
475 south (Figure 2); thus, five out of eight neighboring pixels were classified as belonging to the mixed  
476 class, which had lower NDVI than the vegetation class during the peak of the growing season. Indeed,  
477 compared to the other vegetation plots, NDVI<sub>fusion</sub> values for Plot 11 were underestimated during the  
478 peak growing season (Figure 8). Presumably, the mixed land class surrounding Plot 11 caused the  
479 negative bias through the weight function. Thus, the correlation between NDVI<sub>fusion</sub> values and  
480 NDVI<sub>in situ</sub> values on Plot 11 was relatively weak (Appendix 5). By contrast, NDVI<sub>CESTEM</sub> values use  
481 each CubeSat pixel without spatial interpolation. Therefore, NDVI<sub>CESTEM</sub> values on Plot 11 captured  
482 the dynamics of NDVI<sub>in situ</sub> (Figure 8).

483 Time-lag effects did not always degrade the performance of NDVI<sub>fusion</sub> when input data were biased

484 against *in situ* data. Previous studies have reported that time-lag effects degrade the performance of  
485 fusion products against input data (Fu *et al.* 2015; Gao *et al.* 2006; Olexa and Lawrence 2014; Walker  
486 *et al.* 2012; Xie *et al.* 2018; Zhu *et al.* 2010). As in earlier studies, the relative difference in NDVI  
487 between fusion products and *in situ* data tended to increase over time from the paired dates, in  
488 particular for NDVI<sub>ESTARFM</sub> and NDVI<sub>FSDAF</sub> in the vegetation land class and NDVI<sub>ESTARFM</sub> and  
489 NDVI<sub>STAIR</sub> in the mixed land class (Figure 10). However, time-lag effects in NDVI<sub>STAIR</sub> and  
490 NDVI<sub>CESTEM</sub> in the vegetation land class were not apparent (Figure 10). In the mixed land class,  
491 NDVI<sub>FSDAF</sub> and NDVI<sub>CESTEM</sub>, which did not show clear time-lag effects, did not perform better than  
492 NDVI<sub>ESTARFM</sub> and NDVI<sub>STAIR</sub>, which showed clear time-lag effects, in predicting *in situ* NDVI (Figure  
493 10). Thus, when the input pair data include biases against *in situ* data, time-lag effects do not  
494 necessarily correspond to a decrease in performance against *in situ* data.

495

### 496 **4.3 Future works and perspectives**

497 Rapid and substantial advancements in image fusion methods have emerged in recent years.  
498 Even during the preparation of this report, we found that each image fusion method used in this study  
499 had been improved. For example, highly scalable STARFM was recently implemented on Google  
500 Earth Engine (Gorelick *et al.* 2017) to produce daily gap-free reflectance products (Moreno-Martínez  
501 *et al.* 2020). FSDAF was improved by the incorporation of constrained least squares theory and the  
502 subpixel class fraction change, called IFSDAF (Liu *et al.* 2019) and SFSDAF (Li *et al.* 2020),  
503 respectively. In the present study, STAIR 1.0 integrated all available Landsat and MODIS data sets.  
504 Recently, STAIR also assimilated Sentinel-2 data (STAIR 2.0) (Luo *et al.* 2020), CubeSats, and the  
505 latter work allows the production of 3 m daily leaf area index (LAI) maps across the U.S. Corn Belt  
506 (Kimm *et al.* 2020). STAIR also develops the real-time production capability that enables the same-  
507 day 10m and 30m daily surface reflectance products for anywhere in the earth, through Aspiring  
508 Universe Corporation for the commercialization applications. CESTEM version 1.0, used in the  
509 present study, did not include a gap-filling algorithm, which led to substantial data gaps in parts of the  
510 images (e.g., DOY246 in Figure 4) or even entire scenes during the growing season (e.g., DOY 225 to  
511 DOY 237). Recently, a significantly updated implementation of CESTEM has become an integral  
512 component towards Planet's vision of producing a next generation, analysis ready, and harmonized  
513 product, which delivers clean (i.e., free from clouds and shadows), gap-filled (i.e., daily, 3 m),  
514 temporally consistent, and radiometrically robust surface reflectance integrating the best features from  
515 both public (e.g., Landsat, Sentinel, MODIS) and private missions (Houborg and Zuleta 2019).  
516 Evaluating these recently updated image fusion methods is beyond the scope of this work, but we  
517 anticipate that our systematic, *in situ* spectral data will be helpful in evaluating those methods.

518 Current study has limitations and suggests further improvements. We focused on a flat, rice paddy

519 landscape to evaluate image fusion products. The data collection and evaluation scheme in this study  
520 might be applied to other ecosystems in short vegetation. For tall woody ecosystems, our manual data  
521 collection scheme will not work. In this case, drone could be used to generate reference maps in space  
522 and time (Candiago *et al.* 2015; Hashimoto *et al.* 2019; Zhang *et al.* 2020). Our results suggest each  
523 image fusion algorithm has pros and cons. Furthermore, each algorithm has been under rapid  
524 improvements as discussed above. Though one convergent lesson is clear. In heterogeneous  
525 landscapes not limited to rice paddy, image fusion algorithm based on spatial or temporal  
526 interpolation methods will involve inherent random errors and biases. Therefore, preparing more fine  
527 resolution images that allow more input pair dates and minimize interval of prediction dates is the key  
528 to reduce uncertainties in image fusion methods.

529 Our results highlight the importance of fine-spatial-resolution images with high revisit frequency.  
530 In the past, MODIS played a major role in capturing temporal variation, as satellite images that  
531 capture fine spatial scales, such as those captured by Landsat, have low revisit frequencies. With the  
532 emergence of new satellites and CubeSat constellations, it is possible to produce images that have  
533 both fine spatial and temporal resolution, such as Harmonized Landsat/Sentinel-2 products (Claverie  
534 *et al.* 2018) and PlanetScope products (Planet Team 2018). These harmonized data sets require  
535 careful, precise cross-calibration across satellites. Recent advancements in inexpensive spectral  
536 sensors (Kim *et al.* 2019; Ryu *et al.* 2010) offer new opportunities to continuously monitor land  
537 surface reflectance in situ, which could be deployed as a network across landscape. Incorporating  
538 geostationary satellites into the image fusion framework will help manage sustained data gaps due to  
539 clouds and decrease uncertainties in fusion products caused by temporal interpolation. We envision  
540 that the integration of ground-based spectral sensing networks and harmonized satellite images will  
541 lead to fundamental advancements in high-quality image fusion products.

542

## 543 **5. Summary and Conclusions**

544 In this study, we comprehensively evaluated four state-of-the-art image fusion products against *in*  
545 *situ* spectral measurements over a heterogeneous rice paddy landscape during the 2017 growing  
546 season. All four fusion products showed strong linear correlations to in-situ data when focusing on  
547 temporal observations or pooling spatial and temporal observations. However, we found large  
548 differences in performance when evaluating performance for spatial variations. Overall, CESTEM  
549 outperformed the other three products that relied on the spatial interpolation and STAIR showed  
550 better performance than ESTARFM and FSDAF. The detailed evaluation of fusion products against *in*  
551 *situ* measurements revealed the following insights on the causes underlying the limited performance  
552 of fusion methods. First, biases in forcing data (i.e., Landsat 8 and MODIS) propagated rather directly  
553 to image fusion products. Second, Landsat 8 NDVI only explained 17–22% of spatial variability

554 against *in situ* data, which explains the poor performance of the Landsat-based fusion products to  
555 predict NDVI spatial patterns. Third, the coarse spatial resolution of MODIS led to NDVI biases in  
556 vegetation and mixed land cover classes. Fourth, image fusion products based on spatial interpolation  
557 showed large biases in vegetation plots surrounded by mixed land cover plots. Finally, time-lag  
558 effects did not always degrade the performance of fusion NDVI when input pair data were biased. Our  
559 results identify key sources of uncertainty, which will be important for improving image fusion  
560 products. Also, constellation-based products such as CESTEM are expected to be more widely used  
561 due to their robust performance and high spatio-temporal resolution without the need for spatial  
562 interpolation.

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568 *in situ* data openly accessible via [http://environment.snu.ac.kr/landscape\\_spectral/](http://environment.snu.ac.kr/landscape_spectral/) and expect that  
569 they will be useful for evaluating future image fusion products.

570

### 571 **6. References**

- 572 Aragon, B., Houborg, R., Tu, K., Fisher, J.B., & McCabe, M. (2018). CubeSats enable high  
573 spatiotemporal retrievals of crop-water use for precision agriculture. *Remote Sensing*, 10, 1867
- 574 Barsi, J., Lee, K., Kvaran, G., Markham, B., & Pedelty, J. (2014). The spectral response of the  
575 Landsat-8 Operational Land Imager. *Remote Sensing*, 6, 10232
- 576 Candiago, S., Remondino, F., De Giglio, M., Dubbini, M., & Gattelli, M. (2015). Evaluating  
577 Multispectral Images and Vegetation Indices for Precision Farming Applications from UAV Images.  
578 *Remote Sensing*, 7, 4026-4047
- 579 Chen, B., Huang, B., & Xu, B. (2015a). Comparison of spatiotemporal fusion models: A review.  
580 *Remote Sensing*, 7, 1798–1835
- 581 Chen, B., Huang, B., & Xu, B. (2015b). Fine land cover classification using daily synthetic Landsat-  
582 like images at 15-m resolution. *IEEE Geoscience and Remote Sensing Letters*, 12, 2359–2363
- 583 Claverie, M., Ju, J., Masek, J.G., Dungan, J.L., Vermote, E.F., Roger, J.-C., Skakun, S.V., & Justice,  
584 C. (2018). The Harmonized Landsat and Sentinel-2 surface reflectance data set. *Remote Sensing of*  
585 *Environment*, 219, 145–161
- 586 Czaplá-Myers, J., McCorkel, J., Anderson, N., Thome, K., Biggar, S., Helder, D., Aaron, D., Leigh,  
587 L., & Mishra, N. (2015). The ground-based absolute radiometric calibration of Landsat 8 OLI.  
588 *Remote Sensing*, 7, 600–626
- 589 Dechant, B., Ryu, Y., & Kang, M. (2019). Making full use of hyperspectral data for gross primary  
590 productivity estimation with multivariate regression: Mechanistic insights from observations and  
591 process-based simulations. *Remote Sensing of Environment*, 234, 111435



592 Dechant, B., Ryu, Y., Badgley, G., Zeng, Y., Berry, J.A., Zhang, Y., Goulas, Y., Li, Z., Zhang, Q.,  
593 Kang, M., Li, J., & Moya, I. (2020). Canopy structure explains the relationship between  
594 photosynthesis and sun-induced chlorophyll fluorescence in crops. *Remote Sensing of Environment*,  
595 241, 111733

596 Ding, Y., Zhao, K., Zheng, X., & Jiang, T. (2014). Temporal dynamics of spatial heterogeneity over  
597 cropland quantified by time-series NDVI, near infrared and red reflectance of Landsat 8 OLI imagery.  
598 *International Journal of Applied Earth Observation and Geoinformation*, 30, 139–145

599 Dwyer, J.L., Roy, D.P., Sauer, B., Jenkerson, C.B., Zhang, H.K., & Lymburner, L. (2018). Analysis  
600 ready data: Enabling analysis of the Landsat archive. *Remote Sensing*, 10, 1363

601 Emelyanova, I.V., McVicar, T.R., Van Niel, T.G., Li, L.T., & van Dijk, A.I.J.M. (2013). Assessing the  
602 accuracy of blending Landsat–MODIS surface reflectances in two landscapes with contrasting spatial  
603 and temporal dynamics: A framework for algorithm selection. *Remote Sensing of Environment*, 133,  
604 193–209

605 Feng, G., Yufang, J., Schaaf, C.B., & Strahler, A.H. (2002). Bidirectional NDVI and atmospherically  
606 resistant BRDF inversion for vegetation canopy. *IEEE Transactions on Geoscience and Remote  
607 Sensing*, 40, 1269–1278

608 Fu, D.J., Zhang, L.F., Chen, H., Wang, J., Sun, X.J., & Wu, T.X. (2015). Assessing the effect of  
609 temporal interval length on the blending of Landsat-MODIS surface reflectance for different land  
610 cover types in southwestern continental United States. *ISPRS International Journal of Geo-  
611 Information*, 4, 2542–2560

612 Gao, F., Anderson, M.C., Zhang, X., Yang, Z., Alfieri, J.G., Kustas, W.P., Mueller, R., Johnson, D.M.,  
613 & Prueger, J.H. (2017a). Toward mapping crop progress at field scales through fusion of Landsat and  
614 MODIS imagery. *Remote Sensing of Environment*, 188, 9–25

615 Gao, F., Anderson, M.C., Zhang, X.Y., Yang, Z.W., Alfieri, J.G., Kustas, W.P., Mueller, R., Johnson,  
616 D.M., & Prueger, J.H. (2017b). Toward mapping crop progress at field scales through fusion of  
617 Landsat and MODIS imagery. *Remote Sensing of Environment*, 188, 9–25

618 Gao, F., He, T., Masek, J.G., Shuai, Y., Schaaf, C.B., & Wang, Z. (2014). Angular effects and  
619 correction for medium resolution sensors to support crop monitoring. *IEEE Journal of Selected Topics  
620 in Applied Earth Observations and Remote Sensing*, 7, 4480–4489

621 Gao, F., Masek, J., Schwaller, M., & Hall, F. (2006). On the blending of the Landsat and MODIS  
622 surface reflectance: Predicting daily Landsat surface reflectance. *IEEE Transactions on Geoscience  
623 and Remote Sensing*, 44, 2207–2218

624 Gorelick, N., Hancher, M., Dixon, M., Ilyushchenko, S., Thau, D., & Moore, R. (2017). Google Earth  
625 Engine: Planetary-scale geospatial analysis for everyone. *Remote Sensing of Environment*, 202, 18-27

626 Hashimoto, N., Saito, Y., Maki, M., & Homma, K. (2019). Simulation of Reflectance and Vegetation  
627 Indices for Unmanned Aerial Vehicle (UAV) Monitoring of Paddy Fields. *Remote Sensing*, 11, 2119

628 Hilker, T., Wulder, M.A., Coops, N.C., Linke, J., McDermid, G., Masek, J.G., Gao, F., & White, J.C.  
629 (2009a). A new data fusion model for high spatial- and temporal-resolution mapping of forest  
630 disturbance based on Landsat and MODIS. *Remote Sensing of Environment*, 113, 1613–1627

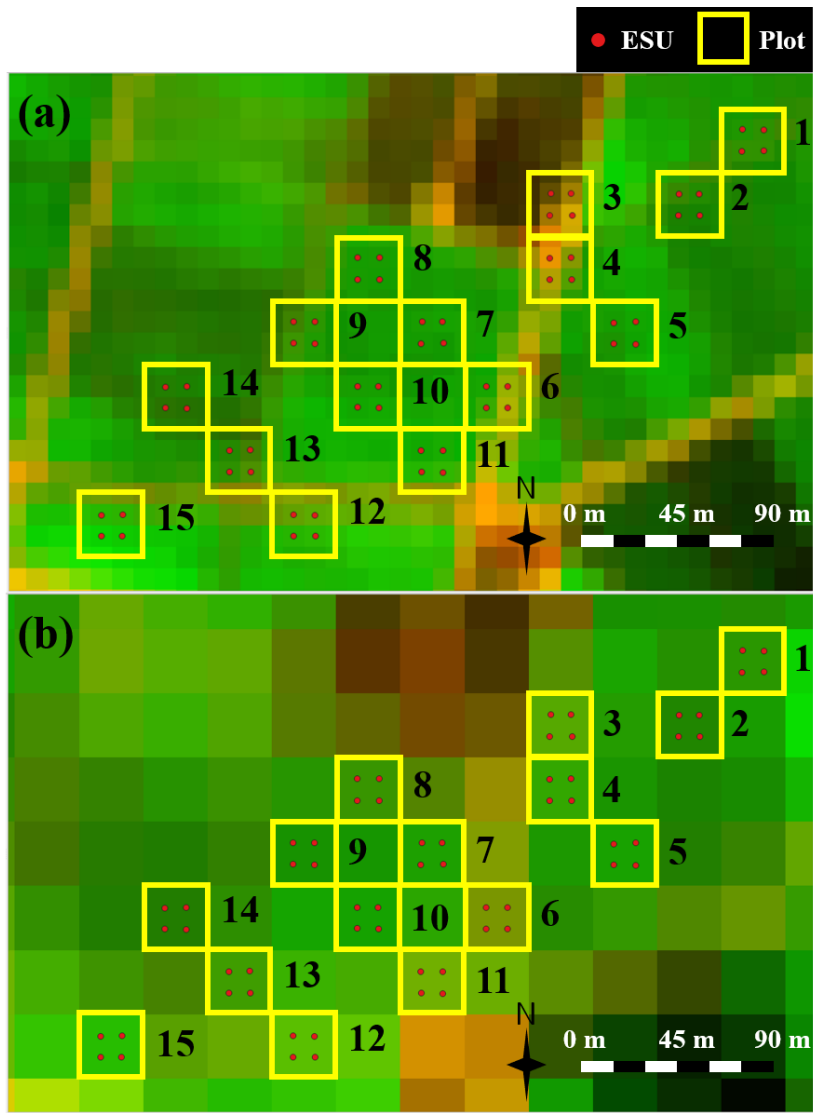
631 Hilker, T., Wulder, M.A., Coops, N.C., Seitz, N., White, J.C., Gao, F., Masek, J.G., & Stenhouse, G.  
632 (2009b). Generation of dense time series synthetic Landsat data through data blending with MODIS  
633 using a spatial and temporal adaptive reflectance fusion model. *Remote Sensing of Environment*, 113,  
634 1988–1999

635 Houborg, R., & McCabe, M.F. (2018a). A Cubesat enabled Spatio-Temporal Enhancement Method  
636 (CESTEM) utilizing Planet, Landsat and MODIS data. *Remote Sensing of Environment*, 209, 211–  
637 226

- 638 Houborg, R., & McCabe, M.F. (2018b). Daily retrieval of NDVI and LAI at 3 m resolution via the  
639 fusion of CubeSat, Landsat, and MODIS data. *Remote Sensing*, 10, 890
- 640 Houborg, R., & Zuleta, I. (2019). Very high resolution (daily, 3 m) data cubes of surface reflectance  
641 synergizing observations from CubeSats, Sentinel-2, Landsat 8, and MODIS. In, AGU Fall Meeting  
642 Abstracts (pp. B54D-01)
- 643 Huang, Y., Ryu, Y., Jiang, C., Kimm, H., Kim, S., Kang, M., & Shim, K. (2018). BESS-Rice: A  
644 remote sensing derived and biophysical process-based rice productivity simulation model.  
645 *Agricultural and Forest Meteorology*, 256–257, 253–269
- 646 Hwang, Y., Ryu, Y., Huang, Y., Kim, J., Iwata, H., & Kang, M. (2020). Comprehensive assessments of  
647 carbon dynamics in an intermittently-irrigated rice paddy. *Agricultural and Forest Meteorology*, 285–  
648 286, 107933
- 649 Jiang, C., Guan, K., Pan, M., Ryu, Y., Peng, B., & Wang, S. (2020). BESS-STAIR: A framework to  
650 estimate daily, 30 m, and all-weather crop evapotranspiration using multi-source satellite data for the  
651 US Corn Belt. *Hydrol. Earth Syst. Sci.*, 24, 1251–1273
- 652 Ke, Y., Im, J., Lee, J., Gong, H., & Ryu, Y. (2015). Characteristics of Landsat 8 OLI-derived NDVI by  
653 comparison with multiple satellite sensors and in-situ observations. *Remote Sensing of Environment*,  
654 164, 298–313
- 655 Kim, J., Ryu, Y., Jiang, C., & Hwang, Y. (2019). Continuous observation of vegetation canopy  
656 dynamics using an integrated low-cost, near-surface remote sensing system. *Agricultural and Forest  
657 Meteorology*, 264, 164–177
- 658 Kimm, H., Guan, K., Jiang, C., Peng, B., Gentry, L.F., Wilkin, S.C., Wang, S., Cai, Y., Bernacchi, C.J.,  
659 Peng, J., & Luo, Y. (2020). Deriving high-spatiotemporal-resolution leaf area index for  
660 agroecosystems in the U.S. corn belt using Planet Labs CubeSat and STAIR fusion data. *Remote  
661 Sensing of Environment*, 239, 111615
- 662 Li, X., Foody, G.M., Boyd, D.S., Ge, Y., Zhang, Y., Du, Y., & Ling, F. (2020). SFSDAF: An enhanced  
663 FSDAF that incorporates sub-pixel class fraction change information for spatio-temporal image  
664 fusion. *Remote Sensing of Environment*, 237, 111537
- 665 Liu, M., Yang, W., Zhu, X., Chen, J., Chen, X., Yang, L., & Helmer, E.H. (2019). An Improved  
666 Flexible Spatiotemporal Data Fusion (IFSDAF) method for producing high spatiotemporal resolution  
667 normalized difference vegetation index time series. *Remote Sensing of Environment*, 227, 74–89
- 668 Liu, X., Liu, L., Hu, J., & Du, S. (2017). Modeling the footprint and equivalent radiance transfer path  
669 length for tower-based hemispherical observations of chlorophyll fluorescence. *Sensors (Basel)*, 17
- 670 Liu, X.L., Bo, Y.C., Zhang, J., & He, Y.Q. (2015). Classification of C3 and C4 vegetation types using  
671 MODIS and ETM plus blended high spatio-temporal resolution data. *Remote Sensing*, 7, 15244–  
672 15268
- 673 Luo, Y., Guan, K., Peng, J., Wang, S., & Huang, Y. (2020). STAIR 2.0: A Generic and Automatic  
674 Algorithm to Fuse Modis, Landsat, and Sentinel-2 to Generate 10 m, Daily, and Cloud-/Gap-Free  
675 Surface Reflectance Product. *Remote Sensing*, 12, 3209
- 676 Luo, Y., Guan, K., & Peng, J. (2018). STAIR: A generic and fully-automated method to fuse multiple  
677 sources of optical satellite data to generate a high-resolution, daily and cloud-/gap-free surface  
678 reflectance product. *Remote Sensing of Environment*, 214, 87–99
- 679 Moreno-Martínez, Á., Izquierdo-Verdiguier, E., Maneta, M.P., Camps-Valls, G., Robinson, N.,  
680 Muñoz-Marí, J., Sedano, F., Clinton, N., & Running, S.W. (2020). Multispectral high resolution  
681 sensor fusion for smoothing and gap-filling in the cloud. *Remote Sensing of Environment*, 247,  
682 111901
- 683 Morissette, J.T., Baret, F., Privette, J.L., Myneni, R.B., Nickeson, J.E., Garrigues, S., Shabanov, N.V.,

- 684 Weiss, M., Fernandes, R.A., Leblanc, S.G., Kalacska, M., Sanchez-Azofeifa, G.A., Chubey, M.,  
685 Rivard, B., Stenberg, P., Rautiainen, M., Voipio, P., Manninen, T., Pilant, A.N., Lewis, T.E., Iames,  
686 J.S., Colombo, R., Meroni, M., Busetto, L., Cohen, W.B., Turner, D.P., Warner, E.D., Petersen, G.W.,  
687 Seufert, G., & Cook, R. (2006). Validation of global moderate-resolution LAI products: A framework  
688 proposed within the CEOS land product validation subgroup. *IEEE Transactions on Geoscience and*  
689 *Remote Sensing*, 44, 1804–1817
- 690 Olexa, E.M., & Lawrence, R.L. (2014). Performance and effects of land cover type on synthetic  
691 surface reflectance data and NDVI estimates for assessment and monitoring of semi-arid rangeland.  
692 *International Journal of Applied Earth Observation and Geoinformation*, 30, 30–41
- 693 Planet Team (2018). Planet satellite imagery products. In. [WWW Document].  
694 <https://www.planet.com/products/planet-imagery/>
- 695 Rouse, J.W., Haas, R.H., Schell, J.A., Deering, D.W. (1974). Monitoring vegetation systems in the  
696 Great Plains with ERTS. Third ERTS Symposium, 309–317
- 697 Roy, D.P., Zhang, H.K., Ju, J., Gomez-Dans, J.L., Lewis, P.E., Schaaf, C.B., Sun, Q., Li, J., Huang,  
698 H., & Kovalsky, V. (2016). A general method to normalize Landsat reflectance data to nadir BRDF  
699 adjusted reflectance. *Remote Sensing of Environment*, 176, 255–271
- 700 Ryu, Y., Baldocchi, D.D., Verfaillie, J., Ma, S., Falk, M., Ruiz-Mercado, I., Hehn, T., & Sonnentag, O.  
701 (2010). Testing the performance of a novel spectral reflectance sensor, built with light emitting diodes  
702 (LEDs), to monitor ecosystem metabolism, structure and function. *Agricultural and Forest*  
703 *Meteorology*, 150, 1597–1606
- 704 Ryu, Y., Lee, G., Jeon, S., Song, Y., & Kimm, H. (2014). Monitoring multi-layer canopy spring  
705 phenology of temperate deciduous and evergreen forests using low-cost spectral sensors. *Remote*  
706 *Sensing of Environment*, 149, 227–238
- 707 Schmidt, M., Lucas, R., Bunting, P., Verbesselt, J., & Armston, J. (2015). Multi-resolution time series  
708 imagery for forest disturbance and regrowth monitoring in Queensland, Australia. *Remote Sensing of*  
709 *Environment*, 158, 156–168
- 710 Senf, C., Leitao, P.J., Pflugmacher, D., van der Linden, S., & Hostert, P. (2015). Mapping land cover  
711 in complex Mediterranean landscapes using Landsat: Improved classification accuracies from  
712 integrating multi-seasonal and synthetic imagery. *Remote Sensing of Environment*, 156, 527–536
- 713 Storey, J., Choate, M., & Lee, K. (2014). Landsat 8 Operational Land Imager on-orbit geometric  
714 calibration and performance. *Remote Sensing*, 6, 11127–11152
- 715 Tucker, C.J. (1979). Red and photographic infrared linear combinations for monitoring vegetation.  
716 *Remote Sensing of Environment*, 8, 127–150
- 717 Walker, J.J., de Beurs, K.M., & Wynne, R.H. (2014). Dryland vegetation phenology across an  
718 elevation gradient in Arizona, USA, investigated with fused MODIS and Landsat data. *Remote*  
719 *Sensing of Environment*, 144, 85–97
- 720 Walker, J.J., de Beurs, K.M., Wynne, R.H., & Gao, F. (2012). Evaluation of Landsat and MODIS data  
721 fusion products for analysis of dryland forest phenology. *Remote Sensing of Environment*, 117, 381–  
722 393
- 723 Wang, P., Gao, F., & Masek, J.G. (2014). Operational data fusion framework for building frequent  
724 Landsat-like imagery. *IEEE Transactions on Geoscience and Remote Sensing*, 52, 7353–7365
- 725 Wolfe, R.E., Nishihama, M., Fleig, A.J., Kuyper, J.A., Roy, D.P., Storey, J.C., & Patt, F.S. (2002).  
726 Achieving sub-pixel geolocation accuracy in support of MODIS land science. *Remote Sensing of*  
727 *Environment*, 83, 31–49
- 728 Xie, D., Gao, F., Sun, L., & Anderson, M. (2018). Improving spatial-temporal  
729 data fusion by choosing optimal input image pairs. *Remote Sensing*, 10, 1142
- 729 Yang, K., Ryu, Y., Dechant, B., Berry, J.A., Hwang, Y., Jiang, C., Kang, M., Kim, J., Kimm, H.,

- 730 Kornfeld, A., & Yang, X. (2018). Sun-induced chlorophyll fluorescence is more strongly related to  
731 absorbed light than to photosynthesis at half-hourly resolution in a rice paddy. *Remote Sensing of*  
732 *Environment*, 216, 658–673
- 733 Zhang, N., Su, X., Zhang, X., Yao, X., Cheng, T., Zhu, Y., Cao, W., & Tian, Y. (2020). Monitoring  
734 daily variation of leaf layer photosynthesis in rice using UAV-based multi-spectral imagery and a light  
735 response curve model. *Agricultural and Forest Meteorology*, 291, 108098
- 736 Zheng, Y., Wu, B.F., Zhang, M., & Zeng, H.W. (2016). Crop phenology detection using high spatio-  
737 temporal resolution data fused from SPOT5 and MODIS products. *Sensors*, 16, 21
- 738 Zhou, J., Chen, J., Chen, X., Zhu, X., Qiu, Y., Song, H., Rao, Y., Zhang, C., Cao, X., & Cui, X.  
739 (2021). Sensitivity of six typical spatiotemporal fusion methods to different influential factors: A  
740 comparative study for a normalized difference vegetation index time series reconstruction. *Remote*  
741 *Sensing of Environment*, 252, 112130
- 742 Zhou, W., & Bovik, A.C. (2002). A universal image quality index. *IEEE Signal Processing Letters*, 9,  
743 81–84
- 744 Zhu, X., Cai, F., Tian, J., & Williams, T. (2018). spatiotemporal fusion of multisource remote sensing  
745 data: Literature survey, taxonomy, principles, applications, and future directions. *Remote Sensing*, 10,  
746 527
- 747 Zhu, X.L., Chen, J., Gao, F., Chen, X.H., & Masek, J.G. (2010). An enhanced spatial and temporal  
748 adaptive reflectance fusion model for complex heterogeneous regions. *Remote Sensing of*  
749 *Environment*, 114, 2610–2623
- 750 Zhu, X.L., Helmer, E.H., Gao, F., Liu, D.S., Chen, J., & Lefsky, M.A. (2016). A flexible  
751 spatiotemporal method for fusing satellite images with different resolutions. *Remote Sensing of*  
752 *Environment*, 172, 165–177
- 753



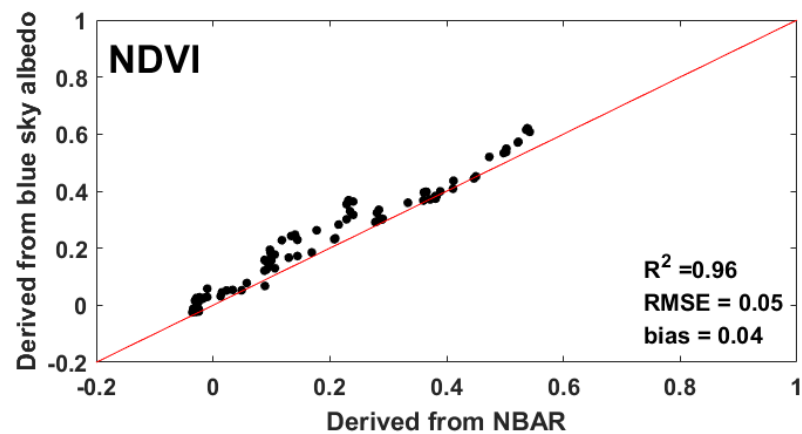
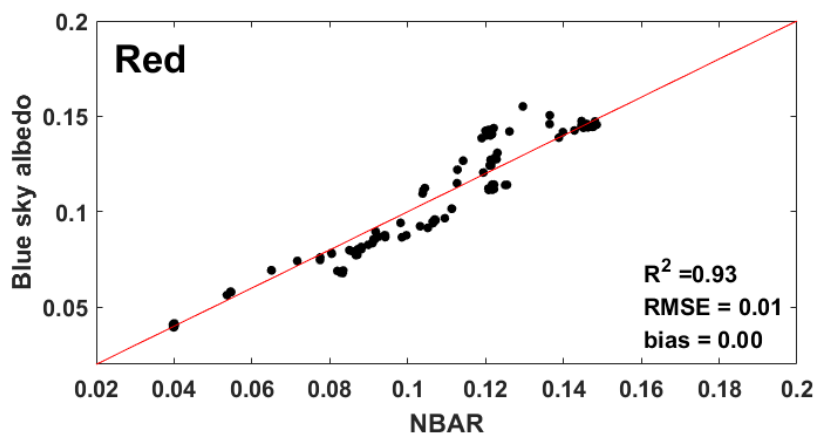
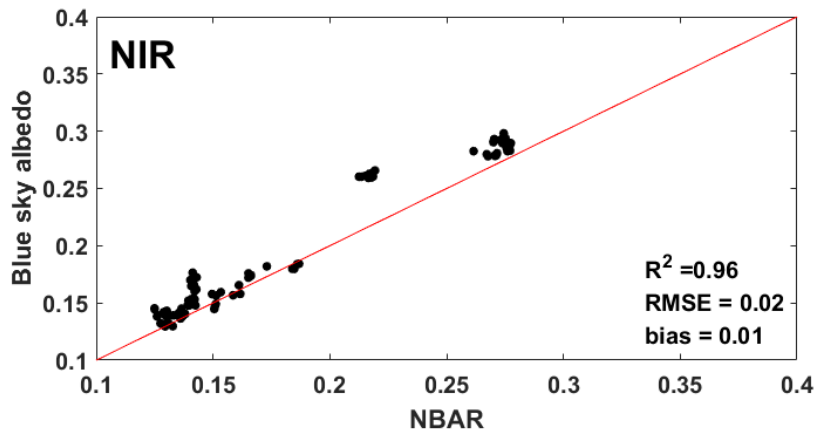
755

756 Appendix 1 The location of ESUs over (a) Sentinel-2 images (DOY 173 since 2017) and (b)  
757 Landsat 8 images (DOY 174 since 2017; RGB rendering: red, NIR, none).

758

759 **Appendix 2 Quality check of Jaz spectrometer data**

760 To check the Jaz spectrometer data, we used an ASD spectrometer (FieldSpec 4 Wide-ResField  
761 Spectroradiometer, ASD, Boulder, CO, USA) on June 4 (DOY 155). Near the flux tower, we  
762 compared *in situ* Jaz spectral measurements with cosine-corrected fiber to 20 ASD measurements  
763 with bare fiber within the Jaz spectrometer footprint. The differences in band between ASD and Jaz in  
764 Landsat 8 band-like data were less than 0.05 (ASD red:  $0.065 \pm 0.007$ , NIR:  $0.127 \pm 0.006$ , NDVI =  
765  $0.327 \pm 0.053$  [95% confidence interval]; Jaz red:  $0.054 \pm 0.0001$ , NIR:  $0.118 \pm 0.004$  NDVI =  $0.376$   
766  $\pm 0.012$  [95% confidence interval]).



767

768 **Appendix 3 Evaluation of view geometry effects on *in situ* spectral measurements. Using**  
 769 **MCD43A1 BRDF parameters, we converted *in situ* bi-hemispheric reflectance measured by Jaz**  
 770 **spectrometer (Yang *et al.* 2018) into blue sky albedo and Nadir BRDF-adjusted reflectance**  
 771 **(NBAR) at 10:30 a.m. (local time, UTC +9). *In situ* data between DOY 100 and DOY 300 (since**  
 772 **2017) were used, excluding cloud-contaminated data.**

773 **Appendix 4 Evaluation of Landsat 8 and fusion NDVI products against *in situ* NDVI on specific**  
 774 **dates. CESTEM on DOY 206 was excluded because of cloud contamination (Table 2). \*: no data**  
 775 **or cloud contamination within  $\pm 3$  days.**

	DOY	155	187	194	206	215	223	239	248	291
Landsat	R <sup>2</sup>	*	*	*	0.22	*	*	*	*	0.17
	rRMSE	*	*	*	11.6 %	*	*	*	*	18.4 %
	rbias	*	*	*	-1.0 %	*	*	*	*	-12.4 %
ESTARFM	R <sup>2</sup>	0.65	0.27	0.36	*	0.59	0.08	0	0.01	0.10
	rRMSE	43.8 %	16.4 %	14.3 %	*	6.8 %	16.5 %	26.1 %	25.7 %	15.8 %
	rbias	-14.1 %	-10.3 %	-6.2 %	*	-1.2 %	-13.7 %	-15.7 %	-5.5 %	-7.0 %
FSDAF	R <sup>2</sup>	0.01	0.26	0.04	0.24	0.25	0.07	0.01	0.11	0.18
	rRMSE	34.7 %	13.5 %	17.6 %	12.3 %	10.1 %	11.1 %	20.5 %	21.2 %	14.6 %
	rbias	-15 %	-3.8 %	-7.9 %	3.9 %	2.8 %	-5.6 %	-10.1 %	-7.6 %	-5.9 %
STAIR	R <sup>2</sup>	0.38	0.39	0.52	0.49	0.87	0.61	0.31	0.25	0.20
	rRMSE	25.7 %	14.4 %	12.4 %	9.6 %	5.8 %	7.7 %	15.2 %	18.1 %	24.7 %
	rbias	-7.2 %	-8.3 %	-3.0 %	0.7 %	3.6 %	4.5 %	-4.6 %	-1.0 %	-20.9 %
CESTEM	R <sup>2</sup>	0.57	0.82	*	*	*	0.87	0.89	0.94	0.47
	rRMSE	20.6 %	7.6 %	*	*	*	4.2 %	7.6 %	16.6 %	21.9 %
	rbias	-3.8 %	-0.9 %	*	*	*	1.7 %	0.7 %	11.5 %	-19.1 %

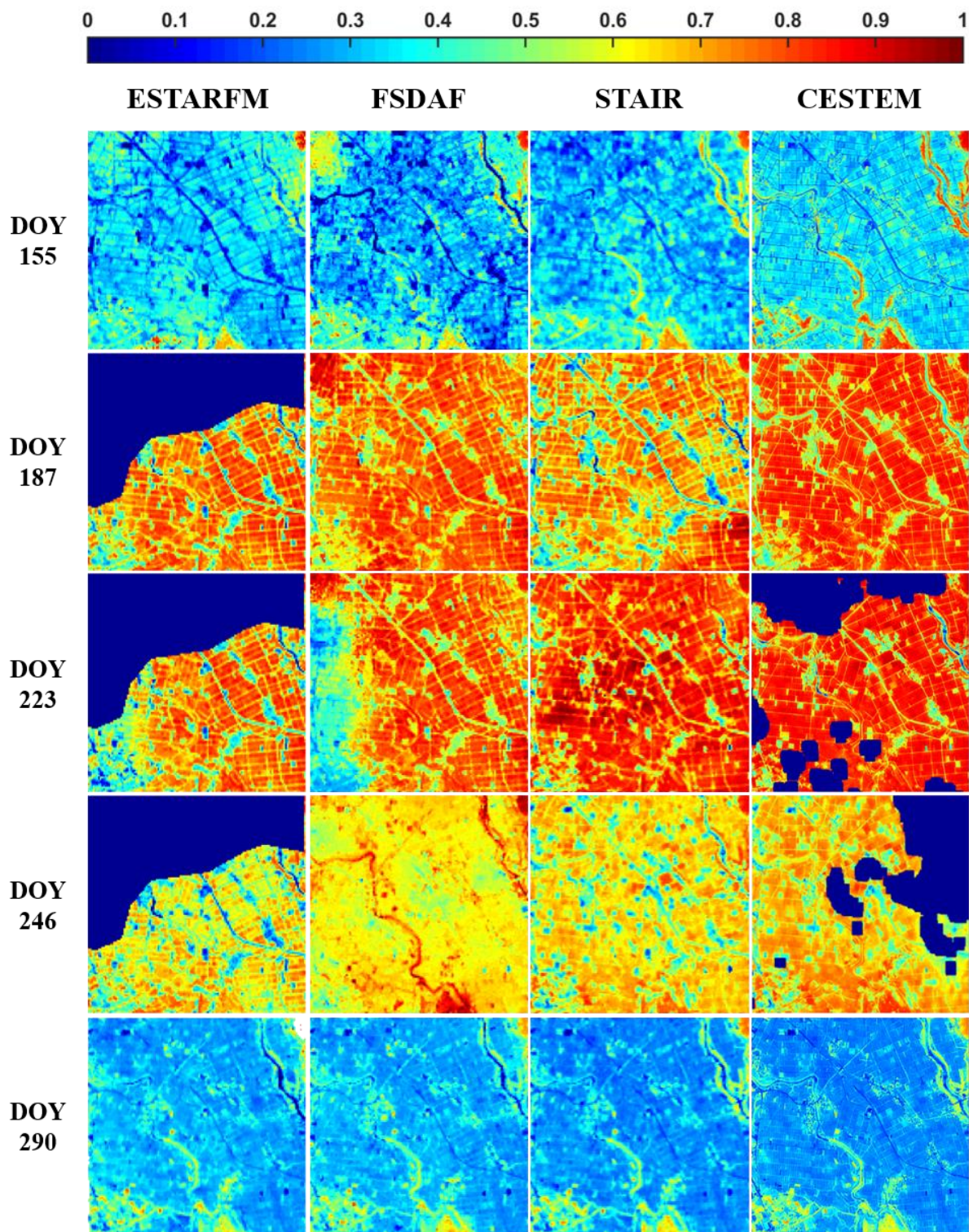
776



777 **Appendix 5 Statistical analyses of each plot during the entire growing season. \*, mixed land**  
 778 **cover plot**

Plot	ESTARFM			FSDAF			STAIR			CESTEM		
	R <sup>2</sup>	<i>rRMSE</i>	<i>rbias</i>	R <sup>2</sup>	<i>rRMSE</i>	<i>rbias</i>	R <sup>2</sup>	<i>rRMSE</i>	<i>rbias</i>	R <sup>2</sup>	<i>rRMSE</i>	<i>rbias</i>
1	0.92	15.5%	-10.3%	0.94	11.6%	-8.2%	0.97	9.4%	-7.5%	0.98	7.5%	-4.1 %
2	0.95	12.8%	-7.9%	0.95	10.9%	-7.2%	0.96	7.5%	-2.8%	0.97	8.1%	-3.4 %
* 3	0.66	24.7%	16.8%	0.75	36.3%	29.4%	0.63	23.9%	11.1%	0.94	17.5%	10.9 %
* 4	0.87	17.0%	12.8%	0.93	21.0%	18.5%	0.92	11.7%	7.3%	0.73	20.4%	7.6 %
5	0.55	31.6%	6.4%	0.73	23.3%	5.3%	0.61	29.3%	10.5%	0.86	20.2%	7.3 %
* 6	0.68	21.9%	-12.3%	0.83	13.7%	3.6%	0.93	9.6%	4.8%	0.97	12.4%	4.3 %
7	0.96	13.0%	-11.4%	0.96	11.8%	-10.1%	0.97	8.4%	-7.0%	0.98	6.4%	-3.9 %
8	0.90	18.4%	-15.5%	0.30	17.4%	-14.9%	0.92	12.6%	-9.4%	0.98	10.2%	-8.7 %
9	0.94	10.3%	-6.9%	0.85	12.7%	-4.0%	0.92	10.9%	-0.1%	0.99	6.9%	-3.2 %
10	0.95	11.8%	-9.2%	0.96	11.9%	-10.0%	0.96	8.6%	-4.2%	0.99	6.5%	-4.9 %
11	0.75	32.2%	-28.2%	0.77	23.0%	-18.4%	0.83	19.5%	-15.3%	0.98	7.0%	-4.7 %
12	0.62	26.9%	-23.3%	0.63	16.9%	-10.2%	0.81	12.8%	-5.9%	0.99	4.5%	-2.4 %
13	0.88	15.1%	-13.0%	0.93	10.2%	-7.1%	0.98	8.0%	-2.0%	0.95	8.5%	-1.9 %
14	0.94	9.7%	-6.7%	0.93	10.2%	-6.5%	0.97	5.2%	-1.6%	0.95	8.0%	0.3 %
15	0.99	10.5%	-10.2%	0.95	9.0%	-7.2%	0.96	7.6%	-3.7%	0.96	6.3%	-3.1 %

779



781

782 **Supplementary 1 NDVI maps ( $3.84 \times 3.84$  km) of fusion products on *in situ***  
783 **measurement dates. The study site is located in the center of each map. For ESTARFM**  
784 **and CESTEM, cloud masks are applied (dark navy).**

785

786 **Supplementary 2 Evaluation of resampling methods for aggregating CESTEM 3m products to**  
 787 **CESTEM 30 m product.**

<b>Interpolation Method</b>	<b>R<sup>2</sup></b>	<b>bias</b>	<b>RMSE</b>
<b>Nearest-neighbor</b>	0.929	0.063	-0.005
<b>Bilinear</b>	0.928	0.063	-0.004
<b>Bicubic</b>	0.928	0.063	-0.003
<b>Nearest-neighbor with antialiasing</b>	0.914	0.069	-0.010
<b>Bilinear with antialiasing</b>	0.907	0.073	-0.010
<b>Bicubic with antialiasing</b>	0.917	0.070	-0.011

788

789 **Supplementary 3 Day-of-year (DOY) data for the NDVI time series.**

<b>Data (Number of Data)</b>	<b>Day of Year since 2017</b>
<i>In situ</i> (9)	155 187 194 206 215 223 239 248 291
Landsat-8 OLI (6)	94 134 174 206 286 302
MODIS (73) (MOD09GQ)	92 93 97 100 102 103 106 109 111 113 116 117 120 121 122 123 126 127 131 134 139 140 141 146 147 152 153 154 155 156 159 161 162 163 166 168 169 173 186 187 195 200 206 213 223 243 244 245 246 256 257 259 263 264 268 269 271 272 275 278 280 286 287 290 293 294 296 297 298 300 301 303 304
ESTARFM (65)	92 97 100 102 103 106 109 111 113 116 117 120 121 122 123 127 131 134 139 140 141 146 152 153 154 155 156 159 161 162 163 166 168 169 186 187 195 200 213 223 243 244 245 246 256 257 259 263 264 268 269 271 272 275 278 280 290 293 294 296 297 298 300 303 304
FSDAF (66)	92 97 100 102 103 106 109 111 113 116 117 120 121 122 123 127 131 134 139 140 141 146 152 153 154 155 156 159 161 162 163 166 168 169 186 187 195 200 213 223 243 244 245 246 256 257 259 263 264 268 269 271 272 275 278 280 286 290 293 294 296 297 298 300 303 304
STAIR (214)	91 to 304 (Daily data, No data gap in STAIR)
CESTEM (60)	92 98 101 102 110 111 112 113 114 119 120 122 126 138 139 141 142 144 147 154 155 161 163 167 168 169 171 174 178 185 187 195 206 215 223 224 238 244 247 258 259 261 263 265 268 269 271 275 277 278 287 290 293 294 296 300 301 303 304

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